

# 1 **Demographic history, linked selection, and recombination shape the genomic landscape of a broadly** 2 **distributed Pacific salmon.**

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## 27 **Abstract**

28 Understanding the impacts of current human activities on within-species genetic variation requires a  
 29 thorough description of the historical factors that have shaped the genomic and geographical distribution of  
 30 nucleotide diversity. Past and current conditions influencing effective population size have important  
 31 evolutionary implications for the efficacy of selection, increased accumulation of deleterious mutations, and  
 32 loss of adaptive potential under the nearly neutral theory. Here, we gather extensive genome-wide data that  
 33 represent the extant diversity of the Coho salmon (*Oncorhynchus kisutch*) to address three issues. First, we  
 34 demonstrate that a single glacial refugium is the source of the majority of present-day genetic diversity, with  
 35 minor but detectable inputs from secondary micro-refugia. We propose a scenario whereby several ancestral  
 36 populations located south of the ice sheets expanded in postglacial time, swamping out most of the diversity  
 37 from other putative micro-refugia. Following this expansion, we identify particular populations having  
 38 undergone continuous declines in population size ( $N_e$ ). Second, we combine multiple evidence from  
 39 demographic modelling, analysis of recombination landscape, and genome-wide landscape of diversity to  
 40 demonstrate that selection at linked sites and Hill-Robertson interference played a major role in shaping  
 41 genetic diversity across the Coho salmon genome. Third, we demonstrate that this demographic history  
 42 generated subtle differences in the load of deleterious mutations among populations, a finding that mirrors  
 43 recent results from human populations. Taken together, we found considerable support for the joint  
 44 contributions of demographic history and linked selection in the load of deleterious mutations. We suggest  
 45 that these inferences should be better integrated in conservation genetics of managed fish species which  
 46 currently focuses largely on within-population adaptation.

## 48 **Author Summary**

49 Reconstruction of a species' past demographic history from genome-wide data allows understanding how  
 50 historical factors interact with intrinsic genomic properties to shape the distribution of genetic diversity along  
 51 its genome and its geographic range. Here, we combine genotyping-by-sequencing and whole genome  
 52 sequence data with demographic modelling to address these issues in the Coho salmon, a Pacific salmon  
 53 species with rapidly declining census size in some parts of its range, notably in the south. Our demographic  
 54 reconstructions indicate a linear decrease in genetic diversity towards the north of the species range,  
 55 supporting the hypothesis of a major southern refugia for the Coho salmon and a northern route of  
 56 postglacial recolonization. Accordingly, the number of candidate deleterious homozygous derived mutations  
 57 was higher in northern populations. Demographic modelling also suggested the existence of cryptic refugia  
 58 that may have been missed with the use of simpler summary statistics. We further showed that the species'  
 59 genome was shaped by linked selection and biased gene conversion. In particular, local variation in  
 60 recombination rates have modulated the efficacy of natural selection. These processes, together with a  
 61 complex demographic history, can contribute to the load of deleterious mutations – an effect we argue should  
 62 be taken into account more routinely in conservation genetics studies.

## 63 Introduction

64 Both plant and animal biodiversity are currently declining or disappearing at unprecedented rates due to  
65 human activity [1]. This leads to population size reduction, reduced genetic diversity, and to the sixth mass  
66 extinction [2]. Before humans became major drivers of changes in species distributions and abundance,  
67 long-term climate change had a major influence [3]. The Pleistocene glaciations resulted in major  
68 contractions in the geographical distributions of many species into refugia that persisted in unglaciated areas  
69 [4]. Postglacial range expansions often led to contacts between ancestral populations previously segregated  
70 in different refugia [4,5]. The effects of long-term climate change combined with recent human-induced  
71 population declines can foster genetic changes including a loss of genetic diversity, increased inbreeding,  
72 increased load of deleterious mutations, and a loss of local adaptation [6].

73 In this context, it becomes important to understand how demographic history interacts with past and  
74 ongoing selection and recombination to shape genetic variation. By disentangling past and current drivers of  
75 range-wide genomic diversity, this information can inform management and conservation decisions [7].  
76 Beyond conservation implications, such context provides a unique opportunity to address outstanding  
77 questions in evolutionary biology. In particular, what is the role of gene flow in shaping heterogeneous  
78 differentiation landscape during population divergence [8,9]? What are the demographic conditions required  
79 to generate substantial differences in deleterious load among populations [10]?

80 A major challenge to understanding drivers of genome-wide patterns of diversity is that different  
81 demographic processes can lead to similar contemporary genomic footprints [11]. As populations diverge,  
82 they accumulate genetic incompatibilities forming barriers to gene flow [12], while the rest of the genome  
83 may continue to be freely exchanged. As a consequence, the genomic landscape of divergence is expected to  
84 vary, with greater differences between populations at genomic barriers as compared to genomic regions  
85 exhibiting ongoing gene flow. However, similar patterns of heterogeneous genome-wide divergence can be  
86 due to genetic hitchhiking of neutral alleles linked to selective sweeps [13] or to background selection (BGS;  
87 [14]). These combined effects, referred to as linked selection reduces polymorphism at sites closely linked to  
88 advantageous or deleterious variants, and therefore reduces local effective population size ( $N_e$ ) along the  
89 genome. The intensity of selection on linked loci will be mostly modulated by variation in local  
90 recombination rate and by gene density [15,16]. Under linked selection, diversity ( $\pi$ ,  $D_{XY}$ ) and differentiation  
91 ( $F_{ST}$ ) metrics are expected to be positively and negatively correlated with genome-wide variation in  
92 recombination rate ( $\rho$ ) respectively, and with the density of targets (e.g., genes, regulatory regions) that are  
93 subject to selection [14–18]. It is now increasingly recognized that neglecting BGS can bias demographic  
94 inferences [19,20] or lead to false adaptive interpretations [21].

95 An understanding of historical demography is also essential for a sound interpretation of patterns of  
96 deleterious mutation load observed among contemporary populations [10,22]. Population bottlenecks are  
97 predicted to reduce potential for local adaptation, but also to reduce standing genetic variation and the  
98 efficacy of selection [23,24]. In turn, a reduced efficacy of purifying selection leads to an increase in the  
99 number of deleterious variants segregating in a population. Moreover, intrinsic genome properties, in  
100 particular local variation in recombination rate or background selection, can favour the accumulation of

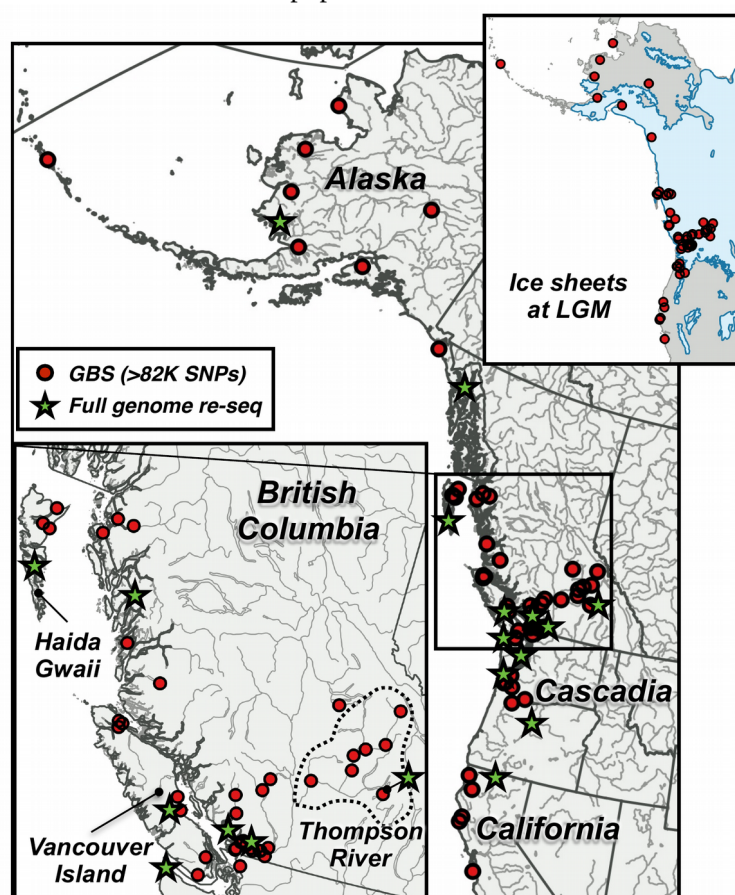
deleterious mutations [25]. From a conservation standpoint, populations harboring an elevated number of deleterious variants might need to be monitored more closely.

Combining population genomics data with demographic modelling represents a powerful strategy to test alternative hypotheses about historical drivers of existing genomic diversity. Previous studies employing a similar approach have focussed mostly on species with a narrow geographic range, such as small islands [26,27], which are on the verge of extinction [28–31] are strongly bottlenecked [32], shedding light on the evolutionary consequences of small population size. Few studies, however, have investigated how historical processes have shaped the geographical patterns in the distribution of genomic diversity in more broadly distributed species, e.g., at the scale of a whole continent. An exception to this observation is the vast literature on demographic reconstructions of human populations. Long-lasting debates in this literature regarding the role of demography in generating mutation load differences among populations [22,33,34] could benefit from studies of species displaying similarly complex demographic histories and broad geographic distributions.

Salmonid fishes are economically important species that have suffered recent demographic declines [35,36]. This is particularly the case for Coho salmon (*Oncorhynchus kisutch*), one of the five anadromous species of Pacific salmon that supports important recreational and indigenous subsistence fisheries, which has suffered dramatic population declines (> 90%) over the last three decades in parts of its range [36,37]. A previous study investigated the range-wide population structure and demographic history of the species and found a cline of decreasing diversity from south to north, as well as some endemic diversity in small putative refugia [38] (see also [39]). This study indicated that Coho salmon may have survived the last glacial maximum (LGM, i.e. the Fraser Glaciation in British Columbia, and the McConnell/McCauley Glaciation in Yukon and Alaska; 23 to 18 Ky ago) in unglaciated areas of Haida Gwaii and Beringia in addition to areas south of the ice sheets. This study, however, predates the genomic era and could not eliminate alternative hypotheses regarding the origin and number of glacial refugia during the LGM. Most importantly, the impacts of confounding factors such as background selection, recombination rate variation, and how these factors may facilitate the accumulation of deleterious mutations could not be studied with the limited number of genetic markers available at the time. In North America, the species is currently distributed from California to Alaska [40]. Unglaciated areas that could potentially serve as glacial refugia persisted both north (e.g. the Beringian refugium in Alaska, the Yukon Territory of Canada and areas of Asia and the Bering Land Bridge) and south (e.g. all of the deglaciated area south of British Columbia, Canada) of the ice sheet [40–42]. Other unglaciated areas (e.g. Haida Gwaii in British Columbia) could also have been micro-refugia [43,44]. In this context, distinct demographic scenarios can be tested. Under a first scenario whereby populations expanded north from a single southern refugium, we predict: *i*) a latitudinal decrease in genetic diversity from south to north along with a pattern of IBD, and *ii*) ancestral populations located in areas south of the ice sheets. Under a second scenario, populations expanded south from a single northern refugium, and we predict the opposite geographic pattern. The third scenario corresponds to the survival of populations in different refugia where we predict: *i*) the existence of clearly distinct genetic clusters, and *ii*) postglacial gene

138 flow with signatures of secondary contacts, with contact zones displaying higher genetic diversity through  
139 postglacial admixture between different genomic backgrounds.

140 In order to test these alternative scenarios, we generated genome-wide data from nearly 2,000 Coho  
141 salmon from California to Alaska, one of the most extensive genomic datasets for a non-model vertebrate  
142 species to date. First, to resolve the species demographic history, we used a modelling approach that accounts  
143 both for barriers to gene flow affecting migration locally, and for linked selection affecting the rate of drift.  
144 Next, we tested the above predictions related to linked selection. Finally, we hypothesized that demographic  
145 history and background selection shaped the pattern of deleterious mutation load, both within and among  
146 populations. In particular, we hypothesized that postglacial re-colonisation influenced levels of standing  
147 genetic variation and favoured the accumulation of deleterious mutations at the expansion front. In these  
148 conditions, we predicted that genetic diversity should decrease as a function of the distance from the  
149 ancestral populations, while the accumulation of putatively deleterious mutations should increase as a  
150 function of the distance to the ancestral populations.



