

1 **Adaptation and evolutionary rescue in a community context**

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15

16 **Abstract**

17 Evolution might mitigate negative effects of environmental change on diversity by rescuing
18 populations from extinction. However, biotic interactions could affect adaptation to abiotic
19 change, and conversely, adaptation to an abiotic stressor could influence biotic interactions.
20 We used experimental evolution of microalgae to investigate reciprocal effects of competition
21 and adaptation to salt stress. We found that evolutionary history influenced competitive
22 interactions and community properties. In high-salt environments, communities assembled
23 from lines with a history of salt stress declined more slowly in diversity than communities
24 assembled from lines without prior exposure to salt. We also found that a history of salt stress
25 can lead to reduced competitive ability in the ancestral environment. While competition had
26 no effect on adaptation to salt, we found that competition can hinder evolutionary rescue by
27 preventing that adaptation translates into population persistence. Collectively, our results
28 highlight the importance of considering evolutionary rescue in a community context.

29 **Introduction**

30 Abiotic environmental change can render populations maladapted, leading to their decline,
31 and ultimately, extinction (Ceballos & Ehrlich 2002; Stuart *et al.* 2004; Both *et al.* 2006).
32 Adaptive evolution may rescue populations from extinction if genotypes rise in frequency that
33 have positive growth rates under the new environmental conditions (evolutionary rescue;
34 Gomulkiewicz & Holt 1995; Bell & Gonzalez 2009). Laboratory experiments have
35 demonstrated that microorganisms can adapt rapidly to a variety of abiotic stressors, including
36 acidification (Lohbeck *et al.* 2012), salt stress (Lachapelle & Bell 2012), resource limitation
37 (Bell 2013), and warming (Schaum *et al.* 2018). Experimental evolution has also helped to
38 reveal the demographic, genetic, and extrinsic factors that influence the probability of
39 evolutionary rescue (Carlson *et al.* 2014). For example, evolutionary rescue is facilitated by
40 larger population size (Bell & Gonzalez 2009), higher genetic variation (Lachapelle & Bell
41 2012), and a slower rate of environmental change (Lindsey *et al.* 2013). However, most
42 empirical studies on evolutionary rescue have considered single species in isolation. Yet,
43 interactions with other species might influence (i) whether populations adapt to abiotic
44 stressors, and (ii) whether adaptation translates into population persistence (Fig. 1a).
45 Moreover, within communities, the ecological consequences of evolutionary rescue could go
46 beyond population persistence and include effects on species interactions and community
47 dynamics (Fig. 1b). This potentially complex interplay of biotic interactions and adaptation to
48 abiotic stressors limits our understanding of whether evolution can contribute to the
49 maintenance of diversity in changing environments.

50 Theory suggests several mechanisms whereby interspecific competition can influence
51 adaptation to abiotic change. First, competition can constrain adaptation to environmental
52 change by reducing population size, and thus the amount of standing genetic variation
53 (Bocedi *et al.* 2013) and the supply of beneficial mutations (Johansson 2008). Second,

54 competition can reduce the time available for adaptation as species are outcompeted by better
55 adapted competitors (de Mazancourt *et al.* 2008). Third, competition can alter the strength and
56 direction of selection (Osmond & de Mazancourt 2013); depending on whether competition
57 and environmental change impose selection on a trait in the same or in the opposite direction,
58 competition can speed up or hinder adaptation to environmental change (Osmond & de
59 Mazancourt 2013). Finally, negative genetic correlations between traits that mediate
60 competition and stress tolerance could constrain adaptation in environments with both biotic
61 and abiotic selective agents (Lau & Terhorst 2015). Empirical tests of these theoretical
62 predictions are rare and focused on biotic interactions other than interspecific competition
63 (Collins 2011; Hiltunen *et al.* 2018) or on environmental amelioration rather than
64 deterioration (Kleynhans *et al.* 2016).

65 Competition could not only alter how species adapt to abiotic change, but could also
66 modulate the effect of adaptation on population dynamics. When competition alters the
67 selection environment generated by abiotic change, adaptation of a focal species in isolation
68 may not translate into increased population size in a community context. Accordingly, an
69 experiment that manipulated CO₂ and diversity of a plant community showed that
70 interspecific competition altered the selection environment created by CO₂ (Kleynhans *et al.*
71 2016): adaptation to elevated CO₂ was only observed when the plants were assayed in the
72 same community context in which selection had occurred. Similarly, an experiment with a
73 marine alga revealed that adaptation to acidification in isolation was not associated with
74 increased population size within communities (Bach *et al.* 2018). In the context of
75 evolutionary rescue, competition could prevent that adaptation translates into population
76 persistence and thus preclude that evolution mitigates negative effects of environmental
77 change on diversity.

78 The field of evolutionary rescue primarily considers the effect of evolution on population
79 dynamics, but within communities, adaptation of a focal species to the local environment may
80 affect interactions with other species (terHorst *et al.* 2014; Pantel *et al.* 2015; Gomez *et al.*
81 2016). Evolution can influence community dynamics directly, by altering traits that underlie
82 species interactions, and indirectly, by influencing population dynamics (Matthews *et al.*
83 2011; Hendry 2019). In a scenario with an abiotic stressor as the selective agent (Fig. 1b),
84 adaptation could influence species interactions if traits that confer adaptation to the stressor
85 are correlated with traits that affect biotic interactions. Some traits may even simultaneously
86 mediate stress tolerance and species interactions. For example, in *E. coli* reduced membrane
87 permeability increases resistance to antibiotics but decreases the ability to take up nutrients,
88 such that adaptation to antibiotics results in lower competitive ability (Phan & Ferenci 2013).
89 In addition to such trait-mediated effects, adaptation of a focal species to an abiotic stressor
90 might influence biotic interactions by increasing its population size. Despite this range of
91 possibilities of how adaptation to abiotic change could affect species interactions, the
92 ecological consequences of evolutionary rescue have been rarely considered in a community
93 context.

94 A further complexity of adaptation in a multi-species context is that species might vary in
95 their ability to adapt to environmental change, which could hamper maintenance of diversity.
96 The rate of evolution depends on selection pressure, genetic variance, and covariance (Lande
97 1979; Agrawal & Stinchcombe 2009). Variation among species in ecological tolerance, life-
98 history traits, and genomic architecture could thus result in variation in evolvability
99 (Barraclough 2015). Species with narrow tolerance of the abiotic stressor will have a higher
100 risk of extinction, but will also experience stronger selection pressure and might thus evolve
101 more than species with broader tolerance (Osmond & de Mazancourt 2013). Species also vary
102 widely in population density (Damuth 1987), mutation rate (Drake *et al.* 1998; Baer *et al.*

103 2007), recombination rate (Stapley *et al.* 2017), and genome complexity (Lynch *et al.* 2011),
104 all of which influence the rate of evolution. Such variation among species in evolutionary
105 potential could lead to changes in the composition and interaction structure of communities in
106 deteriorating environments.

