

Multiple resource use strategies confer resilience to the socio-ecosystem in a protected area in the Yucatan Peninsula, Mexico

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2 ABSTRACT

3 Natural Protected Areas (NPAs) are the main biodiversity conservation strategy in Mexico.
4 Generally, NPAs are established on the territories of indigenous and rural groups driving
5 important changes in their local resource management practices. In this paper we study the
6 case of *Otoch Ma'ax Yetel Kooh*, an NPA in the Yucatan Peninsula, Mexico, that has been
7 studied in a multidisciplinary way for more than twenty years. This reserve and its buffer zone
8 is homeland to Yucatec Mayan communities that until recently used to manage their resources
9 following a multiple use strategy (MUS), which involves local agricultural practices and has been
10 proposed as resilience-enhancing mechanism. However, due to the restrictions imposed by the
11 decree of the reserve and the growth of tourism in the region, some of these communities have
12 started to abandon the MUS and specialize on tourism-related activities. We build a dynamical
13 computational model to explore the effects of some of these changes on the capacity of this NPA
14 to conserve the biodiversity and on the resilience of households to some frequent disturbances in
15 the region. The model, through the incorporation of agent-based and boolean network modelling,
16 explores the interaction between the forest, the monkey population and some productive activities
17 done by the households (milpa agriculture, ecotourism, agriculture, charcoal production). We
18 calibrated the model, explored its sensibility, compared it with empirical data and simulated
19 different management scenarios. Our results suggest that those management strategies that

20 do not exclude traditional activities may be compatible with conservation objectives, supporting
21 previous studies. Also, our results support the hypothesis that the MUS, throughout a balanced
22 integration of traditional and alternative activities, is a mechanism to enhance household resilience
23 in terms of income and food availability, as it reduces variability and increases the resistance to
24 some disturbances. Our study, in addition to highlighting the importance of local management
25 practices for resilience, also illustrates how computational modeling and systems perspective are
26 effective means of integrating and synthesizing information from different sources.

27 **Keywords:** Yucatan Peninsula, Mexico, socio-ecological system, milpa agriculture, ecotourism,
28 agent-based model, boolean network, resilience

1 INTRODUCTION

29 Natural Protected Areas (NPAs) are one of the main conservation strategies in the world to face
30 biodiversity loss. In 2016, 14.7% of terrestrial and inland water ecosystems worldwide were under this
31 type of protection (UNEP-WCMC and IUCN, 2016). This strategy, however, has been highly controversial
32 for their impact on the rural communities that inhabit the protected territories, namely, on their forms of
33 resource management and appropriation (Durand and Jímenez, 2010; García-Frapolli, 2015). Nowadays,
34 conservation strategies in NPAs range from total exclusion of human activities, to the promotion of the
35 sustainable use of the territory (Brockington and Wilkie, 2015).

36 Mexico, as one of the biologically megadiverse territories in the world, has adopted NPA as its main
37 governmental conservation strategy. Nationwide, 10.87% of the terrestrial and inland water ecosystems
38 are under protection (CONANP, 2018). In addition, the country also harbours a great cultural diversity
39 and indigenous groups territories generally are on zones of high biodiversity (Toledo, 2013). Therefore,
40 many of NPAs in Mexico are inhabited by indigenous communities with long-lived productive and cultural
41 practices, and as such, they constitute a highly diverse and complex social-ecological systems (SESs).
42 SESs are indeed conceptualized as complex systems in which humans are considered as part of nature
43 (Berkes et al., 2003). In contrast to traditional ecological and social research, studies with a SES perspective
44 explicitly consider ecological and human components and their interactions (Liu et al., 2007). Additionally,
45 in SES research there is a particular interest in studying and understanding feedbacks, nonlinear dynamics,
46 thresholds, unintuitive behaviors, time lags, heterogeneity and resilience (Liu et al., 2007).

47 As the world undergoes unprecedented global changes, SES research has been particularly interested in
48 understanding resilience and characterizing resilience-enhancing mechanisms (Biggs et al., 2012). This
49 quest has lead to recognizing the value that some traditional or local practices, largely overlooked and
50 blamed as unproductive and environmentally damaging, may actually have for resilience (Berkes et al.,
51 2000). An example of these is the multiple-use strategy (MUS) on which indigenous communities often
52 base their resource management and family household (Toledo et al., 2003). MUS involves the development
53 of a set of different productive activities (e.g., agriculture, agroforestry, gathering, etc.) on a diversity of land
54 units (e.g., milpa, successional forest, mature forest, etc.). It has been argued that this strategy amplifies the
55 subsistence options available for the households to ensure a continuous flow of goods and services, thus
56 minimizing the vulnerability associated with different disturbances (Barrera-Bassols and Toledo, 2005).
57 However, integral assessments of coupled environmental and social outcomes of MUS are largely lacking.

58 In this paper we study the MUS practiced by a Mayan community inhabiting *Otoch Ma'ax Yetel Kooh*
59 (OMYK, “house of spider monkey and jaguar” in Yucatec Maya), an NPA located at the northeastern part
60 of the Yucatan Peninsula, Mexico. This NPA was declared in 2002 in response to a local initiative, after

61 years of community-based conservation. It consists of a body of lakes surrounded by forest in multiple
62 successional stages that harbours a large population of spider monkeys and other ecologically important
63 species. OMYK and its buffer zone are inhabited by Yucatec Maya communities, which, until recently, used
64 to manage their resources following a MUS based on traditional swidden milpa agriculture (García-Frapolli
65 et al., 2008). The milpa system consists of a polyculture which typically combines maize, beans and
66 squash, along with other domesticated, semi-domesticated and tolerated species (Benítez et al., 2014).
67 Milpa system is highly diversified and adapted to an ample variety of environments, playing a key role in
68 the maintenance of biological diversity and food sovereignty (Altieri et al., 2012). However, due to the
69 growing tourism industry in the region and the restrictions imposed by the decree of the NPA, the local
70 communities have experienced important changes in their management strategies. In fact, some households
71 from the community within the NPA have shown a trend towards abandoning their MUS and specializing
72 in the provision of ecotourism services (García-Frapolli et al., 2012; Ríos-Beltrán, 2016).

73 These important changes lead us to ask the following questions: is the elimination of the traditional
74 management practices, particularly the milpa agriculture, the only way to ensure the conservation of
75 biodiversity in OMYK? And, how do different productive strategies affect the resilience of this SES in
76 face of some typical disturbances of the region (e.g. hurricanes, fires, tourism-related fluctuations)? These
77 questions have been previously explored in separate works, focusing on either environmental or social
78 aspects. García-Frapolli et al. (2007), using a probabilistic model, proposed that traditional activities are
79 compatible with conservation objectives in the reserve. In another work, García-Frapolli et al. (2012) have
80 proposed that the MUS is a mechanism that promotes household resilience as it diversify the subsistence
81 means. In this work, we address these questions together, using an integrative and dynamical computational
82 model, a novel approach to understand the SES.

83 Dynamical computational models allow us to make reproducible experiments in fully controlled virtual
84 SESs, on relatively short periods of time and without directly affecting the people that inhabit them
85 (Barreteau et al., 2001). Agent Based Models (ABMs) are a type of dynamical model that has been widely
86 used in the study of SESs (An, 2012). In ABMs a system is modeled as a set of individual interacting
87 elements, or agents. Each agent owns a set of variables that describe its state and a set of rules that
88 determine its behaviour (Railsback and Grimm, 2012). Some examples of the use of ABM to study SESs
89 include the study of ecological degradation scenarios on indigenous communities on Amazonian Guyana
90 (Iwamura et al., 2014, 2016); the study of the impact of demographic human changes on the deforestation
91 of the panda habitat on the Wolong reserve, China (An et al., 2005); and the study of the effect of different
92 management strategies on “La Sepultura” reserve, Mexico (Braasch et al., 2018). Boolean Network Models
93 (BNMs) are another type of dynamical models that provides a qualitative representation of a system. BNM
94 are discrete models in which a system is represented as a directed graph whose nodes represent the variables
95 of the system and edges represent regulatory relationships between them (for a detailed description see
96 Saadatpour and Albert (2013)). BNM have been mainly used in the study of molecular and cellular scale
97 systems and it has not been until recently they have been used to study ecological and agroecological
98 systems (e.g. Robeva and Murrugarra (2016); Gaucherel et al. (2017); López-Martínez (2017)).