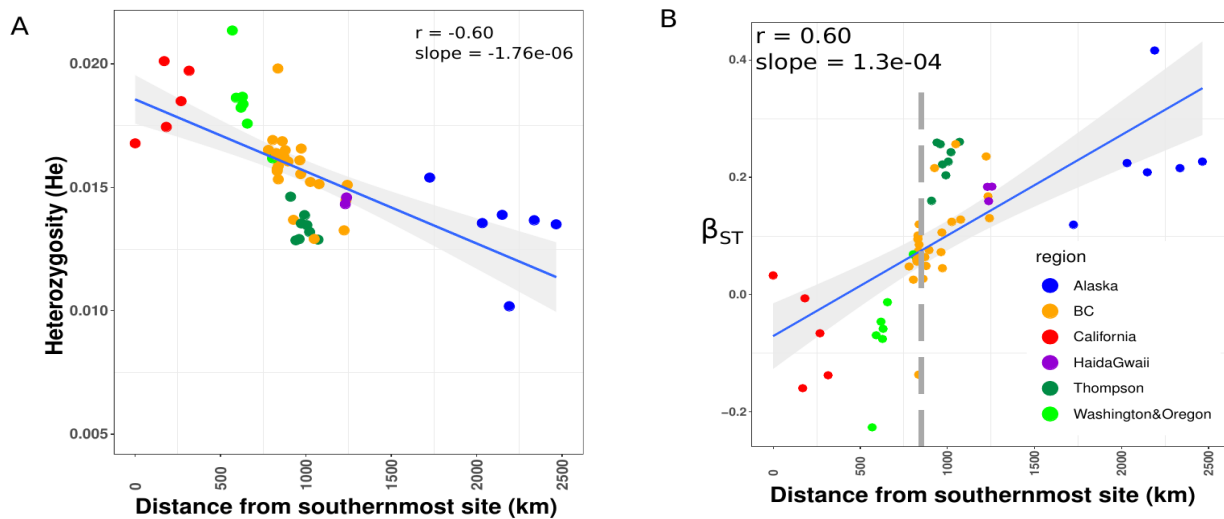
152 **Figure 1) Sampling design** A) Sampling locations of 58 Coho salmon populations distributed across the species' North  
153 American range of distribution. Each dot represents a sampling location. Inset: Map showing the extend of ice-sheet  
154 during the Last Glacial Maximum 13 KyA. Data modified from [115]

## 155 Results

### 156 Overall genetic diversity and population structure

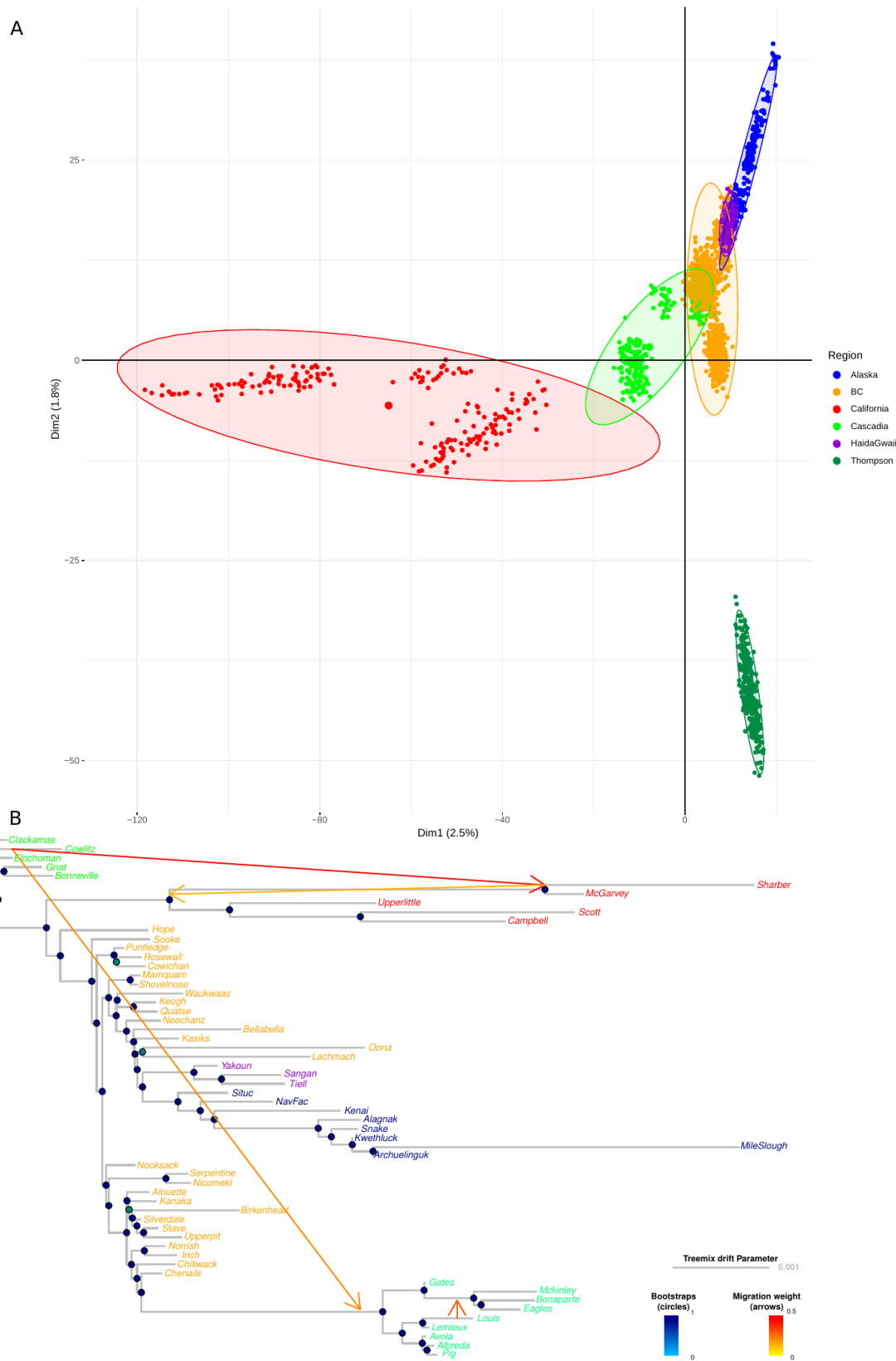
157 A total of 1,957 individuals was sampled from California to Alaska representing 58 sampling locations  
158 (mean  $n = 34$  fish per location, Fig 1a, S1 Table) and genotyped using a genotype by sequencing (GBS)  
159 method that generated 82,772 high quality filtered single nucleotide polymorphisms (SNPs). Another set of  
160 55 individuals representing 11 sampling locations from the same range (Fig 1a), were whole genome  
161 sequenced (WGS ) to ~30X coverage, and used in specific analyses (S2 Table).

162 Levels of genetic diversity (observed and expected heterozygosity,  $\pi_{\text{SNP}}$ ) were highest in formerly  
163 deglaciated areas in the south (California, Cascadia, Fig 2a, Fig 1b) and decreased as a function of distance  
164 from the southernmost site up to Alaska ( $r = 0.64$ ,  $p < 0.0001$ , Fig 2a, S1 Fig). The Thompson River  
165 watershed (Thompson R. hereafter) in southern British Columbia was an exception to this latitudinal pattern  
166 and displayed the lowest average level of regional genetic diversity of all sampling locations which we  
167 hypothesized to results from bottlenecks in this area. The remaining samples from British Columbia were  
168 intermediate in genetic diversity.



### 170 Figure 2) Genetic diversity and differentiation

171 A) Linear relationship between expected heterozygosity and distance from the southernmost population located in  
172 California. B) Linear increase in genetic differentiation as measured by  $\beta_{ST}$  as a function of the distance from the  
173 southernmost population located in California. Negative values indicate the most likely ancestral population. The  
174 relationship in A and B was tested using linear models. The grey vertical bar in panel B) indicates the approximate  
175 location of the southern limit of the ice-sheet at the end of the last glacial maxima.



177 **Figure 3) Genetic structure and gene flow**

178 A) Principal Component Analysis (PCA) summarising population genetic structure among 1,957 individuals based on

179 the principal component axes 1 and axes 2. Each point represents an individual and the colours represent the major

180 regional groups. B) Inference of population splits and mixture by Treemix with 4 statistically significant migration

181 events ( $p < 0.0001$ ). Bootstrap supports and migration weights are indicated according to the colour scale.

The distribution of singletons provided further information regarding the most ancestral populations, with older populations expected to have accumulated a higher density of singletons [45]. Counting the number of singletons by sampling site and averaging by regional groups revealed the following differences: Cascadian samples contained the highest number of singletons, with a mean of 1,263 singletons per site. Californian samples had the fewest number of singletons ( $n_{\text{MEAN}} = 55$ ) while Alaska harbored intermediate density ( $n_{\text{MEAN}} = 966$ ). Consistently, WGS data revealed 2.7 times more singletons in Cascadia (Tsoo-Yess River) as compared to Alaska (Kwethluk River), whereas Thompson R. samples contained 6.7 times less singletons, supporting the hypothesis of a pronounced bottleneck in these populations (S3 Table). Similarly, the occurrence of private alleles was more prevalent among southern than northern populations, being nearly twice as high in Cascadian ( $n = 10,097$ ) than in Alaskan populations ( $n = 5,270$ ). Again, populations from the Thompson R. were an exception to this pattern with the lowest level of private polymorphism ( $n = 1,479$ , S2 Fig).

Comparing the decay of linkage disequilibrium (LD) across populations based on the WGS data indicated rapid LD decay across all samples (S3 Fig). The Thompson R. populations again departed from the general pattern with a much reduced LD decay. Indeed,  $r^2$  values decreased to 0.4 at approximately 342 kb for Thompson R. whereas this  $r^2$  value was attained between 13 and 34 kb for all other populations.

Next we used the  $\beta_{\text{ST}}$  coefficient to identify ancestral population [46]. Unlike  $F_{\text{ST}}$  estimates [47], this index can account for the non-independence among populations and negative values are indicative of ancestral populations [46]. Here,  $\beta_{\text{ST}}$  indicated that ancestral populations were located in previously unglaciated areas corresponding to Cascadia ( $n = 5$  localities), California ( $n = 3$  localities) as well as one site from southern British Columbia (Fig 2b, S4 Table). A linear decrease in  $\beta_{\text{ST}}$  as a function of distance from the southernmost site was observed ( $r = 0.60$ , slope =  $1.03 \times 10^{-4}$ ,  $p < 0.0001$ ) as expected under IBD. Support for this IBD pattern was also observed using  $F_{\text{ST}}$  (S4 Fig,  $r = 0.66$ , slope =  $4 \times 10^{-5}$ ,  $p < 0.0001$ ) as well as Mantel tests ( $r = 0.64$ ;  $p < 0.0001$ ,  $r = 0.72$ ;  $p < 0.0001$  when removing Thompson R. populations). Average pairwise  $F_{\text{ST}}$  across all populations was 0.095 and varied from 0.002 to 0.334 (S5 Fig), indicating moderate population structure, typical of anadromous species connected by gene flow [48].

Model-based analysis of population structure failed to reveal a clear number of distinct populations (K value). Instead, K values ranging from 30 to 60 all fit the data well (S6 Fig), due to the confounding effect of IBD. The first axis of a Principal Components Analysis (PCA, Fig 3a) revealed a separation of the sample from South to North with the most divergent samples found in California. The second axis revealed the divergence from East to West but with a strong separation of the Thompson R. populations. Along these axes populations followed an IBD pattern. These results were also supported by an MDS analysis (S7 Fig). The third and fourth axis did not yield further information (S8 Fig).

We then used Treemix [49] to infer population splits and gene flow (Fig 3b). A first tree assuming no migration (i.e. drift only) explained 98.1% of the variance observed. Adding up to four significant migration events ( $p < 0.0001$ ) explained over 99.1% of variance, then the proportion of explained variance plateaued (S9 Fig). Populations from Cascadia occupied basal positions. California populations displayed pronounced genetic drift, corroborating the high divergence observed in the PCA. Populations from Alaska (MSL River)

220 and Thompson R. also displayed higher genetic drift, in line with evidence based on analyses of genetic  
221 diversity and genetic structure. The two most supported migration events occurred from Cascadia south to  
222 California and north to the Thompson R. We note that populations followed the south to north arrangement,  
223 with the samples from Cascadia displaying less drift than those further north.