107 Here we investigated if adaptation to abiotic change influences competitive interactions
108 and community properties, and, vice versa, if competition influences adaptation to
109 deteriorating environmental conditions. In a laboratory experiment, we propagated
110 monocultures of six species of microalgae for ~ 180 generations with and without increasing
111 salt stress. We then quantified monoculture growth rates of all lines in a reciprocal transplant
112 assay to test the hypothesis that species vary in their evolutionary response to salt stress
113 (Hypothesis 1). We then assembled communities, either from lines with a history of salt stress
114 or from lines without prior exposure to salt, and investigated if evolutionary history affected
115 community dynamics in environments with and without salt, respectively. We predicted that
116 adaptation to salt in isolation would translate into longer persistence of species within salt-
117 stressed communities and consequently into maintenance of diversity (Hypothesis 2).
118 Alternatively, competitive interactions could prevent that adaptation of monocultures
119 translates into persistence within communities. In a second set of experiments, we propagated
120 two algae species for ~ 200 generations as monocultures and as pairs, respectively, in
121 environments with and without increasing salt. We then re-isolated the two species, and
122 measured monoculture growth rates and competitive abilities in reciprocal transplant assays.
123 We predicted that interspecific competition during the selection experiment would constrain
124 adaptation to salt stress (Hypothesis 3). As in the first set of experiments, we expected that a
125 history of salt stress would lead to increased competitive ability at high salt (Hypothesis 4).
126 Our results highlight the complexities that arise from adaptation to abiotic change in a
127 community context. We found that adaptation to abiotic stress influenced competitive

128 interactions and community-level properties, but we observed no effect of competition on
129 adaptation to salt stress.

130

131 **Material and methods**

132 *Model organisms*

133 We tested our hypotheses using six species of microalgae from three different taxonomic
134 groups: the cyanobacteria *Synechococcus leopoliensis* (Canadian Phycological Culture Centre
135 (CPCC) 102) and *Anabaena variabilis* (CPCC 105), the diatoms *Navicula pelliculosa* (CPCC
136 552) and *Nitzschia palea* (CPCC 160), and the chlorophytes *Pseudokirchneriella subcapitata*
137 (CPCC 37) and *Scenedesmus acutus* (CPCC 10). The algae were cultivated in Bold's basal
138 medium (Bold 1949), which was modified by adding silicate (0.58 g L⁻¹ of Na₂SiO₃) and
139 vitamins (Low-Décarie *et al.* 2011) to allow growth of diatoms. In all experiments, the algae
140 were cultivated at 25 °C with light continuously provided at 100 µE m⁻² s⁻¹. Cultures were
141 propagated under sterile conditions. We did not make cultures isogenic at the start of the
142 experiment; the observed evolutionary responses could thus have resulted either from *de novo*
143 mutations or from sorting of standing genetic variation.

144

145 *Salt tolerance of ancestral lines*

146 We determined the limit of salt tolerance of the six species by measuring their growth along a
147 salt gradient implemented on 48-well plates. The length of the gradient was based on results
148 of a preliminary experiment, and was 0 to 10 g NaCl L⁻¹ in 1 g increments for *Nitzschia* and
149 *Pseudokirchneriella*, and 0 to 20 g NaCl L⁻¹ in 2 g increments for the remaining species. After
150 six days of acclimation at the respective salt concentration, we inoculated fresh plates with
151 acclimated cultures (20 µl culture in 1 ml medium), sealed the plates with sterile air-
152 permeable membranes (VWR), and cultivated them on shakers (350 rpm). Fifteen times over

153 the course of ten days, we measured absorbance at 660 nm on an optical plate reader
154 (Synergy-HT, BioTek, Winooski, VT). We estimated the limit of salt tolerance for each
155 species as the salt concentration where we observed no increase in absorbance. Each salt ×
156 species combination was replicated twice.

157

158 *Adaptation to salt in monoculture*

159 In a serial transfer experiment, we propagated monocultures of the six species with and
160 without increasing concentrations of NaCl (Fig. S1, S2). The algae were cultured in 125-ml
161 glass flasks filled with 50 ml of Bold's medium and continuously shaken at 250 rpm. Every
162 3.5 days, we transferred 1 ml of culture to 50 ml of fresh medium for a total of 37 transfers (~
163 180 generations). For each of the six species, three replicate control lines were cultivated
164 without addition of salt and three replicate selection lines were cultivated with increasing salt.
165 We increased the salt concentration in the selection lines by 0.25 g L^{-1} every transfer until
166 reaching 4 g NaCl L^{-1} at transfer 16 (Fig. S2). We maintained this salt concentration for 10
167 transfers and then continued the increase in salt until reaching 6 g NaCl L^{-1} at transfer 34. We
168 chose a final salt concentration of 6 g NaCl L^{-1} because this was the limit of salt tolerance for
169 the species with the lowest salt tolerance among the six species.

170 To test if the six species of algae differed in their evolutionary response to salt stress
171 (Hypothesis 1), we measured the growth rate of each line in the two final salt environments (0
172 and 6 g NaCl L^{-1}). Before starting the growth assay, we removed non-evolutionary effects by
173 cultivating all lines at identical environmental conditions for three transfer cycles, that is two
174 cycles in the ancestral environment (0 g NaCl L^{-1}) and one cycle in the assay environment (0
175 and 6 g NaCl L^{-1}). We then inoculated flasks containing 50 ml medium of either 0 or 6 g NaCl
176 L^{-1} with 1 ml of acclimated culture and measured absorbance through time. Depending on the
177 time of individual cultures to reach carrying capacity, we measured absorbance over 2.5 to 7.5

178 weeks with 23 to 34 measurements. Absorbance was measured twice a day during the first 5
179 days, once a day during the intermediate phase of the experiment, and once every other day
180 towards the end of the experiment. At each measurement, we removed 200 μ l of culture from
181 each flask and measured absorbance in 96-well plates on an optical plate reader.

182

183 *Community assay*

184 To test if evolutionary history can influence community dynamics (Hypothesis 2), we
185 assembled communities either of control or of selection lines, transplanted the communities
186 into the two final salt environments of the selection experiment, and quantified species
187 abundances and diversity of the community over four transfers. Prior to the community assay,
188 the lines had been cultivated for 37 transfers with or without increasing salt, followed by two
189 transfers in the ancestral environment (see above). We assembled communities with equal
190 biovolume of the six species. To this end, we determined species abundances in the
191 monocultures by counting Lugol-fixed samples with an inverted microscope, and then
192 converted abundances to biovolume using cellular biovolume estimates based on
193 measurements of 20 individuals per species (Hillebrand *et al.* 1999). We assembled three
194 replicate communities of control lines and three replicate communities of selection lines.
195 From each of the six communities, we inoculated two flasks, one with a salt concentration of
196 0 and one of 6 g NaCl L⁻¹. Every four days, we transferred 1 ml of the community into 50 ml
197 of fresh medium, with the flasks continuously shaken at 250 rpm. Over the course of four
198 transfers, we sampled the communities at each transfer, fixed the samples with Lugol's
199 solution, and counted species abundances with an inverted microscope.

200

201

202

203 *Adaptation in monoculture vs. adaptation in pairs*

204 To test if competition can influence adaptation to salt stress (Hypothesis 3), we propagated
205 monocultures and pairs of *Anabaena* and *Scenedesmus* with and without increasing salt for 46
206 transfers, then re-isolated the two species and quantified their fitness with and without salt
207 (Fig. S1b). The monocultures and pairs of the two species were part of the selection
208 experiment described above and experienced the same changes in salt concentration (Fig. S2).
209 To quantify how strongly competition affected the abundances of the two species during the
210 selection experiment, we sampled the pairs and monocultures of *Anabaena* and *Scenedesmus*
211 ten times over the course of the selection experiment and microscopically determined the
212 abundances of the two species in Lugol-fixed samples. We chose the combination of
213 *Anabaena* and *Scenedesmus* because in other species combinations the inferior competitor
214 was outcompeted quickly both with and without increasing salt.