99 The objective of this study is to explore the effect of different productive and management strategies
100 on biodiversity conservation and on the resilience of the SES associated to OMYK. Using an integrative
101 dynamical computational model, we tested the hypothesis that: (1) traditional milpa agriculture is
102 compatible with biodiversity conservation in OMYK, and (2) that a balanced diversification between
103 traditional and alternative productive activities is a mechanism that promotes resilience.

2 METHODS

104 2.1 Study site

105 OMYK is located in the northeastern region of the Yucatan Peninsula, Mexico, and it has an extension of
106 5367 ha. Mean annual temperature in the region is 24.3°C and the mean annual precipitation is 1,120.2
107 mm (SMN, 2019). The region has a tropical wet and dry climate (Aw2) with a dry season from December
108 to April and a wet season from May to November (CONANP, 2006).

109 Dominant vegetation is medium semi-evergreen forest in different successional stages (García-Frapolli
110 et al., 2007). The forest forms a heterogeneous landscape composed of a mosaic of vegetation in
111 different successional stages as a result of many years of traditional milpa agriculture and multiple
112 natural disturbances such as hurricanes and forest fires (Bonilla-Moheno, 2008; Rangel-Rivera, 2017).

113 The site has a high faunistic diversity and is inhabited by multiple endangered species such as spider
114 monkey (*Ateles geoffroyi*), jaguar (*Panthera onca*), puma (*Puma concolor*), howler monkey (*Alouatta*
115 *pigra*) and others (CONANP, 2006). The spider monkey population inhabiting the reserve has been widely
116 studied (Ramos-Fernández et al., 2018) and has been the main tourist attraction of the site.

117 The management plan of the reserve formally recognises three user communities: Punta Laguna,
118 Campamento Hidalgo and Nuevo Yodzonot (CONANP, 2006). On this project we focus on Punta Laguna,
119 which given its size and closeness to the reserve has experienced important changes after the decree of the
120 NPA. This community is located in the southeast of the reserve over the roadway Cobá-Nuevo Xcan. Punta
121 Laguna has 136 inhabitants distributed in 28 households. All the inhabitants are indigenous Yucatec Maya
122 (Rivera-Núñez, 2014).

123 Before the management plan of the reserve entered into force in 2006, households used to manage their
124 resources following a MUS. They used to manage five different land units (milpa, homegardens, secondary
125 forest, old-growth forest and aquatic systems) and implement 13 different productive activities (milpa
126 agriculture, beekeeping, charcoal production, gather firewood, gather medicinal plants, gather wood for
127 home construction, hunting, fishing, ecotourism, home gardening, sheep herding and scientific research
128 assistance; García-Frapolli et al. (2007)). Nowadays, there is a tendency towards an abandonment of the
129 MUS and of the traditional activities, and a growing trend towards a specialization on tourism-related
130 activities (Rios-Beltrán, 2016; García-Frapolli et al., 2012).

131 There has been continuous research on the site since 1996. The research lines are diverse and include:
132 behavioral ecology of the spider monkeys (Ramos-Fernández et al., 2018), forest succession and restoration
133 (Bonilla-Moheno, 2008; Bonilla-Moheno and Holl, 2010; Bonilla-Moheno, 2010), local systems of nature
134 management (García-Frapolli, 2006; García-Frapolli et al., 2008; Rios-Beltrán, 2016), land cover change
135 analysis (García-Frapolli et al., 2007; Rangel-Rivera, 2017), conservation (Bonilla-Moheno and García-
136 Frapolli, 2012; García-Frapolli, 2015; Rivera-Núñez, 2014) and local institution analysis (García-Frapolli
137 et al., 2013; Rivera-Núñez, 2014).

138 2.2 Description of the model

139 The model was built integrating an ABM with a BNM. As the BNM can be understood as part of the
140 ABM (Table S1; Figure 1) we describe the whole model following the ODD (Overview, Design concepts
141 and Details) protocol for ABMs (Grimm et al., 2010).

142 **2.2.1 Purpose**

143 The purpose of the model is to explore how different productive and management strategies: (1) affect the
144 capacity of the reserve to conserve biodiversity, defined as the forest area and the size of the spider monkey
145 population, and (2) affect the resilience of some elements of this SES to some frequent disturbances in the
146 region. The model has been designed for scientists and managers, mainly those interested in conservation
147 and forest management in the region, and it is aimed at exploring some hypotheses that have been previously
148 proposed, integrating and synthesizing some of the information that has been generated after more than
149 twenty years of multidisciplinary research in the area.

150 **2.2.2 Entities, state variables and scales**

151 The model consists of four agents: landscape patch, household, monkey and fuel biomass.

152 Landscape patches represent a square of 3 ha that can be of one of 5 different types: forest, agriculture
153 (milpa), burned, water or household. Patches of type forest, milpa and burn have a successional age that
154 increases year by year, simulating the regeneration of the forest. When the milpa and burned patches
155 reach a successional age of two years they change their type to forest. Patches of type forest and milpa
156 can accumulate fuel biomass, which makes them susceptible to become burned. When a patch is burned
157 or is converted to milpa, its successional age becomes zero and the fuel biomass accumulated on it is
158 consumed. Patches of type forest can be converted into milpas by households that practice agriculture.
159 After a milpa has been used for two consecutive years, it is abandoned and allowed to regenerate, and the
160 household opens a new plot (García-Frapolli et al., 2007). Water-type patches represent water bodies and
161 the household-type patches represent family homes (built-up constructions). Water and household-type
162 patches remain fixed throughout the simulation.

163 Households have a fixed position in the virtual ABM world. They can only work on four different
164 productive activities: milpa agriculture, apiculture, charcoal production and provide ecotourism services.
165 The activities that households perform remain fixed throughout the simulation. Households obtain a
166 monetary value from each of the productive activities they do. Disturbances, such as hurricanes, forest fires
167 and fluctuations in tourism, have an effect on the productive activities that households do, thus affecting
168 their monetary value.

169 A monkey agent represents a spider monkey. These can live and move in the forest-type patches with a
170 successional age greater than or equal to 30 years. Monkeys can be born or die following a local logistic
171 growth equation (see 2.2.7.8. Monkey population dynamics).

172 Fuel biomass represents an aggregate of decomposing biomass that can ignite a fire. Fuel biomass
173 accumulates on patches after storms and hurricanes. Each fuel biomass unit has a duration time before it
174 disappears. While the fuel biomass is present on a patch, it can cause it to burn.

175 One time step of the model represents two months (see 7.1. Climate). The model is run on a 37 x 94 grid
176 where each grid cell represents 3 ha of landscape.

177 **2.2.3 Process overview and scheduling**

178 Within a bimester, a sequence of processes takes place in the following order. First, the weather conditions
179 of the bimester are determined (i.e., rainy season or dry season). Second, depending on the weather
180 conditions, different meteorological events may occur (e.g., tropical storms, hurricanes) and depending on
181 the type of event a proportion of patches accumulates fuel biomass. Third, the flow level of tourists visiting
182 the reserve is determined. Fourth, households make their productive activities and calculate how much

183 monetary value they obtain. Fifth, forest fires can occur on patches with fuel biomass accumulated. Finally,
184 monkeys can move from patch. Annually (i.e., every 6 iterations), two additional processes follow: patches
185 regenerate and spider monkey population dynamics take place.

186 As simulations are performed along discrete time steps, all the state variables of the agents are updated
187 synchronously.

188 **2.2.4 Design concepts**

189 **2.2.4.1 Emergence**

190 The interaction and feedback between the social and ecological systems emerge as a result of land use
191 changes and the productive activities done by the households.

192 **2.2.4.2 Adaptation/Objectives/Learning/Prediction/Sensing**

193 Agents do not present any adaptive behavior, objectives, learning, prediction nor sensing. The set of
194 activities a household does makes up a household strategy, but households do not change their strategy
195 along the simulation.