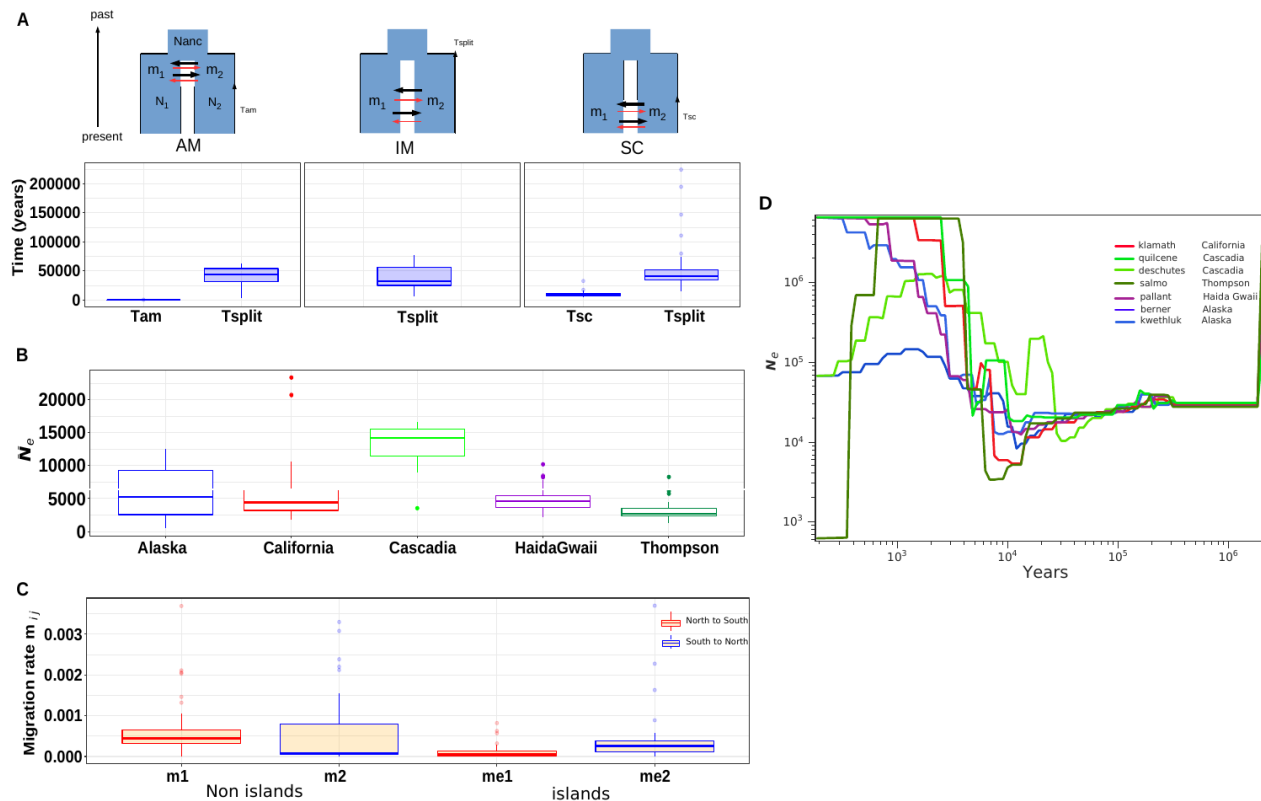
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## 225 *Demographic history*

226 In order to assess more formally the occurrence of one or more refugial origins for contemporary  
227 populations, we performed the following explicit model-based inferences of population divergence scenarios  
228 using  $\partial\text{a}\partial\text{i}$  [50]. Our models account for the confounding effects of selection at linked sites and that of the  
229 accumulation of local barriers to gene flow in the genome [19,51]. Four major demographic models were  
230 statistically compared using groups identified in the PCA and with a focus on previously hypothesized  
231 refugia (i.e., Cascadia, California, Haida Gwaii, Alaska). The following models were tested: strict isolation  
232 (SI model), divergence with ancient migration (AM model), divergence with continuous gene flow (IM  
233 model) and secondary contact (SC model; S10 Fig). A total of 69 pairwise comparisons was performed with  
234  $\partial\text{a}\partial\text{i}$ . In each pairwise comparison one single representative population from a putative refugium was  
235 compared against one population from another refugium. Due to the high local structure within groups, we  
236 avoided pooling samples into higher order groups (e.g., regional groups) as this would unavoidably bias our  
237 results. Models incorporating linked selection and restricted introgression along the genome always received  
238 highest support (S5 Table).  $\Delta\text{AIC}$  (minimum value  $> 5$ ) confidently discriminated models in 87% of the  $\partial\text{a}\partial\text{i}$   
239 comparisons (S5 Table). The SC model received the highest support in 46% of the comparisons, the AM  
240 model in 30%, and the IM model in 14%, with 13% remaining unclassified. Scenarios including periods of  
241 gene flow clearly outperform scenarios assuming no gene flow. The fact that secondary contact was the best-  
242 supported model suggests that more than one glacial refugia contributed to the recolonization of the  
243 contemporary range occupied by Coho salmon.

244 Assuming a generation time of 3.5 years [52] and mutation rate of  $8\text{e}^{-9}$  bp/generation revealed  
245 similar divergence time estimates among models (i.e., 57 Kya under SC, 40 Kya under IM, and 42 Kya under  
246 AM, S6 Table). The median time of secondary contact (SC) was 10 Kya [min = 4,800 – max = 32,900],  
247 corresponding roughly to the onset of the last glacial retreat (Fig 4A). Parameter estimates under the AM  
248 model supported a very recent reduction in gene flow (mean  $< 1\text{Kya}$ ), which therefore represents a  
249 demographic model similar to an IM model. Pronounced variation in effective population size ( $N_e$ ) was  
250 observed with the highest values in Cascadia populations (mean 12,800; range [3,500 – 16,500]). On  
251 average, the smallest  $N_e$  were observed in the Thompson R. populations (mean 3,000; range [1,300 – 8,300],  
252 Fig 4B), again consistent with a population bottleneck. However, smaller  $N_e$  values were also observed in  
253 the isolated Navfac Creek population in Alaska ( $N_e = 500$ ) and the Scott Creek population in California ( $N_e$   
254 = 1,800), which are the most divergent populations identified in the PCA. The McGarvey (California) and  
255 Clackamas (Cascadia) populations identified as “ancestral” according to the  $\beta_{\text{ST}}$  displayed the highest  $N_e$   
256 (23,000 and 16,500). We also note that incorporating linked selection and barriers to gene flow further  
257 reduced  $N_e$ , regardless of the model (S11 Fig). Intrinsic barriers to gene flow reduced the estimated

258 migration rate by half of its value and affected 20 to 40% of the genome (Fig 4C, S6 Table). Similarly, the  
259 Hill-Robertson factor suggested that  $N_e$  was reduced to 37%-50% of its initial estimate and that  
260 approximately half of the genome was affected (S6 Table).



**Figure 4) Inferences of demographic history using  $\partial\text{a}\partial\text{i}$  and  $\text{smc}++$**

262 A) Estimates of divergence time (in years) between each major region as inferred by  $\partial\text{a}\partial\text{i}$  under the best model  
263 (displayed in blue) based on the GBS SNP data set. B) Estimates of effective population size  $N_e$  for each major region  
264 as inferred by  $\partial\text{a}\partial\text{i}$  under each best demographic model based on the GBS SNP data set. C) Estimates of migration rate  
265 among populations in neutral regions ( $m$ ) of the genome and in areas of restricted recombination ( $me$ ) based on the  
266 GBS SNP data set. D) Estimates of effective population size ( $N_e$ ) change through time (in years) for the whole genome  
267 data using  $\text{SMC}++$ .

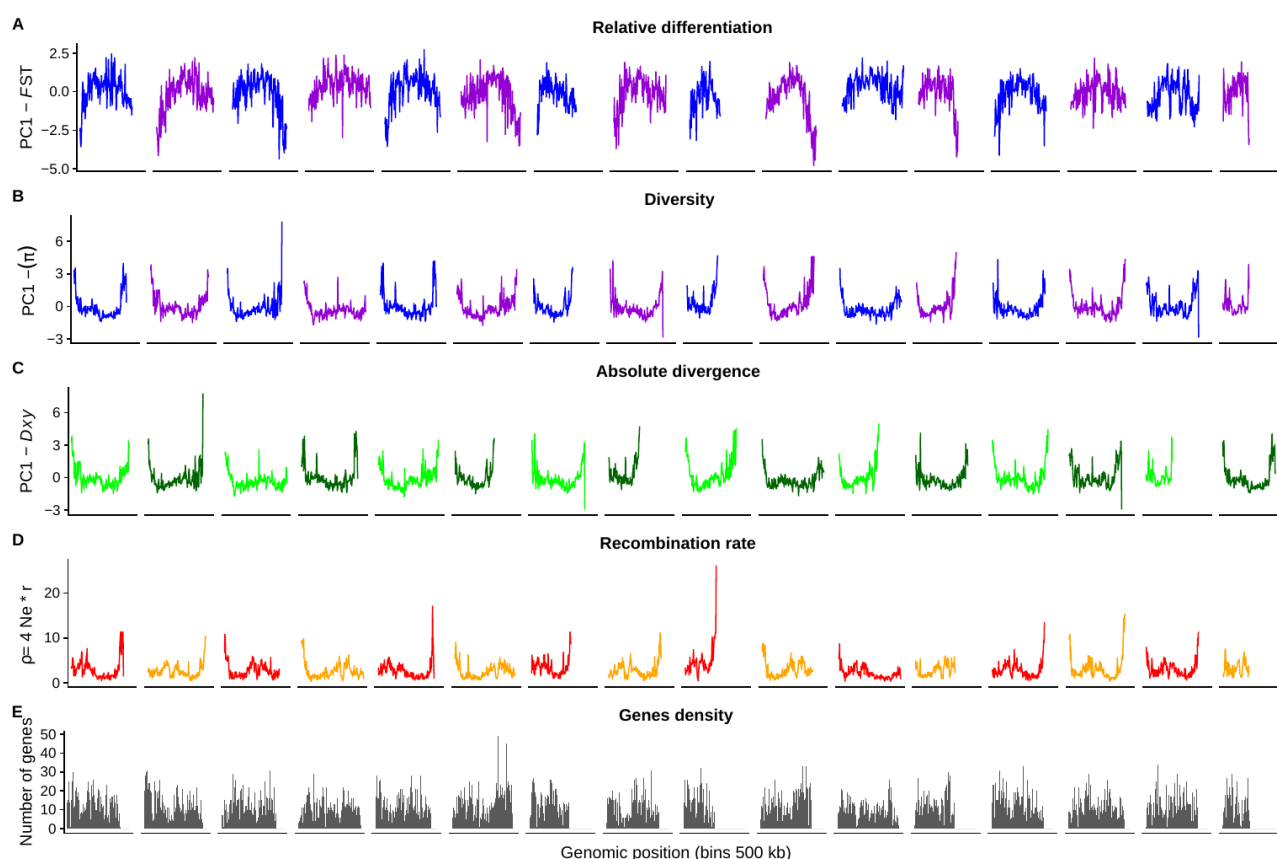
269 We then investigated historical change in effective population size using the Sequentially Markovian  
270 Coalescent in  $\text{SMC}++$  [54] and the 55 whole genome sequences, which are representative of 11 populations  
271 from California to Alaska (Fig 1, S2 Table). For ease of visualization, results for a subset of the populations  
272 are presented in Fig 4D and details for all populations are displayed in S12 Fig. This analysis revealed: i) an  
273 expansion of most populations approximately 12-20 KyA, concomitant with postglacial re-colonisation, ii) a  
274 slow and steady decline in the Thompson R. (Fig 4D, S12 Fig), and iii) a split time between all pairwise  
275 combinations of populations (median = 16,3 KyA, range = 6,7KyA - 61KyA, S13 Fig) compatible with the  
276 onset of postglacial population expansion (Fig 4D) which was accompanied by an increase in  $N_e$  across  
277 samples (Fig 4D). We note that using a different mutation rate ( $1,25\text{e-}8$  mutations/bp/generations) yielded  
278 estimates of split time that were more in line with the estimates of postglacial expansion (median = 9,6 KyA,  
279

280 S13 Fig) and with less variance among compared population (min = 5 KyA – max = 18 KyA). Overall,  
 281 SMC++ results are consistent with previous analyses indicating that until recently, all populations shared a  
 282 similar demographic history. We caution against a strict interpretation of the most recent time. Indeed,  
 283 reliable estimates for present time require a high number of samples, whereas we only had five samples per  
 284 populations [54]. Similarly, exact  $N_e$  estimates should be interpreted cautiously and only the trend should be  
 285 considered rather than the exact values.