215 After 46 transfers in the selection experiment, we re-isolated *Anabaena* and
216 *Scenedesmus* by plating the cultures on agar and isolating individual colonies. Monocultures
217 were treated the same way as the pairs. Individual colonies were resuspended in liquid Bold's
218 medium on well plates, and microscopically checked for contamination after a week of
219 growth. We then merged eight clean isolates per sample to one culture. The two isolation
220 steps were conducted in medium without addition of NaCl. We then acclimated the cultures to
221 the two assay environments (0 and 6 g NaCl L⁻¹) for two transfer cycles. Hence, before
222 starting the growth and competition assays, the cultures were cultivated under identical
223 environmental conditions for four cycles (two isolation cycles without salt, two transfer cycles
224 in the assay environment) to remove any non-evolutionary effects.

225 The acclimated cultures were used to start growth and competition assays. Growth rates
226 of all lines were determined in monoculture assays in the two final salt environments by
227 measuring absorbance through time as described above. To test if salt history and community

228 history influenced competitive interactions (Hypothesis 4), we assembled pairs of *Anabaena*
229 and *Scenedesmus* in the two final salt environments and determined changes in biomass of the
230 two species through time. We conducted two competition assays that differed in how we
231 assembled the pairs. First, we assembled pairs of co-evolved *Anabaena* and *Scenedesmus*, i.e.
232 lines that had the same community and salt history. However, because both competitors
233 differed between flasks in this assay, differences among treatments could not unequivocally
234 be attributed to one of the two species. We thus conducted a second competition assay where
235 we competed each line of *Anabaena* with the same reference line of *Scenedesmus*, and each
236 line of *Scenedesmus* with the same reference line of *Anabaena*. As reference line, we
237 randomly selected one of the three replicate control lines (i.e. community history:
238 monoculture, salt history: no salt). The second competition assay was started three weeks after
239 the first competition assay; cultures were stored meanwhile as monocultures at their
240 respective salt selection environment and then acclimated for one cycle to the salt assay
241 environment.

242 In the competition assays, we measured chlorophyll a concentrations of the two
243 competitors as proxy for biomass. To quantify chlorophyll a concentrations of the two
244 species, we used a FluoroProbe (bbe Moldaenke, Kiel-Kronshagen, Germany), a fluorometer
245 which can discriminate among algal groups (Catherine *et al.* 2012). Before the start of the
246 assay, we calibrated the FluoroProbe using cultures of *Anabaena* and *Scenedesmus* with
247 known chlorophyll a concentrations. We started the competition assay with equal chlorophyll
248 a concentrations (2 µg) of both species in flasks filled with 50 mL of Bold's medium. We
249 transferred the cultures every 3.5 days for ten transfers in the first competition assay and for
250 four transfers in the second competition assay and measured chlorophyll a concentrations of
251 both species at each transfer.

252

253 *Data analysis*

254 We estimated growth rates in monoculture assays by fitting linear models to the exponential
255 part of the absorbance vs. time curves, using ln-transformed absorbance data as response
256 variable. We computed two-way ANOVAs to investigate effects of salt history and assay
257 environment on growth rates of the six species. We analyzed the time series data of the
258 community assay with linear mixed models, with ln-transformed abundance as response
259 variable, transfer, salt history and assay environment as fixed effects, and flask as random
260 effect to account for repeated sampling of communities. We used the same approach to
261 investigate treatment effects on species richness. To test whether competition affected
262 adaptation of *Anabaena* and *Scenedesmus* to increasing salt, we computed three-way
263 ANOVAs and investigated the effects of community history, salt history and assay
264 environment on monoculture growth rates (growth assay) and relative chlorophyll a
265 (competition assay). All analyses were computed with R version 3.4.1 (R Core Team 2015);
266 the package *lme4* was used to calculate linear mixed effect models.

267
268 **Results**

269 *Salt tolerance of ancestral lines*

270 Among the six species, *Nitzschia* and *Pseudokirchneriella* had the lowest limit of salt
271 tolerance (5 and 6 g NaCl L⁻¹, respectively), *Anabaena*, *Scenedesmus* and *Navicula* had
272 intermediate limits of salt tolerance (12-16 g NaCl L⁻¹), and *Synechococcus* had the highest
273 limit of salt tolerance (18-20 g NaCl L⁻¹).

274

275 *Adaptation of monocultures*

276 After 37 transfers with and without increasing salt, respectively, we detected adaptation to salt
277 in two of six species, *Nitzschia* and *Pseudokirchneriella* (Fig. 2). In both species, the growth
278 response to the assay environment depended on salt history: selection lines had higher growth

279 rates at 6 g NaCl L⁻¹ than control lines, but did not differ from control lines in the no-salt assay
280 environment (Fig. 2 c, d). The effect of salt history was particularly strong in *Nitzschia*; the
281 control lines were unable to grow at the final salt concentration, while all three replicate
282 selection lines had positive growth rates. Among the remaining four species, *Scenedesmus*
283 and *Synechococcus* were affected neither by salt history nor assay environment, while
284 *Navicula* and *Anabaena* had reduced growth rates at high salt, irrespective of salt history (Fig.
285 2).

286

287 *Community assay*

288 When assaying communities of either control or selection lines, we found that the effect of
289 salt history varied among species and assay environment (Table 1, Fig. 3). The abundances of
290 most species declined over the four transfers except for *Synechococcus*, which increased.
291 Assay environment had no effect on the abundances of *Navicula* and *Synechococcus*, while
292 the other four species declined faster when the community was assayed at high salt than
293 without salt. Salt history affected three of the six species (*Nitzschia*, *Pseudokirchneriella*, and
294 *Scenedesmus*), but the effect depended on the assay environment (Table 1). The strongest
295 effect of salt history was observed for *Pseudokirchneriella* (Fig. 3): in the high-salt assay
296 environment, selection lines declined more slowly in abundance than control lines, resulting
297 in extinction being delayed by one transfer. Similarly, selection lines of *Scenedesmus* declined
298 more slowly at high salt than control lines, but the effect was of small magnitude (Fig. 3). In
299 contrast, selection lines of *Nitzschia* declined slightly faster than control lines when assayed in
300 the no-salt environment.

301 The effect of evolutionary history on species abundances translated into an effect on the
302 diversity of the community (Table S1, Fig. 4). Species richness declined with time (transfer: P
303 < 0.001), and this decline was faster when communities were assayed with than without salt

304 (assay: $P < 0.001$). However, communities assembled from selection lines declined more
305 slowly in species richness than communities assembled from control lines (salt history \times
306 assay; $P = 0.047$), because of slower extinction of *Pseudokirchneriella*.
307

308 *Adaptation in monoculture vs. adaptation in pairs*

309 Interspecific competition strongly influenced species abundances during the selection
310 experiment, but did not lead to any evolutionary responses. In the selection experiment,
311 *Anabaena* reduced the abundance of *Scenedesmus* as long as salt concentration was low (Fig.
312 S3). However, at the final salt concentration of 6 g NaCl L⁻¹ the competitive hierarchy
313 reversed and *Anabaena* was negatively affected by *Scenedesmus* (Fig. S3). When assayed as
314 monocultures after the selection experiment, neither of the two species showed evolutionary
315 responses (Table S2, Fig. 5a, b). Salt had a negative effect on the growth rate of both species
316 ($P < 0.001$), irrespective of salt history and community history.