196 **2.2.4.3 Interactions**

197 Households interact indirectly with monkeys through the fragmentation and habitat loss caused by milpa
198 agriculture. Monkeys indirectly affect the monetary values of the households by regulating the tourist
199 flow. Households also interact directly with the patches by cultivating them and by removing fuel biomass.
200 Spider monkeys are also affected by high levels of tourist flow.

201 **2.2.4.4 Stochasticity**

202 The occurrence of the different meteorological events is stochastically determined based on occurrence
203 probabilities (see sections below). The selection of the patches that accumulate fuel biomass is random.
204 Tourist flow has a stochastic component. The selection of new parcels for cultivation, the site for placing
205 the beehives and the fuel biomass used by the households is partly random. Forest fires are modeled using
206 a stochastic percolation model. And the movement of the monkeys is partly random.

207 **2.2.4.5 Collectives**

208 Contiguous patches of the same type and within the same range of successional age aggregate into patch
209 neighborhoods. Monkeys living in the same patch neighborhood are aggregated into a local population
210 with its own population dynamics.

211 **2.2.4.6 Observation**

212 The average annual monetary value of the productive activities done by the households (Table S3) is
213 observed to assess the economic state of the households. The mature forest area (successional age >50
214 years), the vegetation type that harbours the greater species richness and most rare species (Bonilla-Moheno,
215 2008), and the total number of spider monkeys, are observed to assess the capacity of the reserve to conserve
216 biodiversity.

217 **2.2.5 Initialization and input**

218 According to literature reports (Rivera-Núñez, 2014), the number of households is set to 28 and the
219 number of monkeys is set to the maximum possible in each habitable patch neighborhood. To create the
220 virtual world, we used a 2003 land use and vegetation map (García-Frapolli et al., 2007). This map was

221 rasterized and inserted into the landscape grid. The successional age of the patches is set to the minimum
222 age according to the range to which they belong on the base map (2-7, 8-15, 16-29, 30-50, >50 years). The
223 model only considers the area that is occupied by the reserve, so all the grid cells outside of the polygon of
224 the reserve are ignored. The model parameters were set to the values shown in Table S2.

225 **2.2.6 Input**

226 The model does not use input data to represent time-varying processes.

227 **2.2.7 Submodels**

228 **2.2.7.1 Climate**

229 Weather conditions are modeled using BNM. This submodel is based on the submodel of the same name
230 of López-Martínez (2017) and is used to give a temporality to the model. The submodel is composed of
231 three nodes: temperature, pressure and precipitation. The interpretation of the state of the nodes, regulation
232 functions and their explanations are shown in Table S1. The dynamics of this submodel generates a periodic
233 attractor of length 6 where the node precipitation remains in one same state for three consecutive states of
234 the attractor and in a qualitatively different state during the next three states. This dynamics is interpreted
235 as an annual climatic dynamic with 6 months of rainy season and 6 months of dry season. This climatic
236 dynamic is characteristic of the region, so each state of the attractor is mapped to one bimester of the year.
237 This mapping allowed us to locate different events throughout the year.

238 **2.2.7.2 Storms**

239 Tropical storms and hurricanes are a frequent disturbance in the region. The strong winds of hurricanes
240 generate defoliation, snapping and uprooting of trees (Bonilla-Moheno, 2010). So, after these disturbances
241 there is a high accumulation of fire biomass that can generate forest fires in the subsequent dry seasons. To
242 simulate this, a proportion of affected patches ($P(s)$) is calculated as a function of the speed of the wind
243 (s). We suppose that the relation between these two variables follows a sigmoid function:

$$P(s) = 100 \frac{1}{1 + e^{p_1(-s+p_2)}}$$

244 where p_1 and p_2 are parameters that determine the shape and position of the inflection point, respectively.
245 The speed of wind (s) takes a value depending on the type of event that occurs on a bimester: no storm (10
246 mph), tropical storm (20 mph) and hurricanes from category one to five (74, 96, 111, 130 and 157 mph). On
247 the rainy season storms and hurricanes can take place with certain occurrence probability estimated with
248 data from NOAA (2019) (Table S2). When an event occurs a unit of fuel biomass is created on an amount
249 of patches of type milpa or forest randomly selected equivalent to the proportion of affected patches ($P(s)$).
250 When created, fuel biomass units are assigned a time of duration from normal distribution (mean duration
251 of fuel biomass, Table S2).

252 **2.2.7.3 Tourism**

253 The tourist flow to the reserve is modeled using BNM. This submodel is composed of two nodes, tourism
254 and tourismH, that together represent three levels of tourists flow: no tourists, low season flow and high
255 season flow. The interpretation of the state of the nodes, regulation functions and their explanation are
256 shown in Table S1. With these rules we obtain a dynamic with two bimester of high season and four of low
257 season that is similar to the tourist dynamic in the region.

258 As hurricanes generally reduce the flow of tourists (García-Frapolli et al., 2012), we assumed that if
259 a hurricane occurs on a bimester, then no tourists arrive (i.e., the nodes tourism and tourismH turn off).
260 Likewise, as the main tourist attraction at the site are spider monkeys (García-Frapolli et al., 2013), we
261 assumed that if there are not at least 20 monkeys at a ratio of 3 km from the households location then no
262 tourists arrive.

263 Finally, a stochastic factor was added to simulate the effect of economic and social disturbances that also
264 reduce the number of tourists. This was simulated through a parameter that with certain probability avoids
265 a high flow of tourists (i.e., the node tourismH is kept turned off; probability of high flow of tourists, Table
266 S2).

267 **2.2.7.4 Household activities**

268 Household activities are also modeled using BNM. This submodel is composed of seven nodes:
269 openMilpa, plantMilpa, youngMilpa, adultMilpa, harvestMilpa, harvestApiculture and charcoalProduction.
270 The interpretation of the state of the nodes, regulation functions and their explanation are shown in Table S1.
271 With these rules we obtained a dynamic that allows to locate the realization of the households' productive
272 activities throughout the year (see 3.1. Submodels build with boolean network modelling). In contrast to
273 the other submodels built with BNM, whose nodes are global (i.e., one set of nodes for all the system), the
274 nodes that constitute this submodel are unique for each household considered in the simulation (i.e., each
275 household has its own set of nodes).

276 To spatially simulate the production of milpa agriculture, when a household opens a milpa it chooses a
277 plot equivalent to 3 ha within a radio of 3 km from its location. Milpas are only opened in patches with
278 successional age greater than or equal to 5 and less than 50 years.

279 To simulate the effect of the hurricanes on the agriculture it is supposed that if a hurricane occurs and the
280 milpa plot of a household is affected by it (i.e., accumulates fossil fuel), then the plants and harvests are
281 lost (i.e., youngMilpa, adultMilpa and harvestMilpa nodes turn off). Likewise, to simulate the effect of
282 hurricanes on apiculture, it is supposed that a household places their beehives in a patch in a radius of 3 km
283 from its location, and if a hurricane occurs and this patch is affected, then the beehives are damaged and
284 there will be no harvests during the next two years. Finally, it is also supposed that households that produce
285 charcoal consume one unit of fuel biomass in a radius of 3 km from its location.

286 Each bimester, households calculate their total bimonthly monetary value from the values shown in Table
287 S3.

288 **2.2.7.5 Forest fires**

289 This submodel uses a modified version of the percolation model of the model library of NetLogo 6.0.4
290 (Wilensky, 2006). In the model forest fires can only occur with certain occurrence probability during the
291 dry season estimated with data from (CONABIO, 2019) (Table S2). If a forest fire occurs, then one of the
292 patches with most fuel biomass accumulated ignites, and with certain probability (burning probability of
293 patches, Table S2) the four neighbour patches containing fuel biomass can also ignite. Then, each burned
294 patch repeats the process.

295 **2.2.7.6 Monkey movement**

296 Spider monkeys can move with certain probability that depends on the successional age of the patch they
297 are on (probability of permanence, Table S2). If a monkey moves, then it randomly shifts to a neighbouring
298 patch with successional age greater than or equal to 30 years. If there are monkeys on patches of type milpa

299 or burned, then they move to one of their nearest patches with successional age greater than or equal to 30
300 years.

301 **2.2.7.7 Forest regeneration**

302 For simplification purposes, we supposed that the regeneration of the forest is deterministic and that the
303 only factor affecting the successional state of the patches is time. So, each year the patches of type milpa,
304 burn and forest increase their successional age.