286

287 *Linked selection shape the Coho salmon genomic landscape*

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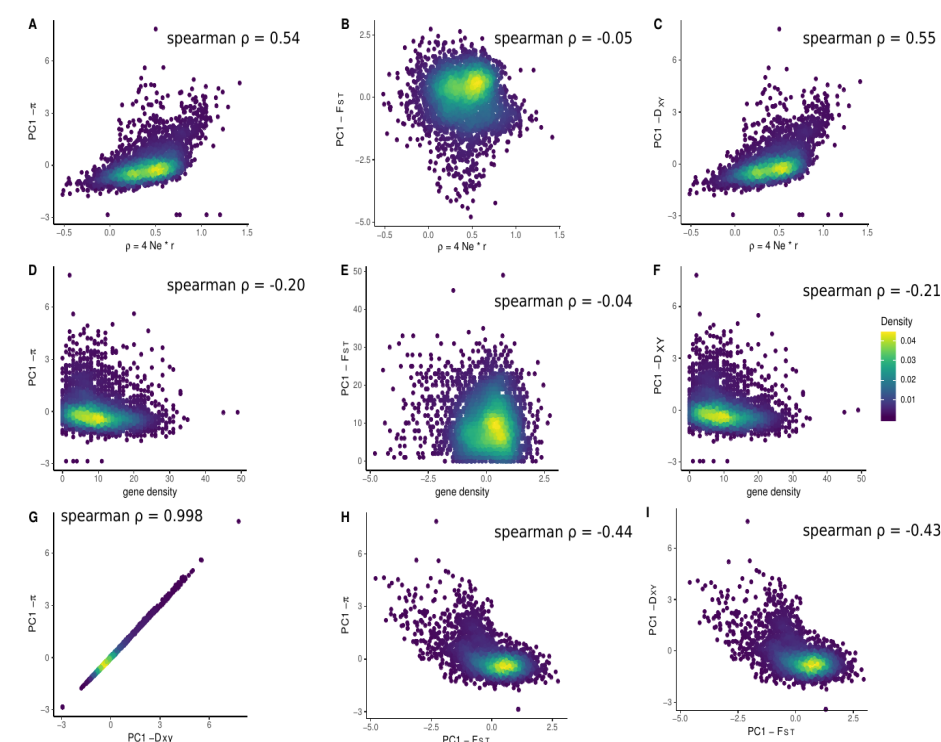
290 **Figure 5) Genome-wide landscapes in Coho salmon with:** A) landscape of differentiation, B) landscape of genetic  
 291 diversity C) divergence, D) recombination and E) gene density. Plots are averaged over 500-kb windows. Only the first  
 292 16 chromosomes are displayed for simplicity. The same figure with all chromosomes is available in Fig. S15.

293

294 We first described patterns of recombination and then tested correlations among genetic diversity ( $\pi$ ),  
 295 divergence ( $D_{XY}$ ), and differentiation ( $F_{ST}$ ) parameters measured across the whole genome using 500 kb  
 296 sliding windows (Fig 5 A-C) and the population-scaled landscape of recombination ( $\rho$ ) (Fig 5D) and gene  
 297 density (Fig 5E) to test our predictions using whole genome sequences.

298 This analysis first revealed a heterogeneous recombination landscape that varied both within and  
 299 among chromosomes. In particular, recombination rate was higher towards the ends of chromosomes (Fig  
 300 5D), as observed across many species [55] Second, we observed a negative correlation between  
 301 recombination rate and chromosome length ( $R^2 = 0.481$ ,  $p < 0.0001$ , S14 Fig). To investigate the effects of

selection at linked sites, we summarised  $\pi$  (across all 11 WGS populations),  $D_{XY}$  and  $F_{ST}$  statistics (55 pairwise comparisons), using the first axis of a PCA (PC1) to obtain the common variation in these metrics and reduce the dimensionality to single summary statistic (Fig 6, S15 Fig, S7 Table). PC1 captured 99% of variance in  $\pi$  and in  $D_{XY}$ , indicating that the PC1 was effective at summarizing information about diversity and divergence. All linear correlations among the tested variables were significant ( $p < 0.0001$ , S8A Table). We found a positive correlation between PC1- $\pi$  and recombination ( $r = 0.54$ , S8A Table) and a negative correlation between PC1- $\pi$  and gene density ( $r = -0.20$ ) (Fig 6A-D). We also found a positive correlation between PC1- $D_{XY}$  and recombination ( $r = 0.55$ ) and a negative correlation between PC1-  $D_{XY}$  and gene density ( $r = -0.21$ ). PC1- $F_{ST}$  was negatively, albeit weakly correlated with both recombination ( $r = -0.05$ ) and gene density ( $r = -0.04$ , Fig 6B-E and C-F). We observed a correlation of 0.999 between  $\pi$  and  $D_{XY}$ , as expected here because of a very recently shared common ancestor (Fig 6G). We also found a negative relationship between PC1-  $\pi$  and PC1-  $F_{ST}$  ( $r = -0.44$ , Fig 6H) as well as between PC1-  $D_{XY}$  and PC1- $F_{ST}$  ( $r = -0.43$ , Fig 6I). The smaller correlations between statistics involving  $F_{ST}$  can be explained by the modest variance in  $F_{ST}$  explained by PC1 (37%).



**Figure 6) Correlations among different metrics.**

Correlations between recombination rate ( $\rho = 4Ne.r$ ) genetic diversity ( $\pi$ ), divergence ( $D_{xy}$ ) and differentiation ( $F_{ST}$ ) statistics (A-C), and correlation between gene density and  $\pi$ ,  $D_{xy}$ ,  $F_{ST}$  statistics (D-F). Correlations between  $\pi$ ,  $D_{xy}$  and  $F_{ST}$  statistics (G-I). The Spearman correlation values are plotted but mixed linear models were fitted to further test interaction among explanatory variables (recombination rate and gene density).

In order to investigate further the effects of linked selection, we used mixed linear models that integrate interactions among tested variables (i.e. recombination, gene density). These revealed a significant effect of both recombination ( $t = -41.17$ ,  $p < 0.0001$ , S8B Table) and gene density ( $t = 5.72$ ,  $p < 0.0001$ ) on PC1- $D_{XY}$  ( $R^2 = 0.348$ ). The same was true when considering the correlation of PC1- $\pi$  with recombination ( $R^2 = 0.55$ , Fig 6A-D,  $t = -40$ ,  $p < 0.0001$ ) and gene density ( $t = 5.22$ ,  $p < 0.0001$ ). Significant effects were also found between PC1- $F_{ST}$  and these variables (recombination:  $t = 6.91$ ,  $p < 0.0001$ , gene density  $t = 6.27$ ,  $p < 0.0001$ , interaction:  $t = -4.073$ ,  $p < 0.0001$ ,  $R^2 = 0.025$ , S8B Table). Similar correlations were found when analysing

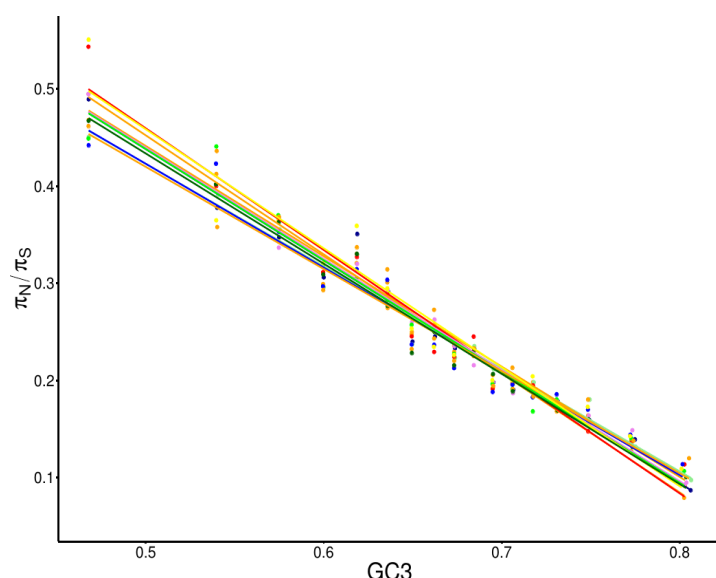
populations separately (S9 Table). Therefore, our results suggest a role of linked selection in shaping the patterns of diversity along the Coho salmon genome.

**Figure 7) Correlation**

**between  $\pi_N/\pi_S$  and GC3 content for each population separately**

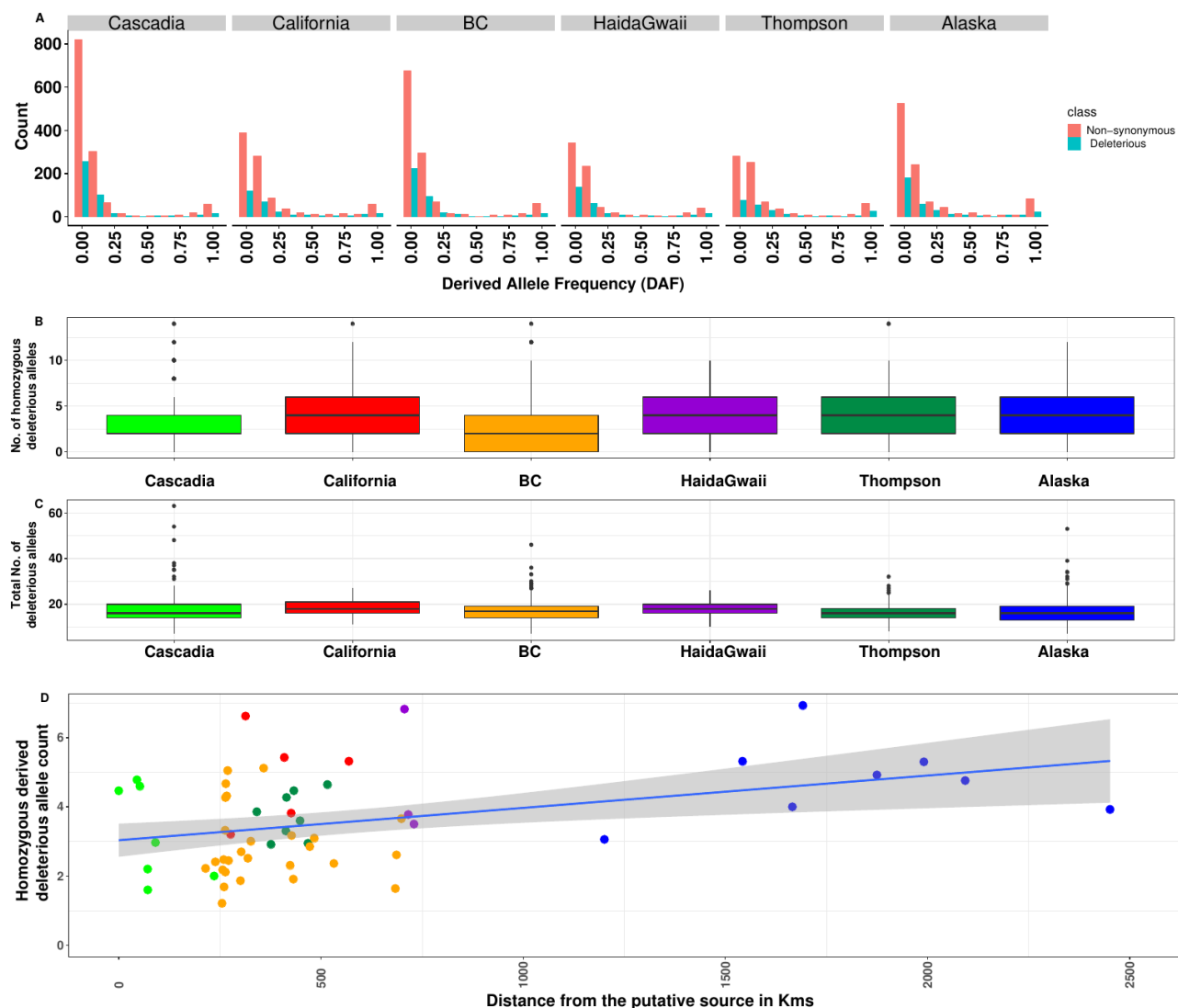
$\pi_N/\pi_S$  is used here as a proxy for the efficacy of purifying selection with smaller values reflecting more efficient purifying selection.