317 The two competition assays revealed that lines of *Anabaena* with a history of salt
318 stress had reduced competitive ability in the ancestral environment. When assayed with co-
319 evolved competitors, high-salt lines of *Anabaena* had lower relative chlorophyll a than no-salt
320 lines in the no-salt assay environment, and consequently high-salt lines of *Scenedesmus* had
321 higher relative chlorophyll a than no-salt lines (salt history \times assay: $P = 0.002$; Table S3, Fig.
322 S4). Because both competitors differed among treatment combinations, this experiment did
323 not allow detecting if the observed pattern was due to altered competitive ability of *Anabaena*
324 or of *Scenedesmus*. However, when assayed in competition with a reference line, only
325 *Anabaena* was affected by salt history, again with lower relative chlorophyll a of high-salt
326 than no-salt lines in the no-salt assay environment (salt history \times assay: $P = 0.029$, Table S4,
327 Fig. 5c, d).

328

329 **Discussion**

330 We found that adaptation to an abiotic stressor can influence competitive interactions and
331 community-level properties, but we observed no effect of competition on adaptation to salt.
332 Communities assembled from lines with a selection history of salt stress maintained diversity
333 longer under salt stress than communities assembled from control lines. The positive effect of
334 salt history on diversity resulted from longer persistence of selection lines of
335 *Pseudokirchneriella* in salt-stressed communities compared with control lines. However, we
336 also found that adaptation of monocultures to salt stress did not necessarily translate into
337 longer persistence within communities: *Nitzschia* adapted to salt in isolation, but selection and
338 control lines went extinct equally fast within salt-stressed communities. This result indicates
339 that competition can override the effect of evolution on population dynamics. Finally, the
340 second set of experiments revealed that a history of salt stress can lead to reduced competitive
341 ability in benign conditions. When assayed in competition, selection lines of *Anabaena*
342 performed worse than control lines in the no-salt assay environment. Taken together, we
343 found different scenarios of how adaptation to an abiotic stressor can play out in a community
344 context, ranging from a positive effect of adaptation on performance within stressed
345 communities, over no effect, to a negative effect on performance under benign conditions.

346 Species varied in their evolutionary response to salt stress. The two species with the
347 lowest limit of salt tolerance adapted to salt, while species that were less negatively affected
348 by salt showed no sign of adaptation (Fig. 2). This finding is in line with the prediction that
349 species with narrow tolerance of a stressor should evolve more than species with broader
350 tolerance (Osmond & de Mazancourt 2013) because they experience stronger selection
351 pressure. Species-specific differences in genetic variance and covariance could be additional
352 reasons for variation among species in the rate of evolution (Lande 1979; Barraclough 2015).
353 However, given that only two species adapted, our sample size is not large enough to evaluate

354 if species-specific differences other than ecological tolerance influenced the evolutionary
355 response to salt. To elucidate such effects, future experiments could expose a set of species to
356 their respective limit of tolerance and then investigate if differences in life-history traits (e.g.
357 population size) influence the rate of adaptation. In our experiment, however, we simulated a
358 scenario where all species experienced the same amount of stress.

359 Evolution transiently mitigated the negative effect of salt stress on diversity, but did not
360 maintain diversity at unstressed levels. In high-salt environments, diversity declined more
361 slowly when communities were assembled from lines with a history of salt stress (Fig. 4),
362 because selection lines of *Pseudokirchneriella* persisted for one transfer longer than control
363 lines (Fig. 3). In *Nitzschia*, however, adaptation to salt in isolation was not associated with
364 increased fitness in salt-stressed communities. Consequently, even in communities of
365 selection lines, diversity declined faster in the high-salt than the no-salt assay environment
366 (Fig. 4). Our results thus suggest that competitive interactions can hinder evolutionary rescue
367 by preventing that adaptation translates into persistence of the population. The scenario of
368 adaptation in isolation followed by competition with pre-adapted species might seem
369 artificial, but it is a plausible scenario in a spatial context. Species may adapt to changes in
370 their local environment before immigration of better-adapted species from the regional
371 species pool (Urban *et al.* 2012). Depending on how well the resident species manage to adapt
372 to the new environmental conditions, they may coexist with or even dominate over pre-
373 adapted, immigrating species (De Meester *et al.* 2016). In our experiment, however, the
374 positive effect of salt history on persistence of *Pseudokirchneriella*, and thus on the diversity
375 of the community, was only transient. By the end of the experiment, this species had gone
376 extinct at high salt irrespective of salt history. However, *Pseudokirchneriella* also declined to
377 low or zero abundance in the no-salt assay environment, indicating that this species was
378 generally a weak competitor, irrespective of salt concentration. The effect of adaptation on

379 diversity and community dynamics might be more pronounced and persistent for species with
380 high interaction strengths and strong ecosystem effects, such as superior competitors or
381 efficient predators (Pantel *et al.* 2015; Matthews *et al.* 2016).

382 Competition had no effect on adaptation to salt stress, neither in terms of growth rate nor
383 in terms of competitive ability (Fig. 5). We had predicted that competition during the
384 selection experiment would constrain adaptation to salt because of reduced population size
385 and/or increased genetic constraints (Johansson 2008; Lau & Terhorst 2015). Alternatively,
386 competition could facilitate adaptation to abiotic stress by increasing the strength of selection
387 (Osmond & de Mazancourt 2013). The two species that we used to test this question,
388 *Anabaena* and *Scenedesmus*, adapted to salt neither in monoculture nor in pairs. Given that
389 these species had not reached their limit of salt tolerance in the selection experiment, low
390 selection pressure could explain why the monocultures did not adapt to salt. However, the
391 combination of salt and competition in the selection experiment drove the inferior competitor
392 to the brink of extinction (Fig. S3), suggesting that the inferior competitor experienced strong
393 selection pressure to evolve faster growth rates. Yet, we did not find any effect of community
394 history on adaptation. One possible explanation is that the competitive hierarchy in flasks
395 with increasing salt changed shortly before the end of the selection experiment when the final
396 salt concentration had been reached, from dominance of *Anabaena* to dominance of
397 *Scenedesmus*. Maybe relaxed selection pressure on *Scenedesmus* towards the end of the
398 experiment and only a short phase of strong selection on *Anabaena* explain why competition
399 had no effect on the evolutionary response of these two species.