305 **2.2.7.8 Monkey population dynamics**

306 This submodel is based on the submodel named “Animal meta-population dynamics” of Iwamura et al.
307 (2014). First, patch neighborhoods (set of neighbor patches within the same range of successional age) of
308 late successional forest (30-50 years) and mature forest (>50 years) are formed. Then, for each of these
309 neighborhoods the local population size (N_t) is estimated using the logistic growth equation:

$$N_t = N_{t-1} \left(1 + R \left(1 - \frac{N_t - 1}{k}\right)\right)$$

Where N_{t-1} is the local population size in the time $t - 1$, R is the discrete intrinsic growth rate and k is the carrying capacity of the neighborhood. R was obtained from the literature, and k for the nth neighborhood, denoted k_n , is calculated in the next way:

$$k_n = D_{max} C_n$$

310 Where D_{max} is the maximum density of a patch and C_n is the number of patches that form the nth
311 neighborhood. D_{max} values were obtained from the literature (Table S2).

312 Finally, as it has been proposed that a high flow of tourists can have an effect on the monkey population,
313 (for instance, the noise generated by tourists and the opening of new trails may distress the monkeys
314 affecting their breeding areas and feeding habits; García-Frapolli et al. (2007)), we supposed that the
315 discrete intrinsic growth rate (R) diminishes by a sixth for each bimester that there is a high flow of tourists.

316 **2.3 Implementation**

317 The model was implemented on NetLogo 6.0.4. The code can be found on:
318 <https://github.com/laparcela/OMYKmodel>

319 **2.4 Calibration**

320 We made a manual calibration of the following four parameters related to the extent of the forest fires for
321 which no information was founded on the literature:

- 322 1. burning probability of patches (values explored: 0.4, 0.5, 0.56, 0.6, 0.7),
- 323 2. sigmoid function parameter 1 (p_1) (values explored: 0.05, 0.045, 0.040, 0.035, 0.03, 0.025, 0.02),
- 324 3. sigmoid function parameter 2 (p_2) (values explored: 74, 96, 111), and
- 325 4. mean duration of fuel biomass (values explored: 6, 9, 12).

326 We explored a total of 315 different treatments that represent all the possible combinations of the selected
327 values for each parameter (Figure S1). We run each treatment 100 times over a time span of 12 years (from

328 2003 to 2015) in which households do not produce milpa agriculture (as it was prohibited in 2006). We
329 selected the combination of values for the parameters by applying the following criteria:

330 1. that the total burned area registered by the model was similar to the total area burned observed by
331 Rangel-Rivera (2017) from 2003 to 2015 (i.e., that the empirical data was inside the interquartile range
332 of the model results).

333 2. that the dispersion of the model results was from the ten smallest.

334 3. that the burning probability of patches was the least possible. This was done with the objective of
335 avoiding the fires that affect the complete reserve that appear in longer simulations.

336 4. that the percentage of affected patches when no storm occurs was the lowest possible. This was done
337 to avoid the Forest Fire and Storms submodels to work independently from each other.

338 After applying this criteria hierarchically (i.e., first applying criteria 1, next criteria 2, and so on) we
339 chose the values for the parameters that are shown in Table S2.

340 **2.5 Sensitivity analysis**

341 We carried out a sensitivity analysis on six parameters as a way of examining the consistency (verification)
342 and the general behaviour of the model. The explored variables and the explored values for each were:

343 1. patch size: 1 and 3 ha;

344 2. burning probability of patches: from 0 to 1, by increments of 0.1;

345 3. sigmoid function parameter 1 (p_1): from 0.01 to 0.05, by increments of 0.005;

346 4. sigmoid function parameter 2 (p_2): 74, 96, 111 and 130;

347 5. mean duration of fuel biomass: 6, 9, 12 and 15 bimesters;

348 6. standard deviation of duration of fuel biomass: 1, 3, 6, 9 and 12 bimesters;

349 7. probability of high flow of tourists: from 0 to 1, by increments of 0.1;

350 8. minimum number of monkeys for tourism: from 10 to 35, by increments of 5.

351 We followed the “one at a time” method, in which one parameter is varied at a time while keeping the
352 other fixed on the values shown in Table S2. We ran 30 simulations for each treatment during a virtual time
353 span of 200 years (so that the output variables stabilize) in which households produce the four productive
354 activities. We assessed the sensitivity visually looking for noticeable changes in the model results.

355 **2.6 Validation test**

356 Model results were compared with empirical data of the different vegetation and land use classes in
357 2015 (Rangel-Rivera, 2017) and estimated monkey population size of 2015 (Spaan, 2017). We ran 100
358 simulations over a virtual time span of 12 years (from 2003 to 2015) in which households do not produce
359 milpa agriculture. Model parameters were fixed on the values shown in Table S2. The values of the
360 four parameters for which no data was found in the literature and no calibration was made, were chosen
361 arbitrarily. We look for values that would make simulations faster and that would not generate noticeable
362 changes in the output variables according to the results of the sensitivity analysis.

363 **2.7 Scenarios**

364 To assess the effect of different management strategies on the capacity of the reserve to conserve
365 biodiversity we explore the 16 different possible combinations of the four productive activities that can be

366 implemented by the households (we only explored the cases where all the households implemented the
367 same set of activities). Each treatment was run 30 times during a time span of 50 years. Combinations were
368 classified and averaged following to the characterization of strategies of García-Frapolli et al. (2007) (Table
369 S4):

370 1. Traditional: combinations where households produce agriculture but no ecotourism.
371 2. Mixed: combinations where households produce both milpa and ecotourism.
372 3. Service Oriented: combinations where households dedicated to ecotourism but no agriculture.
373 4. Other: combinations that did not correspond to any of the previous ones.

374 To explore the effect of different management strategies on the resilience of the SES to some frequent
375 disturbances we ran the next set of scenarios:

376 1. We increased the frequency of hurricanes and tropical storms by arbitrarily multiplying the probabilities
377 of occurrence of tropical storms and of each type of hurricane by 3.
378 2. We increased the frequency of forest fires by arbitrarily multiplying their probability by 3.
379 3. We reduced the flow of tourism by arbitrarily reducing the probability of high flow to 0.3.

380 Each scenario was run 30 times during a time span of 50 years.

3 RESULTS

381 3.1 *Submodels built with boolean network modelling*

382 The sequence of system states (periodic attractor) recovered by the three submodels built with BNM (just
383 considering a single household) is shown in Figure 2. Each of the states of the attractor can be interpreted
384 as a “photography” of the approximate conditions of the SES in a given bimester. For instance, the second
385 state of the attractor may be interpreted as the March-April bimester, a bimester of the dry season (node
386 precipitation turned off) in which a high flow of tourists can arrive (nodes tourism and tourismH turned
387 on); in this bimester a household generally prepares the plot for the milpa (node openMilpa turned on);
388 and also in these months a household can obtain good harvests of apiculture and produce charcoal (nodes
389 harvestApiculture and charcoalProduction turned on). Similarly, the fifth state of the attractor may be
390 interpreted as a rainy season bimester (September-October), with a low flow of tourists on which the milpa
391 is almost ready to be harvested and on which households can produce charcoal.

392 3.2 *Sensitivity analysis*

393 We observed the following behavior patterns for the three output variables:

394 1. Mature forest area. In most simulations this variable presents three events of remarkable growth as a
395 result of the initial conditions of the mode and of the wide length of the age categories used to classify
396 the forest. These events take place at 21, 35 and 50 years and correspond to the moment when the
397 patches that started at 30, 16 and 2 years, respectively, reach the mature forest category. After these
398 three events the variable stabilizes.
399 2. Number of monkeys. In most cases this variable decreases slightly during the first 21 years. Next, it
400 gradually increases until it stabilizes.
401 3. Average monetary value. In most simulations this variable stabilizes since the beginning and fluctuates
402 around certain equilibrium values. In some cases the equilibrium values gradually decrease until they
403 reach a new, lower equilibrium.