All correlations are significant ( $p < 0.001$ ).



We then tested whether the efficacy of selection also played a role in shaping patterns of diversity along the Coho salmon genome using the whole genome sequences. We measured ratios of non-synonymous to synonymous polymorphisms ( $\pi_N/\pi_S$ ) as a measure of the efficacy of natural selection and GC content at the third codon position (GC3) as a proxy of local recombination rates. This metric was preferred as an indicator of localized recombination rate variation at the gene scale, as opposed to the large population scale estimates of recombination ( $\rho$ ) over 250kb windows used above. Indeed, biased gene conversion (gBGC) occurs because base mismatch during homologous recombination is known to be biased toward G and C (conversion preferentially favours G+C over A+T bases) [56,57]. Here, we observed that the correlation between GC3 and  $\rho$  ranged from  $r = 0.3$  to  $0.72$ , depending on population. These modest correlations likely reflect the loss of information about fine scale recombination rate variation when using 250kb sliding windows, which also provides added justification to use the GC3 as a proxy. We found a strong correlation between  $\pi_N/\pi_S$  and GC3 for all populations (Fig 7, linear models all  $R^2 > 0.9$ ,  $p < 0.0001$ , S10 Table). Next, we observed a negative correlation (linear model,  $R^2 = 0.28$ ,  $p = 0.051$ ) between historical  $N_e$  measured before the onset of population expansion  $\sim 13,000$  years ago and  $\pi_N/\pi_S$  ratio (S16 Fig), in agreement with the nearly neutral theory of molecular evolution and indicating that our  $N_e$  estimates were good proxies of genetic diversity. Second, among-population differences in  $\pi_N/\pi_S$  across all genes were modest (mean =  $0.252$ ), with a minimum value of  $0.248$  observed in Tsoo-Yess R. (Cascadia) and Robertson Creek (BC) and maximum values of  $0.257$  and  $0.256$  observed in the Thompson R. and Inch Creek (BC) respectively. The  $\pi_N/\pi_S$  differences among populations were more contrasted in areas of low recombination of the genome (Fig 7) with lowest  $\pi_N/\pi_S$  values observed in the Kwethluck (Alaska,  $\pi_N/\pi_S = 0.442$ ), Quilcene (BC,  $\pi_N/\pi_S = 0.449$ ), and Deschutes (Cascadia,  $\pi_N/\pi_S = 0.451$ ) river populations. The highest values were observed in the Thompson ( $\pi_N/\pi_S = 0.551$ ), Klamath (California;  $\pi_N/\pi_S = 0.543$ ) and Capilano (BC;  $\pi_N/\pi_S = 0.495$ ) river populations.

# 380 *Spatial patterns of variation in deleterious mutations* 381



## 383 **Figure 8) Analysis of deleterious load in the Coho salmon genome**

384 A) Derived allele frequency spectra across major groups. The non-synonymous variants and putatively deleterious  
385 variants frequency are shown. Data are normalized for a sample of size  $n = 102$  corresponding to the smallest size for  
386 the combined samples in Haida Gwaii. B) Distribution of the count of homozygous derived deleterious alleles in each  
387 major group. C) Distribution of the count of total derived deleterious alleles in each major group. D) Correlation ( $p =$   
388  $0.0034$ ,  $R^2=0.13$ ) between geographic distance to the putative ancestral source and the distribution of homozygous  
389 derived putatively deleterious mutations.

390

391 Given the complex demographic history of population size changes and spatial expansions leading to  
392 secondary contacts in Coho salmon and the inference of linked selection, we predicted that deleterious  
393 mutations should segregate at higher frequencies at the edge of the norther expansion front in Alaska, but  
394 also in California likely representing a southern expansion from the Cascadian refugium, as well as in the  
395 bottlenecked Thompson R. populations.

To do so, we tested for an increase in the derived allele frequency (DAF), and homozygosity of predicted deleterious mutations at non-synonymous sites. We estimated derived alleles using whole genome information from three outgroups: 1) Chinook salmon (*Oncorhynchus tshawytscha*) [58], 2) Rainbow trout (*Oncorhynchus mykiss*) [59] and 3) Atlantic salmon (*Salmo salar*) [60]. Out of 4,427 non-synonymous mutations identified with the GBS dataset from all 58 populations, PROVEAN (63) predicted a total of 1,297 deleterious mutations in at least one of these populations and for which we were able to identify the derived allele. Deleterious mutations were maintained at lower DAF than synonymous variants (Wilcoxon-tests, all  $p < 0.001$ ). The DAF spectrum (Fig 8A) were computed for each major region (using a sample of size  $n = 100$ ) and for each population separately (S17 Fig). DAF were significantly different, among region (Kruskal-Wallis chi-squared = 100.57,  $df = 5$ ,  $p\text{-value} < 2.2e-16$ ) as well as among population (Kruskal-Wallis chi-squared = 638.59,  $df = 57$ ,  $p\text{-value} < 2.2e-16$ , S11 Table). They showed that Alaska, Thompson R. and California populations displayed more fixed deleterious mutations ( $n = 18$ , 12 and 11 respectively) than those from the Cascadia region ( $n = 3$ ). Haida Gwaii and BC populations were intermediate with 7 and 6 fixed deleterious mutations respectively (Fig 8A). The three former regions also displayed higher frequencies of polymorphic derived deleterious mutation (S18 Fig) compared to British Columbia or Cascadia (Wilcoxon test,  $p < 0.001$ ).

Next, we examined the count of homozygous (proportional to the load under a recessive model [33,34]), heterozygous and of total derived deleterious mutations (proportional to the load under an additive model [33,34]). In particular, we expected that mutations in a heterozygous state should be more frequent than in a homozygous state, especially in populations with higher effective sizes, where selection should be more effective at purging these mutations [23]. We found that 77% of deleterious mutations were maintained in heterozygous states across all samples. Also, fish from Alaska, Haida Gwaii and California populations harbored a significantly higher number of deleterious mutations in a homozygous state when compared to Cascadia or British-Columbia (Wilcoxon-test,  $p < 0.01$ ) (Fig 8B, S12 Table). When considering the total load of derived deleterious mutations, we found that, on average, there were significantly more putatively deleterious variants per individual in California, Cascadia, and Haida Gwaii populations than in fish from Alaska, British-Columbia or the Thompson R. watershed (S12 and S13 Table, Fig 8C, Wilcoxon-test,  $p < 0.01$ ), although these differences were modest. Finally, we tested the prediction that distance from the likely ancestral source predicts the deleterious load, as observed in human populations [33]. We found a nearly linear relationship between the distance from the putative origin of ancestral populations (the site with the lowest  $\beta_{ST}$  in Cascadia) and the number of derived, homozygous, putatively deleterious mutations (HDD; linear models,  $p < 0.0001$ ;  $R^2 = 0.13$ , Fig 8D). Under the hypothesis that higher recombination rate leads to more efficient purging of deleterious mutations we also expected deleterious mutations to be preferentially located in areas of low recombination (S19 Fig). As expected, the GLM revealed that the occurrence of deleterious mutations decreased as recombination rate increased ( $\chi^2 = 4.90$ ,  $DF = 1$ ,  $p = 0.027$ ). This effect was stronger when considering non-synonymous variants instead of putatively deleterious variants ( $\chi^2 = 10.07$ ,  $DF = 1$ ,  $p = 0.0015$ ). Given the negative correlation of recombination with chromosome length, we also found a positive correlation ( $r = 0.76$ ,  $p < 0.0001$ ) between the chromosome length and the number of

deleterious variants by chromosomes, as expected when purging is more efficient in areas of higher recombination (S20 Fig). To further examine this, we classified regions by recombination rate into high ( $\rho > 4.516$ , see methods), low ( $\rho < 1.411$ ) and intermediate classes and found that there was a significant excess of candidate deleterious mutations in areas of low recombination and depletion of candidate deleterious mutations in areas of high recombination ( $\chi^2 = 13.33$ , DF= 2,  $p = 0.0012$ ).

## Discussion

Coho salmon is an emblematic fish species that has undergone population declines in recent decades throughout its North American range. We generated one of the largest collections of GBS and WGS data to date for a non-model vertebrate species, which revealed: *i*) a complex demographic history involving population splits, gene flow, and secondary contacts; *ii*) linked selection and Hill-Robertson interference shaped genetic diversity along the genome, and *iii*) this demographic history has resulted in modest yet detectable differences in the frequency of deleterious variants across regions and populations, which also varied as a function of recombination rate along the genome. Together, these observations help illuminating the drivers of variation in genetic diversity throughout the genome of a broadly distributed species of economic and cultural importance.

## Expansion from a major southern refugium and secondary contact with micro-refugia

Our results revealed the existence of a major ancestral refugium located south of the ice sheets in Cascadia (i.e., Washington and Oregon) where contemporary populations contain most of the genomic diversity present over the entire North American range of the species. This conclusion was supported by observations of: *i*) a pronounced south-north gradual decrease of genetic diversity, singleton density, and private polymorphism, *ii*) a pattern of IBD, and *iii*) the occurrence of ancestral populations south of formerly glaciated areas. Therefore, although the so-called Beringian refugium that persisted north of the ice sheet (mainly comprising Alaska and the Yukon Territory) was important for many temperate species (6), this was not the case for Coho salmon. The gradual northward decrease in diversity also suggests that populations subsequently expanded from Cascadia to British-Columbia and Alaska [4,5] and is indicative of a serial founder effects due to small proportions of individuals issued from ancestral populations colonising new locations, a process amply documented in humans (e.g., [61]). This expansion was likely postglacial, although GBS-based parameter estimates of divergence time under AM, IM, and SC (~ 30 to ~ 45 KyA) were not in full agreement with SMC++ estimates using WGS (~18 KyA). This, however, is expected given the different assumptions made by the various models compared and that gene flow is expected to delay divergence [62]. The last glacial period (Wisconsin Glaciation) which lasted from ~ 120 Ky to ~11 KyA was interrupted by an interglacial period from 55 to 25 KyA ([40]). Therefore, our results indicate that neither Beringia nor Haida Gwaii were refugia before the onset of the Wisconsin Glaciation. However, it is possible that individuals from Cascadia already colonized and subsequently diverged in Alaska (Beringia) 55 to 25 KyA. If this was the case, however, the contribution of Beringia is likely minor as no strong footprint of ancestral diversity and no signal of secondary contact was inferred in this area.

While a single major southern refugium hypothesis is well supported by our data, our analyses also revealed patterns consistent with minor contributions of micro-refugia on extent genomic diversity. Indeed, our demographic modelling supported a model of postglacial (10-20 KyA, S6 Table) secondary contact (SC) detected in several pairwise comparisons among populations between California and Haida Gwaii ( $n = 10$  pairwise comparisons), Cascadia and Haida Gwaii ( $n = 2$ ), California and Thompson R. ( $n = 11$ ), and between Cascadia and Thompson R. ( $n = 1$ ). These results are counter-intuitive, because contacts between northern sites and California, but bypassing Cascadia, seem unlikely. A possible explanation is that our statistical power to detect SC involving putative micro-refugia was reduced. Here, indeed, the SC period represented a large proportion (22% on average) of the total divergence time across all models. However, models of SC can easily be confounded with models of isolation with migration when SC represents a large period of time ( $> 10\%$ ) relative to the total period of divergence, as observed in our investigations [19,63]. Still, our inference of SC supports the hypothesis that smaller micro-refugia have persisted along the Pacific Coast [43,64]. In particular, the Haida Gwaii archipelago is known for high endemism, and its role as a refugium for mammals, invertebrates, and angiosperms is well established [64–66]. The divergence time inferred between populations from California and Haida Gwaii was relatively recent (39 KyA), in the range of inferred divergence times for bird species from the region [67] and in the time frame of the last interglacial period 55 to 25 KyA. Our results also indicated a pronounced divergence between California and Cascadia, suggesting that these were likely two separated refugia, thus supporting the hypothesis of “refugia within refugia” [43,67]. However, California populations have suffered from strong census size reductions in recent decades [68], possibly increasing genetic structure and lowering effective population size. Given our estimates of split times and population sizes, the possible refugium in Haida Gwaii, and California were demographically small and relatively recent, such that populations in these areas were unlikely to have accumulated significant endemic genetic diversity. The case of the Thompson River, on the other hand, is rather intriguing as this region was entirely covered by ice during the Pleistocene. A possible explanation is that a strong and continuous bottleneck has led to a false signal of SC. The SC was the most parameterized model and was therefore more likely to be falsely supported. A model including more drastic and continuous post-divergence changes in population size may fit the data well but would likely be difficult to statistically separate from a model of secondary contact. British-Columbia populations, located at intermediate latitudes, also displayed intermediate levels of genetic diversity, with some populations displaying similar or higher diversity than those located in Cascadia, further supporting the hypothesis of post-colonization admixture between different refugia [69]. Complex history of divergence in different refugia and possible mixing was also inferred in the Chinook salmon along the Pacific Coast and was associated to different migratory ecotypes [70]. A potential limitations of our current approach is that we did not incorporate samples from Asia, where the Coho salmon also occurs [41]. Whether this would change our inferences remains an open question. In summary, our modelling approach revealed the presence of multiple cryptic refugia that would otherwise have been missed because their endemic genetic diversity appears to have been largely wiped out by a major demographic expansion out-of-Cascadia, which has clearly been the primary contributor of extent

509 genomic diversity in North American Coho salmon. Knowledge of this non-equilibrium demography will  
510 important to help interpret adaptive and deleterious genomics landscape in future studies.