400 Evolutionary change in response to salt influenced competitive interactions (Fig. 5c).
401 We had predicted that a history of salt stress would lead to increased competitive ability at
402 high salt. In contrast, however, strains of *Anabaena* with a history of salt stress had reduced
403 competitive ability in the no-salt assay environment. Such a pattern is an indication of

404 conditionally deleterious mutations (Mee & Yeaman 2019), i.e. mutations that are neutral in
405 the local environment and deleterious in an alternative environment. Because we have neither
406 molecular nor trait data, we cannot elucidate potential mechanisms behind the observed mal-
407 adaptation in the ancestral environment. The evolutionary response of *Anabaena* was
408 independent of community history and of whether fitness was assayed in competition with a
409 co-evolved or a naïve competitor (Fig. 5, Fig. S4). Importantly, however, we observed
410 reduced fitness of selection lines of *Anabaena* only when assayed in competition, not in the
411 monoculture growth assay. It thus seems that the evolutionary response of *Anabaena* to salt
412 was associated with changes in traits that influenced competitive interactions. Similar
413 evidence comes from a functional genomics study on *Daphnia*, which suggested that
414 evolution of salt tolerance could entail fitness costs in environments with predators (Latta *et*
415 *al.* 2012). Overall, our results suggest that evolutionary responses to abiotic change can
416 influence biotic interactions, such that populations with different histories of abiotic stress
417 might vary in their effects on communities and ecosystems.

418 The emerging field of eco-evolutionary dynamics investigates the effects of evolution
419 on populations, communities, and ecosystems (Hendry 2016). Evolutionary rescue, as one
420 component of this field, has its focus on the effect of evolution on population dynamics,
421 specifically on population persistence. Here we showed that the ecological consequences of
422 evolutionary rescue can include effects on species interactions and the diversity of
423 communities. Moreover, we found that competition can override the effect of evolution on
424 population dynamics, illustrating that community context can modulate the ecological
425 consequences of evolutionary change. We have observed a range of complexities arising from
426 a multi-species context despite using a fairly simple model system. Higher diversity and
427 trophic structure of natural systems will pose an even greater challenge to understanding eco-
428 evolutionary dynamics (De Meester *et al.* 2019; Hendry 2019). Field studies have

429 demonstrated evolutionary rescue of populations embedded in diverse communities (Reid *et*
430 *al.* 2016; Oziolor *et al.* 2019), but it remains to be determined if stress tolerant genotypes have
431 the same ecosystem effects as ancestral genotypes. If adaptation to an abiotic stressor is
432 associated with changes in traits that mediate biotic interactions, evolution may maintain
433 populations but fail to maintain community and ecosystem structure. Collectively, our results
434 suggest that considering community context is of fundamental importance if we want to
435 evaluate the capacity of evolution to maintain diversity and species interactions in changing
436 environments.

437

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443

444 **References**

445 Agrawal, A.F. & Stinchcombe, J.R. (2009). How much do genetic covariances alter the rate of
446 adaptation? *Proc. R. Soc. B*, 276, 1183-1191.

447 Bach, L.T., Lohbeck, K.T., Reusch, T.B.H. & Riebesell, U. (2018). Rapid evolution of highly
448 variable competitive abilities in a key phytoplankton species. *Nat. Eco. Evo.*, 2, 611-
449 613.

450 Baer, C.F., Miyamoto, M.M. & Denver, D.R. (2007). Mutation rate variation in multicellular
451 eukaryotes: causes and consequences. *Nat. Rev. Genet.*, 8, 619-631.

452 Barraclough, T.G. (2015). How do species interactions affect evolutionary dynamics across
453 whole communities? *Annu. Rev. Ecol. Evol. Syst.*, 46, 25-48.

454 Bell, G. (2013). Evolutionary rescue of a green alga kept in the dark. *Biol. Lett.*, 9, 20120823.

455 Bell, G. & Gonzalez, A. (2009). Evolutionary rescue can prevent extinction following
456 environmental change. *Ecol. Lett.*, 12, 942–948.

457 Bocedi, G., Atkins, K.E., Liao, J.S., Henry, R.C., Travis, J.M.J. & Hellmann, J.J. (2013).
458 Effects of local adaptation and interspecific competition on species' responses to
459 climate change. *Ann. N. Y. Acad. Sci.*, 1297, 83-97.

460 Bold, H. (1949). The morphology of *Chlamydomonas chlamydogama*, sp. nov. *Bull. Torrey
461 Bot. Club*, 76, 101-108.

462 Both, C., Bouwhuis, S., Lessells, C.M. & Visser, M.E. (2006). Climate change and population
463 declines in a long-distance migratory bird. *Nature*, 441, 81-83.

464 Carlson, S.M., Cunningham, C.J. & Westley, P.A.H. (2014). Evolutionary rescue in a
465 changing world. *Trends Ecol. Evol.*, 29, 521-530.

466 Catherine, A., Escoffier, N., Belhocine, A., Nasri, A.B., Hamlaoui, S., Yepremian, C. *et al.*
467 (2012). On the use of the FluoroProbe®, a phytoplankton quantification method based
468 on fluorescence excitation spectra for large-scale surveys of lakes and reservoirs.
469 *Water Res.*, 46, 1771-1784.

470 Ceballos, G. & Ehrlich, P.R. (2002). Mammal population losses and the extinction crisis.
471 *Science*, 296, 904-907.

472 Collins, S. (2011). Competition limits adaptation and productivity in a photosynthetic alga at
473 elevated CO₂. *Proc. R. Soc. B*, 278, 247-255.

474 Damuth, J. (1987). Interspecific allometry of population density in mammals and other
475 animals: the independence of body mass and population energy-use. *Biol. J. Linn.
476 Soc.*, 31, 193-246.

477 de Mazancourt, C., Johnson, E. & Barraclough, T.G. (2008). Biodiversity inhibits species'
478 evolutionary responses to changing environments. *Ecol. Lett.*, 11, 380–388.

479 De Meester, L., Brans, K.I., Govaert, L., Souffreau, C., Mukherjee, S., Vanvelk, H. *et al.*
480 (2019). Analysing eco-evolutionary dynamics - the challenging complexity of the real
481 world. *Funct. Ecol.*, 33, 43-59.

482 De Meester, L., Vanoverbeke, J., Kilsdonk, L.J. & Urban, M.C. (2016). Evolving perspectives
483 on monopolization and priority effects. *Trends Ecol. Evol.*, 31, 136-146.

484 Drake, J.W., Charlesworth, B., Charlesworth, D. & Crow, J.F. (1998). Rates of spontaneous
485 mutation. *Genetics*, 148, 1667-1686.

486 Gomez, P., Paterson, S., De Meester, L., Liu, X., Lenzi, L., Sharma, M.D. *et al.* (2016). Local
487 adaptation of a bacterium is as important as its presence in structuring a natural
488 microbial community. *Nat. Commun.*, 7, 12453.

489 Gomulkiewicz, R. & Holt, R.D. (1995). When does evolution by natural selection prevent
490 extinction? *Evolution*, 49, 201-207.

491 Hendry, A.P. (2016). *Eco-evolutionary dynamics*. Princeton University Press, Princeton and
492 Oxford.

493 Hendry, A.P. (2019). A critique for eco-evolutionary dynamics. *Funct. Ecol.*, 33, 84-94.

494 Hillebrand, H., Durselen, C.D., Kirschtel, D., Pollingher, U. & Zohary, T. (1999). Biovolume
495 calculation for pelagic and benthic microalgae. *J. Phycol.*, 35, 403-424.

496 Hiltunen, T., Cairns, J., Frickel, J., Jalasvuori, M., Laakso, J., Kaitala, V. *et al.* (2018). Dual-
497 stressor selection alters eco-evolutionary dynamics in experimental communities. *Nat.*
498 *Eco. Evo.*, 2, 1974-1981.

499 Johansson, J. (2008). Evolutionary responses to environmental changes: How does
500 competition affect adaptation? *Evolution*, 62, 421-435.

