

1 Core species and interactions prominent in fish- 2 associated microbiome dynamics

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4 Daii Yajima¹, Hiroaki Fujita¹, Ibuki Hayashi¹, Genta Shima¹, Kenta Suzuki² and Hirokazu
5 Toju^{1†}

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7 ¹Center for Ecological Research, Kyoto University, Otsu, Shiga 520-2133, Japan

8 ²Integrated Bioresource Information Division, BioResource Research Center, RIKEN,
9 Tsukuba, Ibaraki 305-0074, Japan

10

11 [†]**Correspondence:** Hirokazu Toju (toju.hirokazu.4c@kyoto-u.ac.jp).

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13 This article includes 6 Figures, 10 Extended Data Figures, and 1 Supplementary Table 1.

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16

17 **Abstract**

18 In aquatic ecosystems, the health of fish depends greatly on the dynamics of microbial
19 community structure in the background environment. Nonetheless, finding microbes with
20 profound impacts on fish's performance out of thousands of candidate species remains a
21 major challenge. We here show that time-series analyses of microbial population dynamics
22 illuminate core components and structure of fish-associated microbiomes. By targeting eel
23 aquaculture microbiomes as model systems, we reconstructed the population dynamics of
24 9,605 bacterial and 303 archaeal species/strains across 128 days. Due to the remarkable
25 increase/decrease of constituent microbial populations, the taxonomic compositions of
26 microbiomes changed drastically through time. We then found that some specific microbial
27 taxa showed positive relationship with eels' activity level even after excluding cofounding
28 effects of environmental parameters (pH and dissolved oxygen level) on population dynamics.
29 In particular, a vitamin B₁₂-producing bacteria, *Cetobacterium somerae*, consistently showed
30 strong positive associations with eels' activity level across the replicate time-series of the five
31 aquaculture tanks. Network theoretical and metabolic modeling analyses further suggested
32 that the highlighted bacterium formed compartments of close microbe-to-microbe interactions
33 with some other bacterial taxa, forming potential core microbiomes with positive impacts on
34 eels. Overall, these results suggest that integration of microbiology, ecological theory, and
35 network science allows us to explore core species and interactions embedded within complex
36 dynamics of fish-associated microbiomes.

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39 Microbial communities are essential factors of the life of vertebrates^{1–4}, playing key roles in
40 the development and homeostasis of their hosts^{5–7}. Gut microbiomes, for example, play key
41 roles in the nutrition and disease prevention of human and other mammal species^{8,9}. Such
42 physiological and ecological effects of gut microbes on hosts have been reported as well for
43 fish^{6,10,11}. Meanwhile, because fish are continuously exposed to numerous pathogenic and
44 non-pathogenic microbial species in the water, their performance (or fitness) depends not only
45 on gut-associated microbes^{6,10} but also on the microbiomes of the background environment^{12–}
46 ¹⁴. Therefore, finding key microbiome components whose dynamics determine fish's health or
47 performance is of interdisciplinary interest spanning from microbiology to zoology and
48 environmental science. However, due to the tremendous diversity of bacteria and archaea in
49 aquatic ecosystems^{15,16}, exploring such core microbial species associated with fish health
50 remains a challenge.

51 A starting point for finding fish-health-associated microbes in aquatic ecosystems is to
52 track the dynamics of microbial community compositions. Nonetheless, we still have limited
53 knowledge of the extent to which structure of fish-associated microbiomes change through
54 time. Although time-series data of microbiomes have become available in pioneering projects
55 of human-associated microbes^{17,18}, few attempts have been made to monitor microbiomes
56 associated with other animals over tens of time points. Moreover, continuous sampling of
57 fecal samples of targeted vertebrate individuals is generally much harder in aquatic
58 environments than in terrestrial environments. Thus, developing model systems for time-
59 series analyses of microbe–fish ecological interactions is a demanding but essential step for
60 exploring core bacteria and/or archaea out of thousands of candidate species in microbial
61 communities.

62 Despite the hardship in gaining time-series microbiome samples at the individual level,
63 fish-associated microbiome dynamics can be monitored at the population or community level
64 by sampling environmental water samples^{19–21}. Because excrements of fish are released to
65 water, samples of background water are expected to reflect gut microbiomes of fish
66 populations or communities. Furthermore, as individual fish are continuously exposed to the
67 background microbiomes, analyses of water samples provide essential insights into
68 surrounding environmental conditions and potential sources of gut microbiomes^{12–14}. In this
69 respect, time-series analyses of aquaculture or aquarium systems offer an ideal opportunity
70 for investigating relationship between microbial community structure, core microbial species,
71 and vertebrate health.

72 By targeting a recirculating aquaculture system of the Japanese eel (*Anguilla*
73 *japonica*), we herein integrate microbiology, community ecology, and network science for
74 detecting key species and structure within fish-associated microbiomes. Based on the DNA
75 metabarcoding of prokaryote (bacterial and archaeal) communities for the 128-day time-
76 series, we revealed to what extent the compositions of aquaculture microbiomes fluctuate
77 through time. We then reconstructed the population dynamics (i.e., increase/decrease) of the
78 9,908 microbial amplicon sequencing variants (ASVs) constituting the aquaculture
79 microbiomes, screening bacteria or archaea whose abundance was tightly linked with the
80 health condition of eels. We then found that several microbial ASVs showed positive
81 associations with eel health consistently across the five replicate aquaculture tanks, even after
82 controlling the effects of their environmental preference (e.g., preference to pH and dissolved
83 oxygen level). With the approaches of network science and metabolic modeling, we further
84 examined potential interactions between the core microbes. Overall, this study illustrates how
85 core species and interactions are detected based on time-series datasets of microbiome
86 dynamics.