511

## 512 **Selection at linked sites shape genome wide variation**

513 We also explored the role of gene flow and linked selection in generating a heterogeneous differentiation  
514 landscape along the genome, a hotly debated topic in evolutionary genomics [8,9,71,72]. Disentangling the  
515 two processes is of fundamental importance for correctly interpreting the origin of genomic islands of  
516 divergence and genome scan results [73,74]. Although a positive relationship between absolute and relative  
517 divergence may reveal the presence of barriers to gene flow [9], this is not relevant for early stages of  
518 divergence [75], as in the present study. Instead, our modelling approach best supported a role of both gene  
519 flow and linked selection. In particular, the role of linked selection is supported by the positive correlations  
520 between  $\pi$  or  $D_{XY}$  and recombination rate, while  $F_{ST}$  was negatively correlated with recombination rate  
521 [73,75]. The negative correlations between  $\pi$  or  $D_{XY}$  and gene density provided further evidence for the role  
522 of linked selection [17]. Here, heterogeneous genomic divergence did not arise after speciation, as suggested  
523 by Cruickshank & Hahn [9]. Instead, our within-lineage study indicated that linked selection arose within  
524 structured populations, in line with recent findings in birds [73]. Similar effects of linked selection were  
525 suggested between the more diverged ancestral lineages within Atlantic salmon (>1 MyA) [76]. Given that  
526 correlated genomic landscapes have been reported among species diverged for over tens of millions of years  
527 [77,78], it would be interesting to investigate how linked selection has also shaped genomic landscapes  
528 within the entire radiation of salmonid fishes. Indeed, salmonid species have undergone a whole genome  
529 duplication approximately 90 MyA [79] affecting recombination throughout their genome [57,80,81]. Precise  
530 mapping of the genomic locations of duplicated regions will allow an improved understanding of how  
531 recombination rate varies in these regions although we do not expect that will affect our main conclusion  
532 related to linked selection. Salmonid genomes are also characterised by pronounced male heterochiasmy  
533 [82,83] and here, we observed a significant correlation between recombination and chromosome length,  
534 indicative of crossover interference [84]. Both heterochiasmy and crossover interference contribute to  
535 heterogeneity in recombination rates and likely favour the effect of linked selection. Given that most non-  
536 neutral mutations are deleterious [85], and given the impact of recombination across the genomes of fishes,  
537 models of background selection could be the null model against which to test adaptive hypotheses [21]. Our  
538 data revealed that the GC3 was a good proxy of recombination and that it was strikingly well correlated with  
539 our proxy for the efficacy of natural selection ( $\pi_N/\pi_S$ ). As expected, this metric revealed differences in the  
540 most bottlenecked populations from the Thompson River drainage, where the severity of the bottleneck  
541 mimics the effect of domestication bottleneck (e.g. [86]). Indeed, the Thompson R. populations displayed a  
542 higher burden of non-synonymous mutations, and we suggest that the strong bottlenecks in these populations  
543 had a dramatic impact on the efficacy of natural selection. When considering only genomic regions of low  
544 recombination, however, we also found that some California populations also displayed an increased  $\pi_N/\pi_S$   
545 ratio. Given that these populations have recently undergone large declines in abundance [68], we expected a  
546 lower efficacy of purifying selection in regions of low recombination where Hill-Robertson interference

should be high [82]. Furthermore, we suggest that the severe and recent demographic declines documented in California populations compared to more long term and putatively less severe declines in the Thompson R. may explain why differences were detected only in genomic regions of low recombination in the former but genome-wide in the latter. Regardless, the non-equilibrium demography, and role of linked selection, in particular background selection, are predicted to favour the accumulation of deleterious mutations, an important finding to manage declining species.

### **Accumulation of deleterious mutations under complex demography**

In the literature, contradictory results have been reported regarding the role of demographic history on the load of deleterious mutations [10,33]. On the one hand, processes such as strong and repeated bottlenecks [24,32], including domestication [86,88], large expansions [89,90] or postglacial colonization [91] can all increase the deleterious load. On the other hand, empirical studies in human indicate that recent demography should not have a strong impact on the load of deleterious variants over the long term [10]. However, other studies indicate that strongly deleterious variants affecting fitness can still display increased frequency due to demographic growth [10]. Here, the inferred demographic history, together with support for linked selection and gBGC, provides ideal conditions to favour the accumulation of deleterious mutations. In particular, we found that populations displaying smaller effective population size (e.g. California), and those at the extreme of the expansion front (Alaska), displayed deleterious mutations at higher derived allele frequencies, and more frequently in homozygous state, in line with the nearly neutral theory. Similar findings have been observed in domesticated species [86] and recently in Isle Royale wolves [26] where decreased population sizes and inbreeding increased the frequency of deleterious recessive mutations. Overall, these results are consistent with recent empirical findings in which small populations or populations at the expansion front carry more homozygous derived deleterious mutations [26,27,86]. These deleterious mutations in homozygous state are expected to be purged by purifying selection [23,92]. Populations from Cascadia, with higher effective population sizes, contain a higher number of putatively deleterious variants in heterozygous state, as expected from population genetics theory. We also found a nearly linear relationship between the number of derived deleterious mutations in homozygous states and geographic distance from the putative refugial source in Cascadia. This relationship mirrors the findings pertaining to the ‘out-of-Africa’ expansion of human populations [33], although the maximum geographic distance in our study is an order of magnitude smaller than in the human studies, and that the presence of multiple refugia (as opposed to the sole African ancestral origin for humans) may have contributed to obscure somewhat the relationship. Moreover, pronounced genetic drift or bottlenecks observed in some populations (e.g. Thompson R. and California) may have contributed to reduce the efficacy of selection [24]. Finally, the lower prevalence of deleterious mutation in fish from British-Columbia might also be explained by post-glacial admixture. In some small populations it is possible that these mutation initially reached fixation but subsequently became masked in heterozygous state due to gene-flow and introgression [93]. Similar effect could be due to artificial supplementation programs, but the origin of source populations and intensity of stocking in our study populations are poorly documented. Finally, we observed that recombination was significantly correlated with the proportion of non-synonymous mutations and with the load of deleterious mutations, with

585 more deleterious mutations being found in genomic regions of low recombination. This is in line with  
586 theoretical predictions (e.g. Hill-Robertson Interference discussed above), and also empirical studies in other  
587 species [94,95]. Overall, our results suggest that Coho salmon populations with highly reduced population  
588 sizes are exposed to higher inbreeding depression – a prediction with major conservation implications and  
589 which should be investigated further in future studies.

## 590 **Conclusion**

591 In this study, we presented a rare combined assessment of the relative role of complex demographic history  
 592 (investigated both by empirical genomic data and modelling) and intrinsic genomic factors (recombination,  
 593 linked selection) in shaping drivers of the genomic landscape of a broadly distributed species. Complex  
 594 demographic processes including population expansion, isolation, and secondary contact were revealed  
 595 through the use of an extensive modelling framework. Moreover, our results highlighted the necessity of  
 596 accounting for local variation in recombination rate, a key driver of linked selection. Altogether, these  
 597 processes influence the efficacy of selection and can favour the accumulation of mutations affecting fitness.  
 598 Our findings suggest that such approaches offer enormous potential in the field of conservation genomics to  
 599 disentangle the impacts of historical vs. recent drivers of demographic declines and for assessing the  
 600 distribution of not only putatively beneficial, but also deleterious variants. We propose that future studies  
 601 should also integrate in-depth analysis of selective sweeps which will be necessary to investigate into more  
 602 details how linked selection, through background selection and hitchhiking, acts to maintain deleterious  
 603 mutations in the genome [96]. Finally, while it has become routine with new genome-wide datasets to focus  
 604 on the effect of positive selection and documenting patterns of local adaptation [7], this focus ignores the  
 605 fundamental prediction that most new mutations are likely to be deleterious [85]. An increased focus on  
 606 deleterious mutations would provide a more nuanced view of genomic evolution in wild species which  
 607 would benefit both the fields of evolutionary and conservation genetics

608

## 609 **Methods**

### 610 ***Genotyping By Sequencing***

611 A total of 2,088 individuals was collected from 58 sample sites located along the Pacific coast from  
 612 California to Alaska (S1 Table and Figure 1). DNA was extracted from all individuals and sequenced using a  
 613 GBS method (protocol detailed in [97]). Reads were aligned to the Coho salmon reference genome v1  
 614 (GCF\_002021745.1) using bwa-mem 0.7.13 [102]. Samtools v1.7 was used to keep reads with a  
 615 mapping quality above 20, remove supplementary alignment and unmapped read. Variants were then  
 616 called with Stacks v1.46 [98]. To do so, the module “pstacks” was used with a minimum depth of 5, and up  
 617 to three mismatches were allowed in catalog assembly. The module “populations” was run to produce a vcf  
 618 file that was filtered with a custom python script. We performed stringent filtering to remove SNPs that were  
 619 1) genotyped in less than 60% of the individuals; 2) at a mean depth of sequencing below 7, and 3) with  
 620 observed heterozygosity above 0.60, thus resulting in 93,000 SNPs. The pipeline for SNP calling is available  
 621 on github at [https://github.com/enormandeau/stacks\\_workflow/releases/tag/coho\\_demography\\_paper](https://github.com/enormandeau/stacks_workflow/releases/tag/coho_demography_paper). Next,  
 622 we removed any individuals with more than 5% missing data and finally only kept SNPs present in at least  
 623 95% of the individuals yielding a total of 82,772 filtered SNPs for 1,957 individuals. Remaining filtration  
 624 was done according to the requirement of each analysis performed below.