404 Most output variables were sensible to the modification of the explored parameters. Mature forest area
405 and the total number of monkeys were sensible to the burning probability of patches (Figure 3), the mean
406 duration of the fuel biomass (Figure 4), the two parameters of the sigmoid function (Figures S2-S3), and
407 patch size (Figure S4). Average monetary value was sensible to burning probability of patches (Figure 3),
408 patch size (Figure S4), the probability of high flow of tourists (Figure S5) and the minimum number of
409 monkeys for tourism (Figure S6).

410 **3.3 Validation test**

411 The model reproduced relatively well the area of forest of 2-15, 16-29 and 30-50 years (Figure 5).
412 Nevertheless, the model overestimated the area of mature forest and the total number of monkeys.

413 **3.4 Scenarios**

414 In the scenario without any extra disturbance (Figure 6 first column), mature forest area was greater
415 for Service Oriented and Other strategies. However, there was a growth trend of this variable in all four
416 strategies. Similarly, there was a growth trend of the monkey population in all the strategies, with a larger
417 population increase in Other strategy, followed by the Service Oriented and Traditional strategies, and
418 finally by the Mixed strategy. Average monetary value was markedly higher for the strategies including
419 ecotourism, though the variability was higher as the households depended more on this activity (Table S5).

420 When the storm occurrence probability was increased there was just a slight reduction of the trajectories
421 of the three output variables for all strategies (Figure 6 second column). In the scenario of increased
422 forest fires (Figure 6 third column) there was a considerable reduction of mature forest area and monkey
423 population in all strategies, but there were just slight changes in the average monetary value. In contrast,
424 when tourist flow was reduced (Figure 6 fourth column), there were no noticeable changes in the trajectories
425 for the mature forest area and the number of monkeys, yet there was a remarkable reduction of the average
426 monetary value for the strategies that included this activity. The reduction of the aggregated monetary value
427 under all disturbances was higher for the Service Oriented strategy than for the Mixed strategy (Figure 7,
428 Table S5).

4 DISCUSSION

429 The model proposed here enabled the formal integration of ecological and social data and information
430 collected over the past 20 years, and allowed us to test how different management strategies affect the
431 biodiversity conservation and socio-ecosystemic resilience in the long term. We show that traditional
432 activities like swidden milpa agriculture are compatible with forest and biodiversity conservation, and that
433 a balance of traditional and alternative activities is a way to promote resilience.

434 The results of the sensitivity analysis suggest that the model has an important degree of uncertainty
435 associated with the modification of its parameters. Nevertheless, the behaviors observed in the output
436 variables can be explained by the assumptions of the conceptual model (i.e., the mechanisms proposed
437 in subsection 2.2.7). This fact increases the confidence on the model implementation. In the case of the
438 validation test, the model was able to reproduce quite well the empirically observed area of successional
439 forest categories but failed to reproduce the observed mature forest area and estimated number of monkeys.
440 This may be partly due to the unusually large forest fires that took place between 2006 and 2011, which
441 damaged large areas at the north of the reserve affecting particularly one of the main patches of mature
442 forest (Rangel-Rivera, 2017). It is important to notice that the calibration and the validation test were done
443 with data of the same study. This can explain the good fit of some of the model results. To better test the

444 predictive capacity of the model it will be necessary to compare it with future independent studies. Even if
445 the model could not be correctly validated and even if it may not be useful to make precise quantitative
446 predictions, it reproduces fairly well the mid-term qualitative system behavior described in the literature.
447 As such, the model proved to be a useful tool to explore the joint effect of the considered variables and
448 gain new insights of this SES.

449 The small effect of milpa agriculture on the prevalence and increment of mature forest and monkey
450 population suggests that the production of traditional milpa agriculture is compatible with the conservation
451 of biodiversity in OMYK. These results agree with model predictions obtained by García-Frapolli et al.
452 (2007). These authors, using a probabilistic model, found no significant differences in the reduction of
453 mature forest in scenarios with and without milpa agriculture in OMYK. This compatibility of milpa
454 agriculture and biodiversity conservation is due to multiple factors. First, the low rate at which milpa
455 agriculture was done prior to its prohibition in 2006 did not compromised the extension of remaining
456 older successional forest. Second, before the decree of the reserve, the local community, incentivized by
457 ecotourism, had already agreed on not using the mature forest surrounding the lagoon for agricultural
458 activities (García-Frapolli et al., 2007). This initiative allowed the conservation of one of the main patches
459 of mature forest used by the spider monkeys. As the model results suggest, this zonification, that did not
460 eliminate completely milpa agriculture from the reserve, may have been a sufficient measure to conserve
461 the spider monkey population, as well as many other species that share their habitat. And third, the spider
462 monkeys may be more flexible to the degradation of their habitat than previously thought (Spaan, 2017). In
463 the reserve, spider monkeys are not restricted to mature forest as they also use and travel on late successional
464 forest fragments (30-50 years old; Ramos-Fernández and Ayala-Orozco (2003); Ramos-Fernandez et al.
465 (2013)). This allows them a greater mobility to exploit a wide range of environments in a fragmented
466 habitat. More generally, in spite of the general perception that agricultural activities are a threat to primate
467 biodiversity, there is evidence of primate presence on a diversity of agroecosystems (Estrada et al., 2006).

468 As our results show, the incorporation of ecotourism has brought significant economic benefits to the
469 households. Regular and relatively high incomes from this activity have significantly reduced the need
470 of the household members to search for temporary work outside their communities, improving family
471 well-being (García-Frapolli et al., 2007). However, the abandonment of traditional activities and the MUS
472 reduces the resilience of this SES. The greater resistance to all disturbances and the lesser variability of
473 the Mixed strategy, support the hypothesis that a balanced diversified strategy between traditional and
474 alternative productive activities is a mechanism that promotes resilience. This is explained by the greater
475 functional redundancy and response diversity that this strategy promotes (Biggs et al., 2012). For instance,
476 while in a Service Oriented strategy a reduction of the tourist flow may be strongly perceived by the
477 household, in a Mixed strategy the income reduction can be buffered by the goods obtained on other
478 activities. In general, these results agree with studies that suggest that economic diversification helps rural
479 households reduce poverty and vulnerability (Ellis, 2008; Martin and Lorenzen, 2016; Thulstrup, 2015)
480 and studies that document the greater vulnerability of tourism specialization (Tao and Wall, 2009; Su et al.,
481 2016).

482 Further, specialization also threatens food sovereignty and biocultural diversity. As specialization on
483 ecotourism increases, there is a greater dependency of the communities on external agents and a loss of
484 control over their subsistence. This means, for example, that communities are forced to buy products that
485 they previously produced and decided how to produce. The abandonment of milpa agriculture is also
486 worrying, given that it is a central component of the Yucatec Maya identity and, as such, it has associated a
487 great diversity of knowledge, practices and beliefs which may become lost (Barrera-Bassols and Toledo,

488 2005; Lyver et al., 2019). Similarly, tourism dynamics have led to a process of acculturation, as shown
489 by the way that the traditional ceremonies and rituals have been commodified and are now commercial
490 products for the tourists (García-Frapolli et al., 2018).

491 The disturbance that most affected the mature forest and the monkey population was the increase in forest
492 fires. The importance of this disturbance on the SES was also observed in the sensitivity analysis when
493 modifying the burning probability of patches and the mean duration of biomass fuel. When varying the
494 burning probability of patches we obtained the behaviour characteristic of the percolation models, with
495 a threshold that when exceeded, prevents the growth of the mature forest and the monkey population. It
496 is worth noting that in order to reproduce the empirically observed total burned area, in the calibration,
497 it was found that this parameter has to be set near the critical point. Indeed, it has been proposed that
498 some ecological systems organize near the critical points, particularly regarding the occurrence of forest
499 fires (Ricotta et al., 1999). Similarly, when varying the mean duration of fuel biomass, we found a clear
500 reduction of the mature forest area and of the spider monkey population. This was due to the greater extent
501 of the forest fires given the greater proportion of affected patches at a given point of time. This result
502 suggests that some traditional activities that involve the reduction of the volume of fuel material, as the
503 gathering of firewood and wood for construction and the controlled burning for traditional agriculture, may
504 be important mechanisms to prevent large-scale forest fires. However, as some of these activities are mainly
505 seen as detrimental to the environment and some evidently involve fire risk, they have been prohibited in
506 the reserve (CONANP, 2006).