501 Kleynhans, E.J., Otto, S.P., Reich, P.B. & Vellend, M. (2016). Adaptation to elevated CO₂ in
502 different biodiversity contexts. *Nat. Commun.*, 7, 12358.

503 Lachapelle, J. & Bell, G. (2012). Evolutionary rescue of sexual and asexual populations in a
504 deteriorating environment. *Evolution*, 66, 3508-3518.

505 Lande, R. (1979). Quantitative genetic analysis of multivariate evolution, applied to
506 brain:body size allometry. *Evolution*, 33, 402-416.

507 Latta, L.C., Weider, L.J., Colbourne, J.K. & Pfrender, M.E. (2012). The evolution of salinity
508 tolerance in *Daphnia*: a functional genomics approach. *Ecol. Lett.*, 15, 794-802.

509 Lau, J.A. & Terhorst, C.P. (2015). Causes and consequences of failed adaptation to biological
510 invasions: the role of ecological constraints. *Mol. Ecol.*, 24, 1987-1998.

511 Lindsey, H.A., Gallie, J., Taylor, S. & Kerr, B. (2013). Evolutionary rescue from extinction is
512 contingent on a lower rate of environmental change. *Nature*, 494, 463-467.

513 Lohbeck, K.T., Riebesell, U. & Reusch, T.B.H. (2012). Adaptive evolution of a key
514 phytoplankton species to ocean acidification. *Nat. Geosci.*, 5, 346-351.

515 Low-Décarie, E., Fussmann, G.F. & Bell, G. (2011). The effect of elevated CO₂ on growth
516 and competition in experimental phytoplankton communities. *Glob. Change Biol.*, 17,
517 2525–2535.

518 Lynch, M., Bobay, L.M., Catania, F., Gout, J.F. & Rho, M. (2011). The repatterning of
519 eukaryotic genomes by random genetic drift. *Annu Rev Genomics Hum Genet*, 12,
520 347-366.

521 Matthews, B., Aebischer, T., Sullam, K.E., Lundsgaard-Hansen, B. & Seehausen, O. (2016).
522 Experimental evidence of an eco-evolutionary feedback during adaptive divergence.
523 *Current Biology*, 26, 483-489.

524 Matthews, B., Narwani, A., Hausch, S., Nonaka, E., Peter, H., Yamamichi, M. *et al.* (2011).
525 Toward an integration of evolutionary biology and ecosystem science. *Ecol. Lett.*, 14,
526 690-701.

527 Mee, J.A. & Yeaman, S. (2019). Unpacking conditional neutrality: genomic signatures of
528 selection on conditionally beneficial and conditionally deleterious mutations. *Am.*
529 *Nat.*, 194, in press.

530 Osmond, M.M. & de Mazancourt, C. (2013). How competition affects evolutionary rescue.
531 *Phil. Trans. R. Soc. B*, 368, 20120085.

532 Oziolor, E.M., Reid, N.M., Yair, S., Lee, K.M., VerPloeg, S.G., Bruns, P.C. *et al.* (2019).
533 Adaptive introgression enables evolutionary rescue from extreme environmental
534 pollution. *Science*, 364, 455-457.

535 Pantel, J.H., Duvivier, C. & De Meester, L. (2015). Rapid local adaptation mediates
536 zooplankton community assembly in experimental mesocosms. *Ecol. Lett.*, 18, 992-
537 1000.

538 Phan, K. & Ferenci, T. (2013). A design-constraint trade-off underpins the diversity in
539 ecologically important traits in species *Escherichia coli*. *Isme J.*, 7, 2034-2043.

540 R Core Team (2015). R: A language and environment for statistical computing. R foundation
541 for statistical computing. Vienna, Austria. URL <http://www.R-project.org/>.

542 Reid, N.M., Proestou, D.A., Clark, B.W., Warren, W.C., Colbourne, J.K., Shaw, J.R. *et al.*
543 (2016). The genomic landscape of rapid repeated evolutionary adaptation to toxic
544 pollution in wild fish. *Science*, 354, 1305-1308.

545 Schaum, C.E., Buckling, A., Smirnoff, N., Studholme, D.J. & Yvon-Durocher, G. (2018).
546 Environmental fluctuations accelerate molecular evolution of thermal tolerance in a
547 marine diatom. *Nat. Commun.*, 9, 1719.

548 Stapley, J., Feulner, P.G.D., Johnston, S.E., Santure, A.W. & Smadja, C.M. (2017). Variation
549 in recombination frequency and distribution across eukaryotes: patterns and processes.
550 *Phil. Trans. R. Soc. B*, 372, 20160455.

551 Stuart, S.N., Chanson, J.S., Cox, N.A., Young, B.E., Rodrigues, A.S.L., Fischman, D.L. *et al.*

552 (2004). Status and trends of amphibian declines and extinctions worldwide. *Science*,

553 306, 1783-1786.

554 terHorst, C.P., Lennon, J.T. & Lau, J.A. (2014). The relative importance of rapid evolution for

555 plant-microbe interactions depends on ecological context. *Proc. R. Soc. B*, 281,

556 20141615.

557 Urban, M.C., De Meester, L., Vellend, M., Stoks, R. & Vanoverbeke, J. (2012). A crucial step

558 toward realism: responses to climate change from an evolving metacommunity

559 perspective. *Evol. Appl.*, 5, 154-167.

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561

562 **Table 1: Community assay.** Results (i.e. P-values) of linear mixed models testing for effects
563 of transfer (T), assay environment (A), and salt history (H) on the abundances of the six
564 species within communities assembled either from control or selection lines. Transfer was
565 treated as a continuous variable. Abundances of species were ln-transformed prior to analysis,
566 P < 0.05 in bold.

567

Species	Transfer	Assay	History	A × H	T × A	T × H	T × A × H
<i>Anabaena</i>	< 0.001	< 0.001	0.807	0.538	0.048	0.770	0.822
<i>Navicula</i>	< 0.001	0.836	0.947	0.749	0.010	0.128	0.664
<i>Nitzschia</i>	< 0.001	< 0.001	0.285	0.031	0.004	0.151	0.648
<i>Pseudokirchneriella</i>	< 0.001	< 0.001	0.100	0.005	0.502	0.450	0.843
<i>Scenedesmus</i>	< 0.001	< 0.001	0.682	0.017	0.005	0.901	0.189
<i>Synechococcus</i>	< 0.001	0.289	0.781	0.545	0.328	0.611	0.633