## 625 ***Genetic diversity and ancestral populations***

626 For each sampling location we estimated the observed heterozygosity and  $\pi$  using vcftools 0.1.16 [99] and  
 627 hierfstat [100]. The most likely ancestral populations were identified using  $\beta_{ST}$  [46]. A total of 1,000  
 628 bootstraps was performed to obtain the 95% confidence intervals around the  $\beta_{ST}$ . Weir and Cockerham's  $F_{ST}$   
 629 estimator  $\theta$  [47] was computed in vcftools. We measured the relationship between observed heterozygosity,  
 630  $\beta_{ST}$ ,  $F_{ST}$  and the distance to the southernmost site using linear models. We also verified the relationship  
 631 between  $F_{ST}$  and the distance to the southernmost site using Mantel tests with 10,000 permutations. Vcftools  
 632 was also used to identify singletons (i.e. variants present in one single individual across the whole dataset).  
 633 Their distributions were then summed in each locality. We then computed the averaged (min, max and  
 634 median) number of singletons at the regional level. The scripts are available on github at [https://github.com/](https://github.com/QuentinRougemont/utility_scripts)  
 635 [QuentinRougemont/utility\\_scripts](https://github.com/QuentinRougemont/utility_scripts)

## 637 ***Population structure, admixture and gene flow***

638 Levels of ancestry and admixture proportions were inferred with the snmf function implemented in the R  
 639 package LEA [101]. We allowed less than 5% of missing data. We then kept a set of SNPs in approximate  
 640 linkage equilibrium by removing pairs of variants with  $r^2$  greater than 0.2 (option --indep-pairwise 50 10 0.2)  
 641 resulting in 40,632 SNPs. K-values ranging between 1 and 60 were tested and cross-validations were  
 642 obtained using the cross-entropy criterion with 5% of masked genotypes. The default value for the  
 643 regularization parameter was used to avoid forcing individuals into groups and hence underestimating  
 644 admixture. Similar results were obtained from Admixture [102] and are not presented here. Genetic  
 645 relationship among all salmon was assessed using a PCA with the R package ade4 [103] based on the LD-  
 646 pruned dataset (40,632 SNPs). We used a 1% minor allele frequency (MAF) threshold and allowed less than  
 647 4% missing data. Formal tests of admixture were performed using Treemix [49] using the LD-pruned dataset  
 648 of 40,632 SNPs and without any MAF threshold. A MDS was also constructed using plink and plotted with  
 649 the ggplot2 [105] R package. We ran Treemix allowing up to 20 migration events and performed 500  
 650 bootstraps replicates of the best model to obtain robustness of the nodes. The “best” model was inferred after  
 651 inspecting the relevant migration edges by measuring the percentage of variance explained as migration edge  
 652 were added to the tree as well as by assessing the p-value associated to each migration edge. A total 500  
 653 bootstraps replicate was performed under the “best” model and under a model without migration to infer the  
 654 robustness of the nodes. The scripts are available on github at  
 655 [https://github.com/QuentinRougemont/treemix\\_workflow](https://github.com/QuentinRougemont/treemix_workflow)

## 657 ***Explicit demographic inferences.***

658 We tested alternative hypotheses of secondary contact between major regional groups (Haida Gwaii,  
 659 California, Cascadia (Washington-Oregon), Alaska, British Columbia, Thompson R.) of populations by  
 660 comparing alternative divergence scenarios represent in Fig S12 and initially described in [19,76].  
 661 Alternative hypotheses of secondary contacts were tested between major groups by testing the significance  
 662 of alternative divergence scenarios. The four major models tested included a model of Secondary Contact

(SC), a model of Strict Isolation (SI), a model of Ancient Migration (AM) and a model of Isolation with Migration (IM).

The models shared the following parameters: the ancestral populations of size  $N_{anc}$ , splits at time  $T_{split}$  into two daughter populations of size  $N_1$  and  $N_2$ . Under the SI model, no gene flow occurs between the two populations. Under AM, gene flow occurred between  $T_{split}$  and  $T_{am}$  and is followed by a period of strict isolation. Under IM, gene flow occurs at a constant rate at each generation between the two populations. Gene flow can be asymmetric, so that two independent migration rates  $m_{12}$  (from population 2 to 1) and  $m_{21}$  (from population 1 to 2) were modeled. Under the SC model, the population evolved in strict isolation between  $T_{split}$  and until  $T_{sc}$  where a secondary contact occurs continuously up to present time. Gene flow is modeled as  $M = 2N_{ref}m$ . In  $\partial a \partial i$ , heterogeneous introgression was modeled using two categories of loci occurring in proportions  $P$  (i.e., loci with a migration rates  $M_{12}$  and  $M_{21}$ ) and  $1-P$  (i.e., loci with a reduced effective migration rates  $Me_{12}$  and  $Me_{21}$ ) across the genome. The same procedure was used to account for linked selection by defining two categories of loci with varying effective population sizes (proportion  $Q$  of loci with a “neutral  $N_e$ ” and a proportion  $1-Q$  of loci with a reduced effective population size due to either linked or background selection). To quantify how linked selection affects reduced  $N_e$ , we used a Hill-Robertson scaling factor ( $Hrf$ ) to relate the effective population size of loci influenced by selection ( $N_r = Hrf * N_e$ ) to that of neutral loci ( $N_e$ ).

Models were fitted using the diffusion theory implemented in  $\partial a \partial i$  [50] and also includes the effect of linked selection and barriers to gene flow as detailed in [19,104].  $\partial a \partial i$  uses the SFS as a summary of the data. For a given demographic model, the SFS is computed using diffusion approximation and compared to the empirical SFS using AIC. Here, we started from the whole file containing 200,000 SNPs and used one single SNP per GBS locus, filtered the data to minimize missing data. No MAF was used and singletons were kept to obtain ascertainment-free estimates of demographic parameters. Ideally, no Hardy-Weinberg Equilibrium (HWE) filter should be used for demographic inferences, as this also biases the distribution of allele frequencies. However, to remove paralogs present in the Coho salmon genome, a permissive HWE filter based on a p-value of 0.0001 was used. Here, a total of 69 pairwise comparisons between populations from the major regional groups was performed in order to test for a prevailing pattern. Here, the regional groups considered were all previously unglaciated areas during the LGM, namely California, Cascadia, Alaska (although some samples were likely under ice at different time periods) and Haida Gwaii. For each model, 20 independent replicate runs were performed and only models with the lowest AIC and  $\Delta AIC$  were kept. A model was classified as “ambiguous” and not used for parameter estimation if  $\Delta AIC$  between the best model and second-best model was below 10. The whole pipeline is available at <https://github.com/QuentinRougemont/DemographicInference>.

### ***Analyses based on whole genome resequencing data.***

We used 55 individuals representing 11 populations from California to Alaska (S2 Table). Each individual was sequenced on an Illumina platform using paired-end 150 bp reads. Reads were processed using fastp for trimming [105], bwa mem v0.7.13 [106] for mapping, samtools v1.7 requiring a minimum quality of 10, and

picard to remove duplicates (<http://broadinstitute.github.io/picard/>). Then SNP calling was performed using GATK [107]. Genotypes were filtered for depth between 10 and 100 reads to remove low confidence genotypes including potential paralogs, displaying high coverage. Then following GATK Best Practices we excluded all sites that did not match the following criterion:  $MQ < 30$ ,  $QD < 2$ ,  $FS > 60$ ,  $MQRankSum < -20$ ,  $ReadPosRankSum < 10$ ,  $ReadPosRankSum > 10$ . We also generated a vcf file using the samtools mpileup pipeline, merging individuals with bcftools and performing the same stringent filtering as with the vcf constructed with GATK. Finally, we also generated a separate vcf file using the emit-all-site option to call variable and invariable sites across the whole genome. This vcf file was used in the sliding windows analysis below to test the effect of linked selection. The whole pipeline is available on github ([https://github.com/QuentinRougemont/gatk\\_haplotype](https://github.com/QuentinRougemont/gatk_haplotype)).

SMC++ [53] was used to infer changes in population size through time. SMC++ works similarly to the PSMC model but takes advantage of information contained in the Site Frequency Spectrum (SFS) in addition to linkage disequilibrium, and is particularly well suited to analyse large sample sizes. Estimates of population size changes were performed for all 11 populations. Splitting time was estimated between all pairs of samples based on the joint SFS. A generation time of 3.5 years and a mutation rate of  $8e^{-9}$  mutation/bp/generation were applied. The pipeline to reproduce the analysis is available on github ([https://github.com/QuentinRougemont/smcpp\\_input](https://github.com/QuentinRougemont/smcpp_input)).

We computed pairwise linkage disequilibrium using vcftools with the  $r^2$  statistics calculated in all 11 populations separately. A window of 1,000,000 base pairs was used and all SNPs were included. To reduce the number of SNPs, we allowed no missing data and used a MAF of 10% and a p-value of Hardy-Weinberg disequilibrium of 0.05 in each population, keeping between 2 and 4 million SNPs depending on the population. We then estimated LD decay by plotting LD against physical distance measured in base pairs and using smoothing functions implemented in ggplot2 [108] package in R. vcftools was also used to identify singletons, for which the distribution were counted by localities, as for the GBS dataset.

We used LDHAT software [109] to estimate effective recombination rates ( $\rho=4.Ne.r$  where  $r$  represents the recombination rate per generation and  $Ne$  is the effective population size) along the genome. Unphased genotypes were converted into LDHAT format using vcftools with a minimum MAF of 10% since only common variants are useful for such inferences [109]. Following the authors' guidelines, the genome was split in chunks of 2,000 SNPs with overlapping windows of 500 SNPs to compute recombination rate and data were then merged together. We measured recombination rates for each river as well as globally including all populations except the population from the Thompson R. watershed that was too divergent from the remaining samples. The pipeline to reproduce the analysis is available on github ([https://github.com/QuentinRougemont/LDhat\\_workflow](https://github.com/QuentinRougemont/LDhat_workflow)).

### ***Clarifying the role of linked selection***

The genetic landscape of divergence was measured by estimating levels of nucleotide diversity ( $\pi$ ), gene density, levels of genetic differentiation ( $F_{ST}$ ), levels of divergence ( $D_{xy}$ ), and scaled recombination rate along the genome. Estimates of  $\pi$ ,  $F_{ST}$ , and  $D_{xy}$  were computed into 500kb windows using Python scripts

739 available from [110]. Gene density was computed directly from the gff file of the Coho salmon genome v1.0  
 740 (NCBI ftp file: [ftp://ftp-trace.ncbi.nih.gov/genomes/Oncorhynchus\\_kisutch/GFF/](ftp://ftp-trace.ncbi.nih.gov/genomes/Oncorhynchus_kisutch/GFF/); Rondeau et al. In prep)  
 741 and measured in 500 kb windows. Recombination rates were averaged into 250 and 500-kb windows using a  
 742 Python script after being estimated with LDhat as described above. We used a PCA to obtain a synthetic  
 743 view across all 11  $\pi$  estimates, and all pairwise 55 Dxy and 55  $F_{ST}$  separately. This allowed capturing the  
 744 common variation affecting these three estimates. First, simple Spearman correlations based on linear models  
 745 testing each (z-transformed) variable projected on the PC1 axis separately were carried out to produce fig 5.  
 746 Then Mixed Linear Models were used to test for correlations between either  $\pi$ , Dxy, or  $F_{ST}$  and either  
 747 recombination landscape ( $\rho$ ) or gene density windows as explanatory variables, with and without  
 748 interaction. Correlations were also calculated considering each  $\pi$ , Dxy, and  $F_{ST}$  by population separately  
 749 without PCA transformation (S9 Table) which returned patterns that were congruent with those observed  
 750 with PCA results.

751

## 752 **Measuring GC content, Non-Synonymous and Synonymous diversity**

753 The pipeline developed by [111] was used to compute Tajima's  $\pi$  estimator of nucleotide diversity and the  
 754 GC content over non-overlapping 10-kb windows. Then we concatenated the gene into different classes  
 755 according to their length and compute  $\pi_N$  and  $\pi_S$  over 4 mb windows, which allows circumventing the  
 756 problem associated to low  $\pi_S$  values. We then verified the correlation between populations scaled  
 757 recombination rate ( $\rho$ ), measured over 250-kb windows and GC at the third codon position (approximately  
 758 neutral). To do so, we assigned a  $\rho$  value for each gene according to its position into each 250-kb windows.  
 759 Then we averaged  $\rho$  values over the same length class as GC3 classes and compared the median population  
 760 recombination rate estimates and GC3s over 4-mb windows. Finally, we measured the correlation between  
 761 GC3s and  $\pi_N/\pi_S$  using linear models.

762

## 763 **Genetic load estimation from the GBS data set**

### 764 *Estimating ancestral and derived alleles*

765 We used the derived allele count as an estimator of deleterious allele count. We used the genomes of three  
 766 outgroup species, the chinook salmon, the rainbow trout, and the Atlantic salmon, to classify SNPs as  
 767 ancestral or derived. Whole genome data for the chinook salmon ( $n = 3$  individuals) were provided by one of  
 768 us (B. Koop, unpublished), whereas rainbow trout ( $n = 5$ ) and Atlantic salmon ( $n = 5$ ) data were downloaded  
 769 from NCBI Sequence Read Archive (rainbow trout, SRA, Bioproject : SRP117091; *Salmo salar* SRA  
 770 Bioproject: SRP059652). Each individual was aligned against the Coho salmon V1 reference genome  
 771 (GCF\_002021745.1) using GATK UnifiedGenotyper and calling every SNP using the EMIT\_ALL\_SITES  
 772 modes. We then constructed a Python script and determined the ancestral state of the GBS SNPs if 1) the  
 773 SNP was homozygous in at least two of the three outgroups, and 2) match one of the two alleles identified in  
 774 Coho salmon. Otherwise, the site was inferred as missing and not used in subsequent analyses.