507 Throughout most of the 20th century, environmental policy on fire on most countries, including Mexico,
508 was based on a fire suppression approach (Martínez-Torres et al., 2016). This approach is responsible
509 for the widely negative perception on fire use on traditional agriculture, as it blamed the practices of
510 rural and indigenous people as the main cause of forest fires (Martínez-Torres et al., 2016). Currently,
511 Mexico environmental policy on fire is on a transition towards an integrated management approach which
512 recognises the role that fires could play in ecosystems (Gutiérrez Navarro et al., 2017). However, as
513 environmental policy on fire responds mainly to ecological concerns, it continues to ignore the practices
514 and knowledge on fire use of rural and indigenous people (Gutiérrez Navarro et al., 2017; Martínez-Torres
515 et al., 2016; Monzón-Alvarado et al., 2014). In the case of the Yucatec Mayas, “wind tenders” (*yum ik’ob*
516 in Yucatec Maya) are the people who carry out the burns in the milpa production. Wind tenders have
517 developed a variety of techniques to control fire, preventing it to expand to the surrounding fields by placing
518 firebreaks, while also assuring the complete conversion of biomass into ashes (Nigh and Diemont, 2013).
519 Given the need of effective large-scale fire prevention strategies and the profound traditional ecological
520 knowledge of the Maya people, it seems essential to revalue local practices and knowledge regarding fire
521 management. We do not suggest that traditional or local practices are ideal, but that they should not be
522 ignored and that the local communities’ perspectives, needs and knowledge should not be excluded from
523 fire management plans. In general, our findings suggest that forest fire prevention strategies in NPAs need
524 to go beyond the recognition and prohibition of productive activities with possible detrimental effects on
525 the environment, and also recognise the role that these activities may play in ecosystem management and
526 fire prevention.

527 The model presented in this work is a simplified representation of a complex SES and has multiple
528 limitations. We discuss four of these. Firstly, the model only considers the area occupied by the reserve.
529 This implies that it ignores all the ecological and social processes that take place outside its administrative
530 borders. For instance, although the habitat of the monkeys may be recovering within the reserve, other
531 processes (e.g., intensification of agricultural activities or expansion of urban areas) may be occurring

532 outside of it that may jeopardize the long-term permanence of monkey populations (e.g., loss of connectivity
533 between habitats in the region). Secondly, the model does not consider the human demographic changes.
534 Thus, we could not explore how population growth may promote the expansion of agriculture and urban
535 areas and the overall consequences of such expansion. However, population growth in OMYK does
536 not necessarily imply an expansion of the agricultural land given the existence of alternative productive
537 activities (García-Frapolli et al., 2007). Thirdly, the income distribution is rarely equitable and accessible to
538 all, as assumed in the model. In OMYK, while the community of Punta Laguna has highly benefited from
539 tourism, other communities have been excluded or have not consolidated initiatives that allow them to enjoy
540 the same benefits (Rivera-Núñez, 2014). Even among the members of Punta Laguna, there has been an
541 exclusion of the households that have remained more attached to their traditional activities (García-Frapolli
542 et al., 2018). Fourth, the model does not consider some of the main current forces of change in this SES.
543 Since the reserve's decree, inhabitants have faced strong conflicts over land tenure and usufruct. There is
544 a conflict between the community members of Punta Laguna and some members of the ejido assembly.
545 For years the ejido, a collective form of land tenure, has demanded a greater economic gratification of
546 ecotourism activities, while the local community members have refused to do so (García-Frapolli et al.,
547 2013). This conflict has led to eviction threats to the local community members. Similarly, in the past
548 years there was an illegal attempt to change the land tenure regime, from collective to private, of the lands
549 occupied by the reserve. With these changes, the land surrounding the reserve would be modified from
550 collective to private parcels, which could then be sold (García-Frapolli et al., 2007). This would enable the
551 entry of external agents, quite likely from the tourism development and real estate industries.

552 To our knowledge, this is the first work on the study of SESs that explicitly integrates BNMs and
553 ABMs. One of the main constraints of the latter is that they require large amounts of empirical data to be
554 parameterized (Iwamura et al., 2014). In contrast, BNMs require no or few parameters, making them a
555 suitable tool for modeling scarcely characterized systems (Saadatpour and Albert, 2013). In this sense, the
556 qualitative approach offered by the BNMs can facilitate and simplify the construction of models for the
557 study of SES. Our model can be used as a guide or support tool for future integrative studies in OMYK.
558 Similarly, the model can be used as a guide towards the development of support tools in participatory
559 processes. This type of tools can help in the communication of knowledge, promote collective reflection and
560 learning, and thus facilitate collective and adaptive decision-making for building sustainable management
561 practices and effective governance (Braasch et al., 2018; D'Aquino et al., 2003).

562 In conclusion, the results of this study suggest that: (1) conservation strategies that do not exclude
563 traditional productive activities can be compatible with biodiversity conservation, and (2) a balanced
564 economic diversification between traditional and alternative activities, is a strategy that promotes the
565 economic resilience of rural households. In addition, the model presented here shows how computational
566 modeling and the SES perspective are effective means of integrating and synthesizing information from
567 different sources. Finally, our work also illustrates some of the main limitations of systems perspective and
568 computational modelling on the study of SES, such as the difficulty in considering political, cultural and
569 historical features.

CONFLICT OF INTEREST STATEMENT

570 The authors declare no conflict of interest.

AUTHOR CONTRIBUTIONS

571 LGGJ, MB and GRF conducted the research. EGF, MBM and CRR provided data and maps for the
572 model. LGGJ wrote original draft. All authors reviewed and edited the paper.

ACKNOWLEDGMENTS

573 We thank the community of Punta Laguna for their support of academic research over the years. We also
574 thank Celene Espadas, Fernanda Figueroa, Fernanda Ríos-Beltrán, Tlacaelel Rivera-Núñez and Ernesto
575 Vega for their commentaries on previous versions of this work. M.B. acknowledges financial support from
576 UNAM-DGAPA-PAPIIT (IN207819). G.R.F. acknowledges support from CONACYT (grant 157353)
577 and Instituto Politécnico Nacional.

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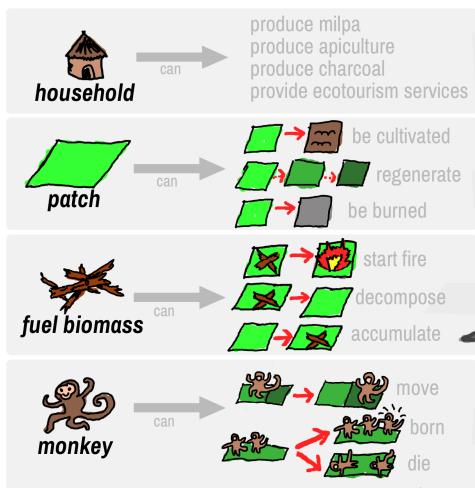
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FIGURE CAPTIONS

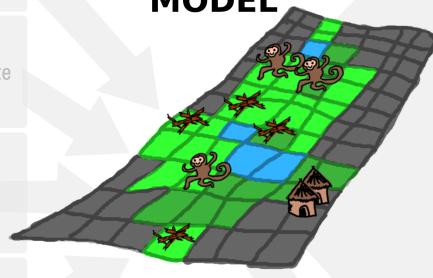
AGENTS



SUBMODELS



MODEL



BASE MAP

García-Frapolli et al. (2006)

OUTPUT

Area of mature forest
Numer of monkeys
Average monetary value

Figure 1. Representation of the model. The model is composed of four different types of agents. The world where the agents interact is based on a land use and vegetation map. The model simulates eight different processes, or submodels, six of which occur every two months (each time step) and the last two only occur annually (each six time steps). The state of the modeled system is monitored through three output variables.