568

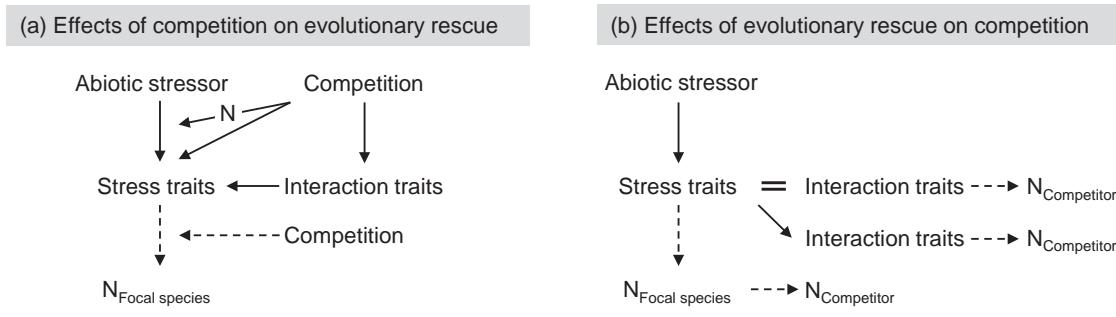


Figure 1: Conceptual figure of how evolutionary rescue from an abiotic stressor (a) is affected by competition, and (b) affects competition. Evolutionary effects are denoted with solid lines, ecological effects with dashed lines. Stress traits indicate traits that mediate the response to a stressor, interaction traits indicate traits that mediate competitive interactions. Evolutionary rescue from a stressor occurs when the focal species responds with the evolution of stress traits such that population size (N) increases and the population persists. (a) Competition can influence adaptation to an abiotic stressor by reducing population size and thus genetic variation of the focal species, by imposing selection on stress traits, or by imposing selection on interaction traits that are genetically correlated with stress traits. Competition could also influence evolutionary rescue by altering the effect of adaptation on population size. (b) Evolutionary rescue can affect competition when stress traits simultaneously act as interaction traits, when stress traits are correlated with interaction traits, or when increased population size of the focal species influences the competitor.

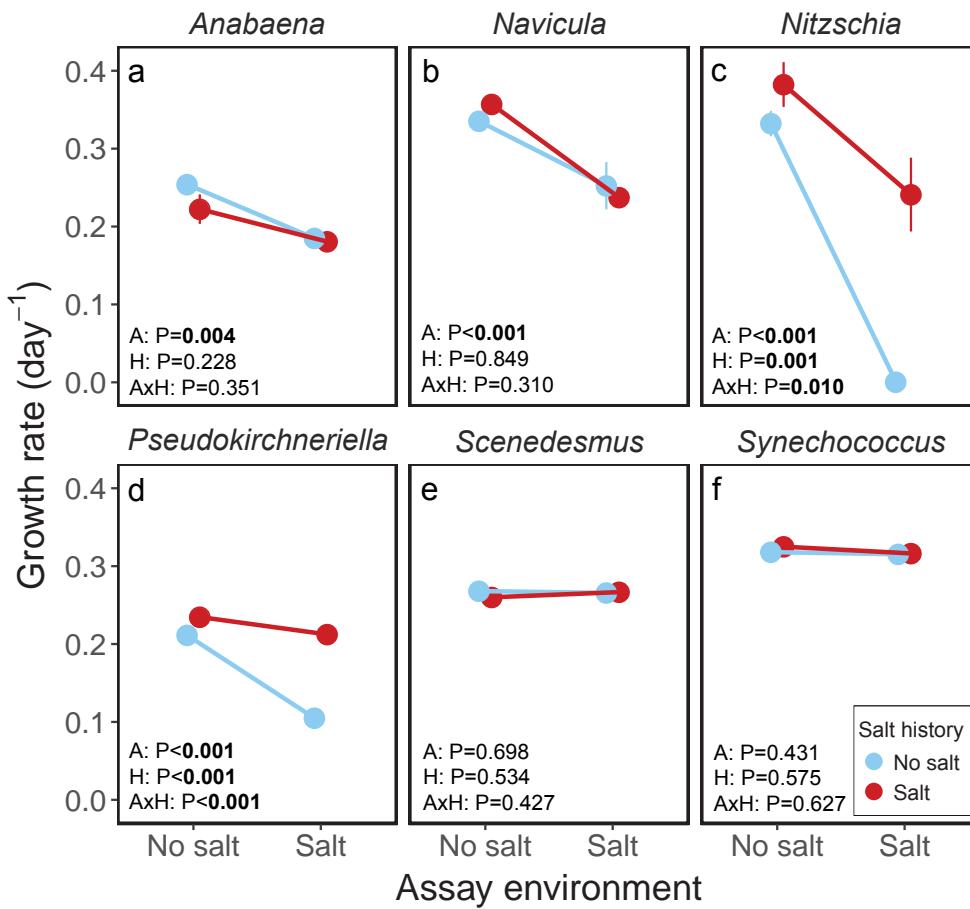


Figure 2: Effects of salt history and assay environment on the monoculture growth rates of six species of microalgae. Control lines (blue) have a salt history of 37 transfers at 0 g NaCl L^{-1} , selection lines (red) have a salt history of 37 transfers at increasing salt stress with a final concentration of 6 g NaCl L^{-1} . Salt concentrations in the two assay environments were the same as in the final selection environments, i.e. 0 and 6 g NaCl L^{-1} , respectively. Insets give the P-values for assay environment (A), salt history (H), and their interaction (AxH). Bold font denotes P-values < 0.05 . Values are means \pm SE, $n = 3$.

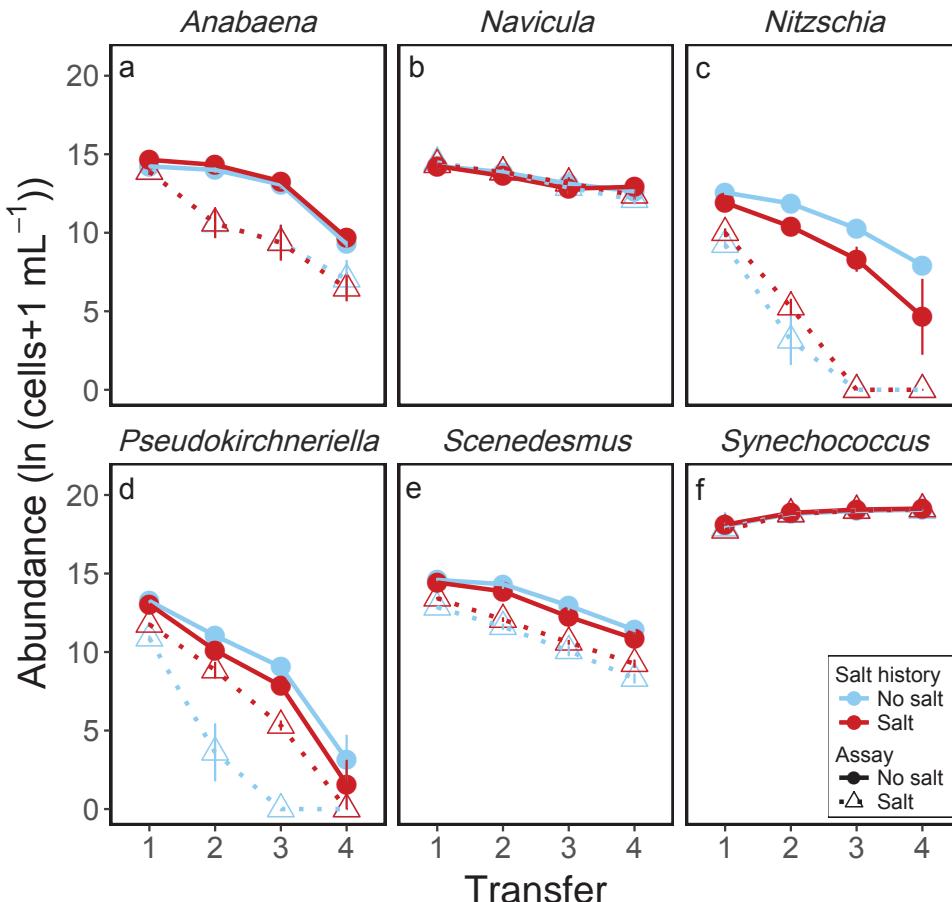


Figure 3: Effects of salt history and assay environment on species abundances in the community assay. Communities of six species were assembled either from control lines (blue) or from selection lines (red). Species abundances were quantified over four transfers in assay environments without salt (solid circles, solid lines) and with salt (open triangles, dotted lines), respectively. Values are means \pm SE, $n = 3$.