## 775 *Measuring damaging impact of non-synonymous alleles*

776 We tested differences in mutation load among populations as follows. The software Provean [112] was used  
 777 to predict whether a given non-synonymous mutation was deleterious with a threshold score of -2.5 or less  
 778 using the pipeline available at [https://github.com/QuentinRougemont/gbs\\_synonymy\\_with\\_genome](https://github.com/QuentinRougemont/gbs_synonymy_with_genome). We  
 779 analysed data in two ways: first we counted the total number of putative homozygous deleterious alleles per  
 780 individual as well as the total number of deleterious alleles (both in homozygous and heterozygous states)  
 781 using:  $N_{total} = 2 \times N_{homo} + N_{hetero}$  [34]. These individual values were then averaged per population and major  
 782 regional groups (i.e., California, Cascadia, British Columbia, Haida Gwaii, Thompson, and Alaska). We then  
 783 computed derived allele frequencies (DAF) in all sampling locations and across mutation categories  
 784 (synonymous, non-synonymous, and non-synonymous deleterious) and tested for significant allele frequency  
 785 differences among populations in non-synonymous and non-synonymous deleterious mutations using  
 786 Wilcoxon rank sum tests. DAF spectra were constructed for all population separately. For the ease of  
 787 visualisation, we also constructed DAF spectra by region for a sample of size  $n = 100$  individuals (Fig 7A).  
 788 This size was chosen according to the smallest sample size of the three combined Haida Gwaii populations.  
 789 Finally, we tested for a preferential enrichment of “hotspots or coldspots” of recombination in deleterious  
 790 mutations. To define a coldspot, we first computed a lower bound that we defined as the  $mean_{RHO} -$   
 791  $5 \times \text{standard errors}$  for each chromosome separately. Similarly, “hotspots” of recombination were identified  
 792 using an upper bound defined as  $mean_{RHO} + 5 \text{ standard errors}$ . We then tested if the average recombination rate  
 793 of each of the 250 kb windows was falling below the threshold (for coldspot) or above it (for hotspots). We  
 794 then tested if recombination hotspots and coldspots contained more or less putatively deleterious mutations  
 795 than “normally” recombining regions using  $\chi^2$  tests. Linear mixed effects models were further used to test if  
 796 there was a relationship between recombination rate and the distribution of putatively deleterious mutations.  
 797 The response variable was the deleterious state considered binomial (0 = non deleterious, 1 = deleterious)  
 798 and the explanatory variable was the recombination rate. The chromosome identity was included as a random  
 799 effect. This model was compared against a null model excluding recombination rate. The analysis was  
 800 replicated but using the synonymous and non-synonymous state as the response variable instead of the  
 801 putatively deleterious state of the considered mutation. Models were carried out in R using the LME4  
 802 package [113]. Finally, we used SnpEff v3.4 [114] to obtain the functional annotations of the putatively  
 803 deleterious variants. The annotations were comprised of mis-sense variants, non-coding transcripts, 3' and 5'  
 804 untranslated regions, 5kb up- and down-stream variants, intergenic and intronic, splice acceptor and splice  
 805 region, stop gained and start loss (S14 Table). We found 35% of deleterious mutations to be missense  
 806 variants, 38% to be non-coding transcript and 22% to be either upstream or downstream gene variants, with  
 807 the remaining being spread over the categories.

## 808 **Ethic Statement**

809 A permit number SIRUL 111722 was obtained to work on DNA sequences.

810

811 Raw data will be deposit on NCBI together with Short Read Archive (SRA) accession number.

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## Supporting Information Legend

**S1 Fig. Linear decrease in genetic diversity when considering  $\pi$ SNP as a function of the distance to the southernmost sample site.** Each points represents a sample site and is coloured by region.

**S2 Fig. Network of shared and private polymorphisms.** The branch (grey) represent shared polymorphism between sample site and are proportional to levels of sharing. Each point represent the number of private polymorphisms and is coloured by region. Computation were based on a sample of size 100 in each region to enable comparison.

**S3 Fig. Measure of linkage disequilibrium decay using whole genome sequencing data.** LD between pairs of loci was measured as  $r^2$  in windows of 1 million base pairs. Means smoothed value are display for each whole genome samples.

**S4 Fig. Patterns of Isolation By Distance.** Increasing  $F_{ST}$  as a function of the distance to the southernmost site. Each points represents a sample site and is coloured by region.

**S5 Fig. Summaries of  $F_{ST}$  values.**

- A.  $F_{ST}$ -based Hierarchical tree depicting relationship among samples. Colors represent the major region.
- B. Heatmap of  $F_{ST}$  values among samples ordered from North to South on the X and Y-axis.

**S6 Fig. Structure and Admixture inferences.**

- A. Admixture Barplot obtained from LEA for various K-values.
- B. Progressive decrease of LEA cross-entropy criterion. Lower cross-entropy values indicates the number of cluster compatible with the data (here from 30 to 60).

**S7 Fig. Multidimensional Scaling (MDS) plot depicting relationship among individuals.** Each points represents an individual site and is coloured by region.

**S8 Fig. Principal Component Analysis recapitulating the relationship among individuals.** The Axis 3 and axis 4 are displayed.

**S9 Fig. Treemix results**

- A. Proportion of variance explained (y-axis) as a function of the number of migration edge (x-axis)
- B. Treemix tree inferred without gene flow and residuals
- C. Residuals for Treemix tree with four migration edges

**S10 Fig. Compared Demographic Models**

Strict Isolation (SI), Isolation with constant Migration (IM), Ancient Migration (AM) and Secondary Contact (SC). The models shared the following parameters:  $T_{split}$ : number of generation of divergence (backwards in time).  $N_{anc}$ ,  $N_1$ ,  $N_2$ : effective population size of the ancestral population, of the first and second daughter population.  $M_1$  and  $M_2$  represent the effective migration rates per generation with  $m$  the proportion of population made of migrants from the other population.  $T_{sc}$  is the number of generations since gene flow started (secondary contact) after a period of isolation.  $T_{am}$  is the number of generations since the two populations have diverged without gene flow. Each models are declined in alternative version allowing homogeneous or heterogeneous gene flow and homogeneous and heterogeneous effective size to account for the effect of linked selection (affecting  $N_e$ ) and barrier to gene flow (affecting  $m$ ).

**S11 Fig. Estimate of effective population size across models with linked selection (2N suffix) and barriers to gene flow (2m suffix).** AM2m = Ancient Migration with heterogeneous migration.

AM2N = Ancient Migration with heterogeneous effective population size. IM2m = Isolation with Migration with heterogeneous migration. IM2N2m = Isolation with Migration with heterogeneous migration and with heterogeneous effective population size, SC2m = Secondary Contact with heterogeneous migration.

875 SC2N2m = Secondary Contact with heterogeneous migration and with heterogeneous effective population  
876 size.  
877

878 **S12 Fig. SMC++ estimates of effective population sizes for the 11 whole genome samples included.**  
879

880 **S13 Fig. SMC++ estimates of divergence time between all possible pairs of joint Site Frequency**  
881 **Spectra based on WGS data.** Two estimates of mutation rate were used. One standard estimate of  $1.25 \times 10^{-8}$   
882  $\mu/\text{bp/generation}$  and a second based on unpublished data.  
883

884 **S14 Fig. Negative Relationship between effective recombination rate ( $\rho$ ) and length of chromosome**  
885 **in bp.**  
886

887 **S15 Fig. Genome-wide landscapes in coho salmon with (A) landscape of differentiation, B) landscape**  
888 **of genetic diversity C) divergence, D) recombination E) gene density, plots are averaged over 500kb**  
889 **windows.** The figure is the same as figure 5 but all the 30 chromosomes are displayed.  
890

891 **S16 Fig. Correlation between  $\pi_N/\pi_S$  and effective population size ( $N_e$ ) during postglacial time.** Each  
892 point represent a sample locality and is color coded by region.  
893

894 **S17 Fig. DAF spectrum of nonsynonymous and putatively deleterious mutation in each region for all**  
895 **samples.**  
896

897 **S18 Fig. Significant mean derived allele frequencies of deleterious mutation.** Displayed are the mean  
898 dervied allele frequencies of polymorphic deleterious sites in each regions  $\pm 2$  standard deviation.  
899 x-axis = chromosome position (in bp) and y-axis = effective recombination landscape. Each red dots  
900 represents a candidate deleterious mutation.  
901

902 **S19 Fig. Location of putatively deleterious mutation along the effective recombination landscape.**  
903

904 **S20 Fig : Positive correlation between chromosome length and occurence of deleterious variants along**  
905 **chromosome.**  
906

907 **Table Legend:**  
908

909 **S1 Table:** Abbreviation, Region and coordinates (Longitude and Latitude) of each river used in the GBS data  
910 with the number of individuals provided (nb. Inds).  
911

912 **S2 Table:** River Name, region and coordinates (Longitude and Latitude) of each river used in the whole  
913 genome resequencing data.  
914

915 **S3 Table:** BST values along with 95% confidence intervals for each river from the GBS data (82 K SNPs).  
916 95% confidence intervals obtained after 1000 bootstraps.  
917

918 **S4 Table:** Distribution of singleton in the WGS data. The Thompson sample displays less singelton than  
919 southern samples.  
920

921 **S5 Table:** model choice results for dadi. AIC and deltaAIC are provided for each pairwise comparison and  
922 model. AM = Ancient Migration, IM = Isolation with Migration, SI = Strict Isolation, SC = Secondary  
923 Contact, the simplest models assume homogeneous migration and homogeneous effective population size.  
924 2N suffix = heterogeneous effective population size, 2Msuffix = heterogeneous migration.  
925 Model with both suffix assumes that both effective population size and migration are heterogeneous.  
926 Modil with a single suffix assumes that either migration or effective population size are heterogeneous.  
927

928 **S6 Table :** Parameter estimates from GBS data obtained under the best demographic model with dadi.  
929  $N_{e1}$  and  $N_{e2}$ , effective population size of the compared pair.  $m1 \leftarrow 2$  and  $m2 \leftarrow 1$ , migration from  
930 population 2 to population 1 and migration from population 1 into population 2.  $me_{12}$  and  $me_{21}$ , effective  
931 migration rate estimated in the most differentiated regions of the genome Ts: Split Time of the ancestral

932 population in two population; Tsc: duration of the secondary contact P: proportion of the genome freely  
 933 exchanged (1-P provides the proportion of the genome non-neutrally exchanged); Q: proportion of the  
 934 genome with a reduced effective population size due to selection at linked sites; hrf = Hill-Robertson factor  
 935 representing the reduction of Ne in the region Q with reduced Ne.

936  
 937 **S7 Table:** PCA loadings for Dxy, Fst and Pi.

938  
 939 **S8 Table:** A) Spearman correlation association to the comparison in Figure 6.  
 940 B) Linear models testing the combined effect of recombination (Rho) and gene density (Gene count).  
 941 Interaction terms were not significant for Pi and Dxy and were removed.

942 **S9 Table:** spearman correlation obtained between recombination and Dxy, Fst and Pi when considering each  
 943 possible pairs of river separately (n = 55) or each river independently (for Pi only).

944  
 945 **S10 Table:** Results of linear models testing the correlation between piN/piS and GC3.

946  
 947 **S11 Table:** Summary of deleterious variation by region.  
 948 1) Derived Allele Frequency (DAF) of deleterious mutation, after averaging by rivers and then by major  
 949 regional group. 2) Count of deleterious mutations in each rivers and then averaged by major regional group.  
 950 3) Number of homozygous derived deleterious mutations by individual, after averaging by rivers and then  
 951 by major regional group. 4) Number of heterozygous mutations by individuals, after averaging by rivers and  
 952 then by major regional group 5) Total load of derived deleterious mutations by individuals, after averaging  
 953 by rivers and then by major regional group.

954  
 955 **S12 Table:** Results of Wilcoxon test (Man-Whitney tests) for differences in derived allele frequencies  
 956 among major groups for a sample of size 100.

957  
 958 **S13 Table:** Results of Wilcoxon test (Man-Whitney tests) for differences in count of derived homozygous  
 959 variants and total load among individuals in each region.

960  
 961 **S14 Table:** SNPeff results classified by categories for the deleterious mutations identified in the GBS data.