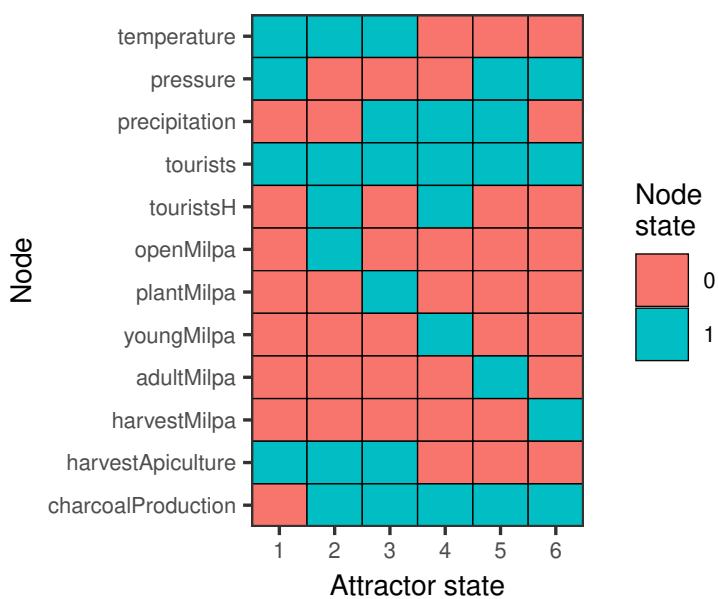


Figure 2. Attractor recovered by the submodels built with boolean network modelling. Rows represent each node of the network, columns are the attractor states and each color represents a different state of the node. Each of the attractor states of this cyclic attractor of period six can be interpreted as a snapshot of the approximate state of the SES in a given bimester (See Subsection 3.1. Submodels built with boolean network modelling).

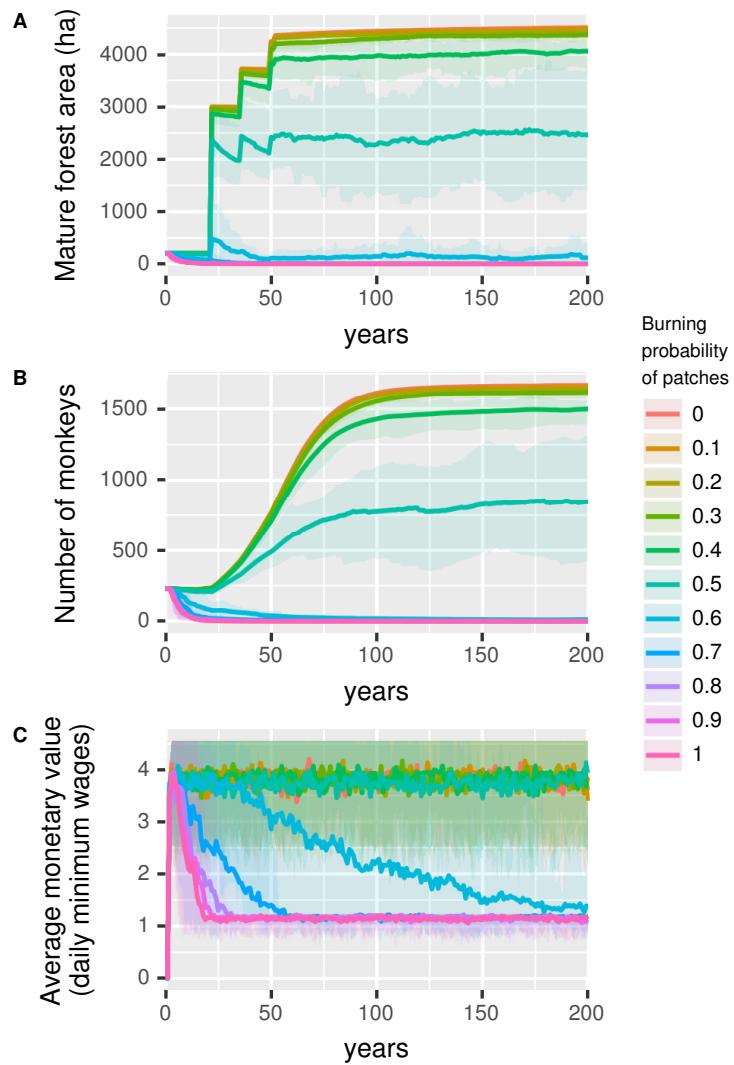


Figure 3. Sensitivity analysis of the burning probability of patches. Each panel represents the change in an output variable for a 200 year period and 30 simulation runs for each value of the burning probability of patches: A): Mature forest area in ha (forest greater than 50 years of successional age). B): total number of monkeys. C): average monetary value of activities done by the households in daily minimum wages. Lines represent the mean and upper and lower shadows represent 5th and 95th quartiles, respectively, for each value of the burning probability of patches (different colors).

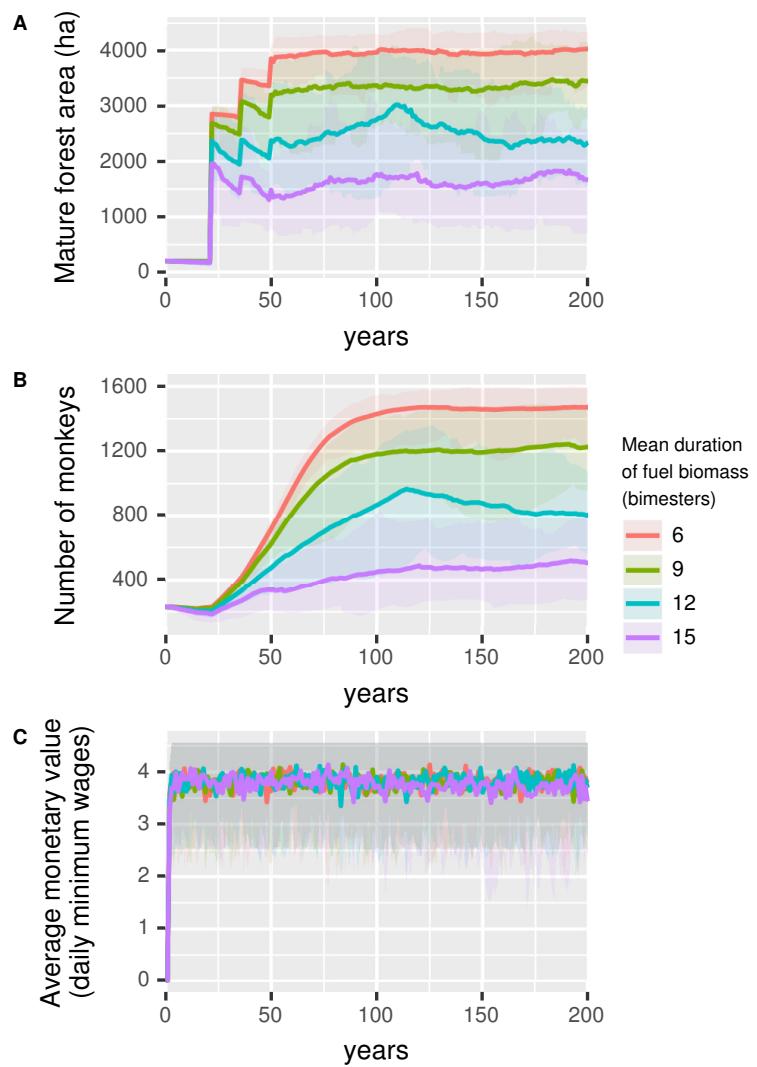


Figure 4. Sensitivity analysis of the mean duration of fuel biomass (bimesters). Each panel represents the change in an output variable for a 200 year period and 30 simulation runs for each value of the mean duration of fuel biomass. A): Mature forest area in ha (forest with successional age greater than 50 years). B): total number of monkeys. C): average monetary value of activities done by the households in daily minimum wages. Lines represent the mean and upper and lower shadows represent 5th and 95th quartiles, respectively, for each value of the mean duration of fuel biomass (different colors).