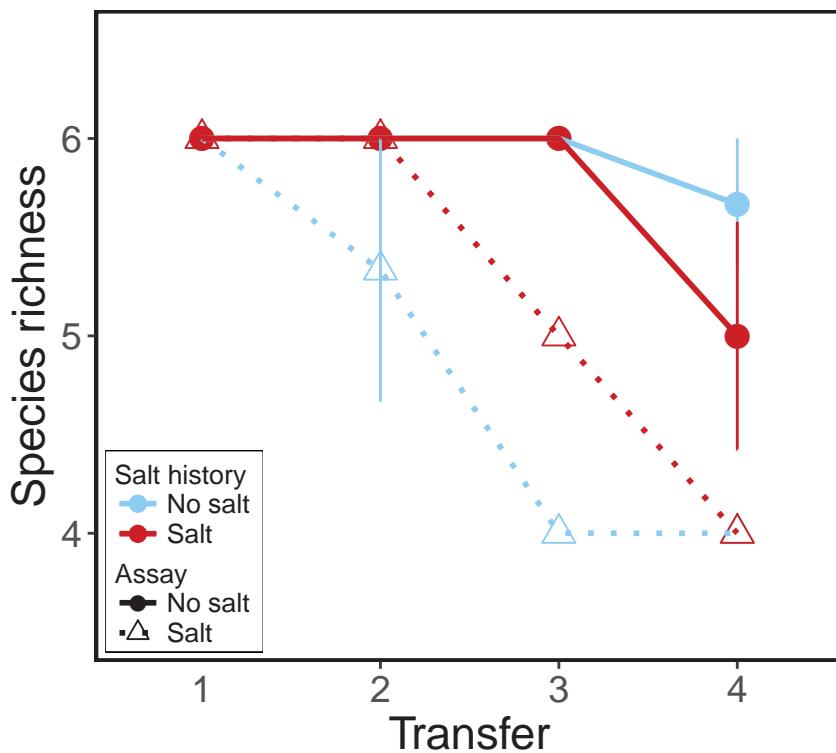


Figure 4: Effects of salt history and assay environment on species richness in the community assay. Communities of six species were assembled either from control lines (blue) or from selection lines (red). Species richness was quantified over four transfers in assay environments without salt (solid circles, solid lines) and with salt (open triangles, dotted lines), respectively. Values are means \pm SE, $n = 3$.

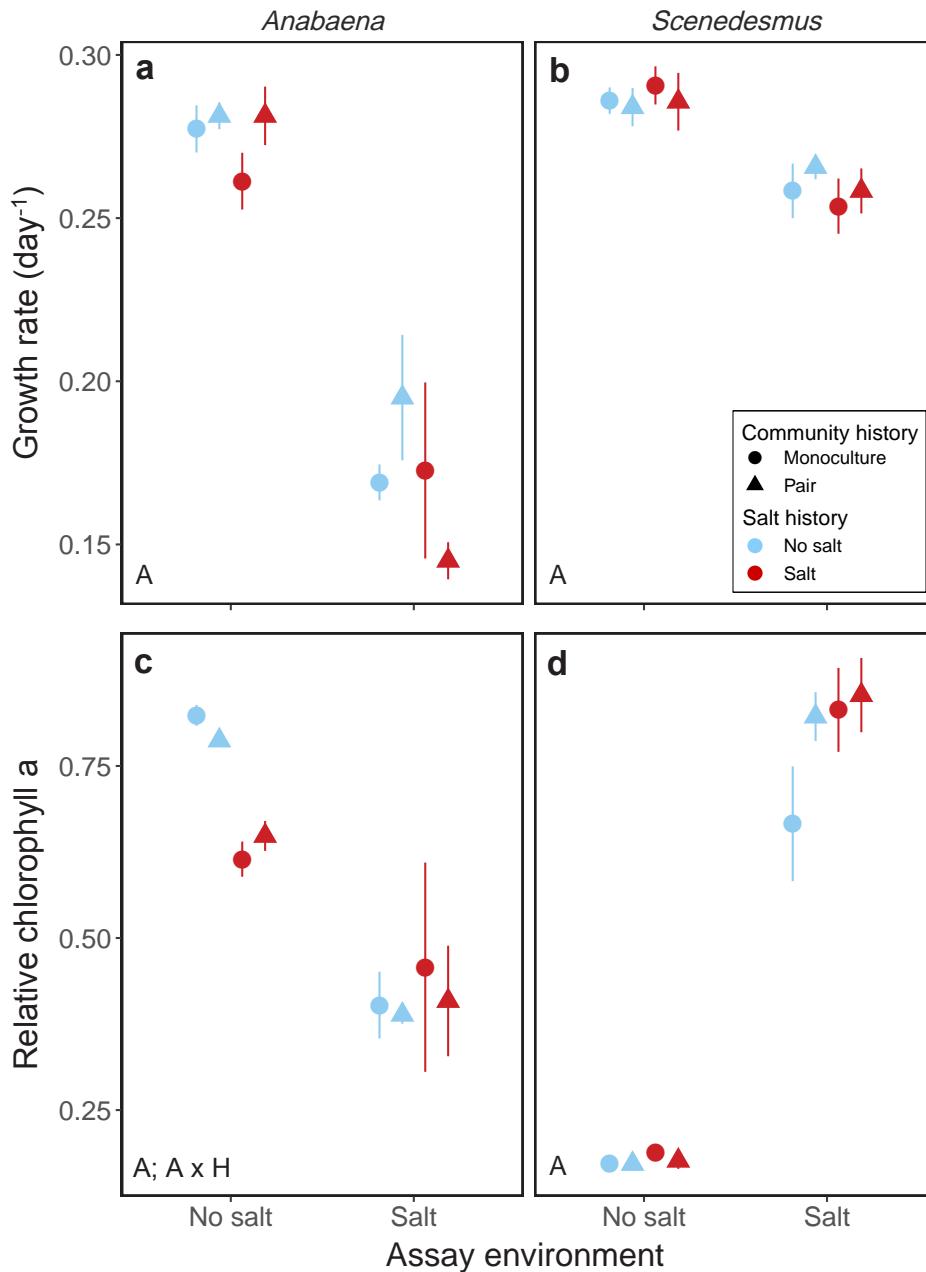


Figure 5: Effects of community history, salt history, and assay environment on the growth rates and competitive abilities of *Anabaena* and *Scenedesmus*. Monoculture growth rate of (a) *Anabaena* and (b) *Scenedesmus*, relative chlorophyll a of (c) *Anabaena* in competition with a reference line of *Scenedesmus* and (d) of *Scenedesmus* in competition with a reference line of *Anabaena*. Inserts indicate significant effects ($P < 0.05$) of community history (C), salt history (H), assay environment (A), or their interactions. Values are means \pm SE, $n = 3$.