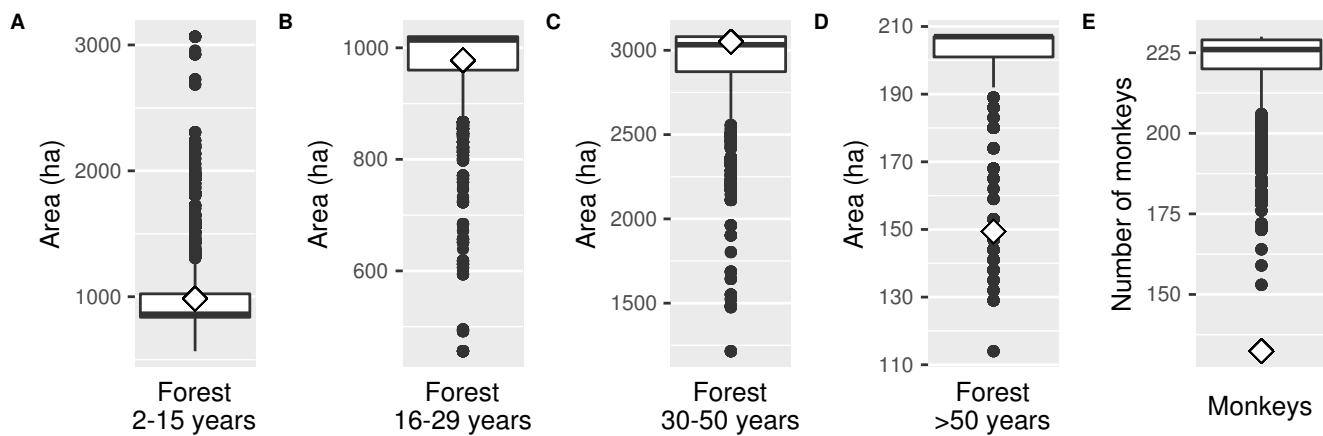


Figure 5. Validation test results. Results of simulations and empirical values for five different variables. A): area of forest with successional age between 2 and 7 years. B): area of forest with successional age between 16 and 29 years. C): area of forest with successional age between 30 and 50 years. D): area of forest with successional age greater than 50 years. E): total number of monkeys in the reserve. Boxplots are the model outputs after 12 years (from 2003 to 2015) for 100 simulation runs. Diamonds represent the empirical observed values for different forest types in 2015 and the total estimated number of monkeys in 2015.

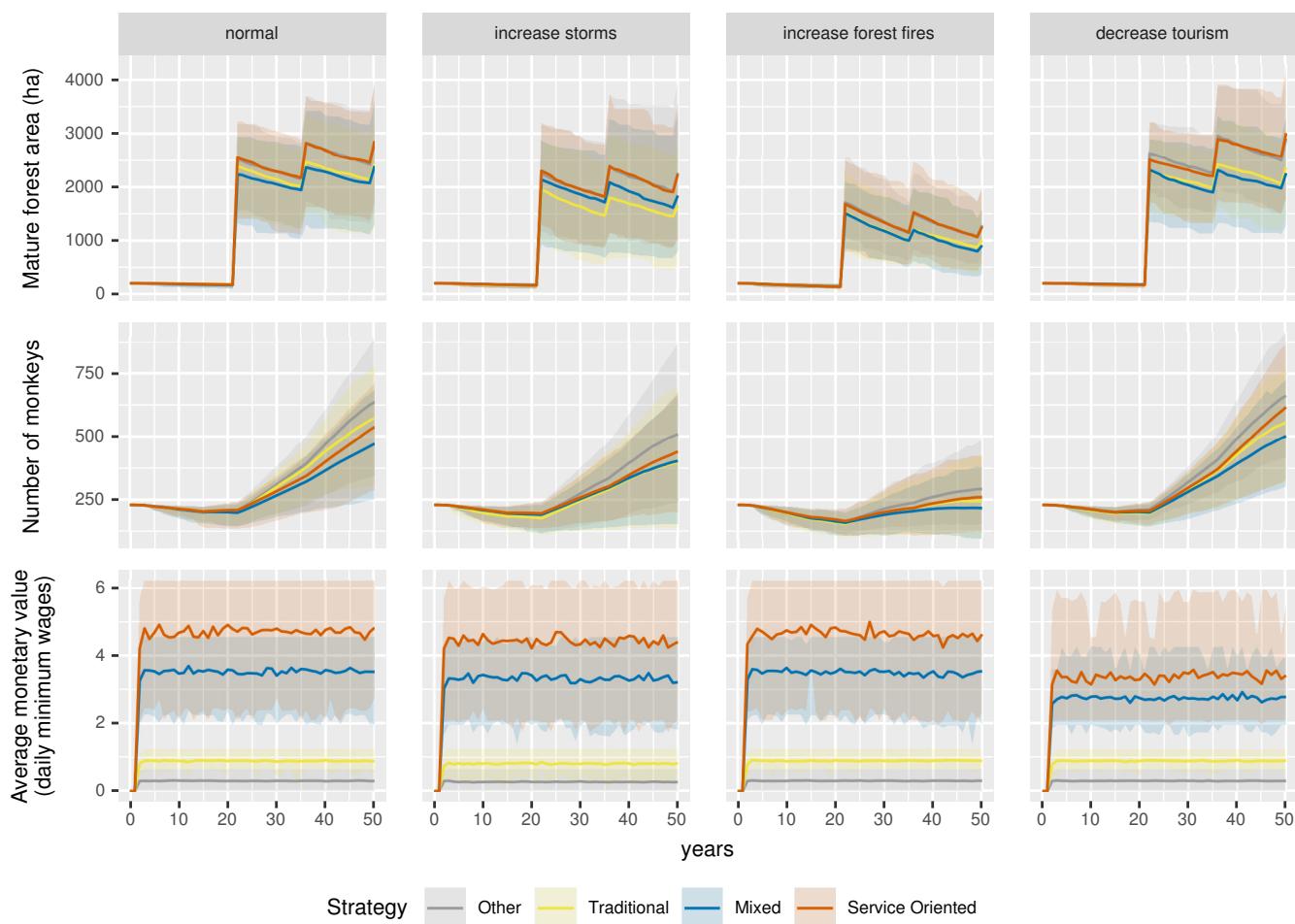


Figure 6. Model results for different strategies under different scenarios. Each column of graphs represents a different scenario and each row a different output variable. The four scenarios shown are: a scenario with normal level of all disturbances, a scenario with an increase in the storms and hurricane occurrence probability, a scenario with an increase in the probability of occurrence of forest fires and a scenario with a decrease in the probability of the flow of tourists. First row: Mature forest area in ha (forest with successional age greater than 50 years). Second row: total number of monkeys. Third row: average monetary value of activities done by the households in daily minimum wages. Lines represent the mean and upper and lower shadows represent 5th and 95th quartiles, respectively, for different strategies (in colors), each ran over a 50 year period and for 30 simulation runs.

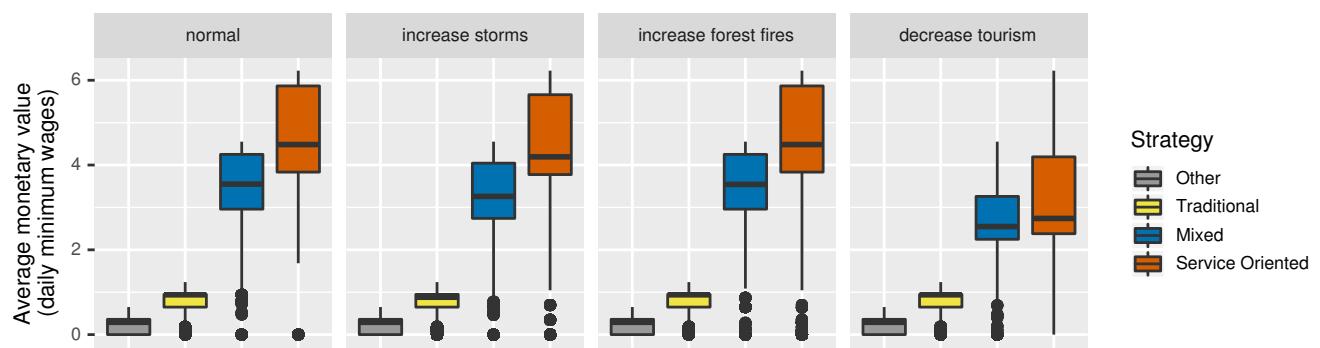


Figure 7. Aggregate average monetary value for different strategies under different disturbance scenarios. The four scenarios shown are: a scenario with normal level of all disturbances, a scenario with an increase in the storms and hurricane occurrence probability, a scenario with an increase in the probability of occurrence of forest fires and a scenario with a decrease in the probability of the flow of tourists. Boxplots represent the yearly average monetary value of different strategies (in colors) for 30 simulations each ran over a 50 year period.