87

88 **Results**

89 **Microbiome dynamics.** Monitoring of microbiome dynamics was conducted targeting the
90 five water tanks of an aquaculture farm of the Japanese eel. In each water tank (diameter = 2.5
91 m; height = 1 m; volume = 20 m³), 1,400–4,300 eel individuals (average weight = 80–130 g)
92 had been kept. The pH and dissolved oxygen (DO) concentrations were recorded for each
93 tank every day. In addition, as a measure of ecosystem-level functions of microbiomes, the
94 health condition of eels was evaluated based on eight criteria, yielding eel activity scores on a
95 scale of 0 to 40 (see Methods). For the analyses of microbiome dynamics, water was sampled
96 from each aquaculture tank every 24 hours during 128 days. By applying a quantitative
97 amplicon sequencing approach for estimating 16S ribosomal RNA gene (16S rRNA) copy
98 concentrations of respective microbes^{22,23}, we obtained time-series datasets representing the
99 increase/decrease of 9,605 bacterial and 303 archaeal ASVs representing 618 genera and 325
100 families (Fig. 1a). Thus, our data offered a novel opportunity to test synchronizations among
101 microbial population dynamics, environmental factors (pH and DO), and vertebrate
102 performance (eel activity level).

103 At the community level, drastic taxonomic turnover was observed in the timeseries of
104 each aquaculture tank (Fig. 1; Extended Data Figs. 1-3). In Tanks 1 and 2, for example, the
105 community structure characterized by the predominance of Fusobacteriaceae and

106 Microbacteriaceae was suddenly altered by a Flavobacteriaceae-dominated state around Day
107 45 (Fig. 1). Meanwhile, microbiomes of Tanks 3–5 displayed more complex dynamics
108 represented by frequent shifts between Flavobacteriaceae-dominated and Chitinophagaceae-
109 dominated states, although clear classification of community states was difficult (Fig. 1).

110 A multivariate analysis of the prokaryote community structure indicated that the
111 community state characterized by dominance of Fusobacteriaceae and Microbacteriaceae was
112 associated with high eels' activity (Fig. 2). In contrast, the Flavobacteriaceae-dominated and
113 Chitinophagaceae-dominated states, which were observed in high-pH conditions, were
114 associated with low eels' activity (Fig. 2). At the genus level, the high-eel-activity-related
115 state of dominance by Fusobacteriaceae and Microbacteriaceae was characterized by high
116 relative abundance of *Cetobacterium*, which includes species potentially contribute to fish
117 physiological homeostasis²⁴. On the other hand, the Flavobacteriaceae-dominated and
118 Chitinophagaceae-dominated states associated with low eels' activity were represented by
119 *Flavobacterium* and *Edaphobaculum*, respectively (Extended Data Fig. 4). Among the
120 genera, *Flavobacterium* includes fish pathogens²⁵, while *Edaphobaculum*²⁶ has been poorly
121 investigated in terms of their effects on fish physiology. These results suggest potential
122 impacts of environmental microbiome dynamics on fish health/behavior in aquaculture
123 systems.

124

125 **Exploring microbes with key roles.** We next evaluated how the population dynamics of
126 each microbial ASV were associated with environmental variables and eels' activity level.
127 Specifically, we examined how population size (absolute abundance) of each ASV varied
128 with pH, DO, and eels' activity level (Fig. 3a) based on correlation analyses with twin-
129 surrogate permutations²⁷ (Fig. 3b-c). The ASVs varied in their environmental preference for
130 pH and DO conditions as well as in their associations with eels' activity level (Extended Data
131 Fig. 5). We also found that ASVs' relationship with eels' activity level displayed tank-
132 dependent complex associations with pH or DO preference (Fig. 3d-e). Thus, for each
133 microbial ASV in each aquaculture tank, we calculated a partial correlation between absolute
134 abundance and eels' activity scores through the time-series by controlling the effects of pH or
135 DO. Because partial correlation coefficients were consistent between the pH-controlled and
136 DO-controlled calculations (Fig. 3f), the pH-controlled partial correlation coefficients were
137 used in the following analyses.

138 The partial correlation coefficients with eels' activity level varied greatly depending on
139 prokaryote taxa (Fig. 2e). Nonetheless, ASVs belonging to some bacterial genera showed

140 consistently positive correlation with eels' activity scores across the five tanks (Fig. 2g;
141 Extended Data Fig. 5). The list of those ASVs included bacteria belonging to the genera
142 *Cetobacterium* (Fusobacteriaceae; Fusobacteriia; ASV ID = X_0002), *Plesiomonas*
143 (Enterobacteriaceae; Gammaproteobacteria; X_0020), *Turicibacter* (Erysipelotrichaceae;
144 Bacilli; X_0041), *Paraclostridium* (Clostridiaceae; Clostridia; X_0014), *Romboutsia*
145 (Peptostreptococcaceae; Clostridia; X_0028), *Edwardsiella* (Hafniaceae;
146 Gammaproteobacteria; X_0027), *Clostridium* (Clostridiaceae; Clostridia; X_0029), and an
147 ASV belonging to Barnesiellaceae (Bacteroidia; X_0064) (Fig. 3g; Extended Data Fig. 5d).

148 An additional database search of the 16S rRNA sequences suggested that some of the
149 ASVs with positive associations with eels' activity level belonged to bacterial species with
150 potential physiological impacts on fish. For example, the *Cetobacterium* ASV, which showed
151 strongest positive partial correlation with eels' activity level, was represented by the 16S
152 rRNA sequences completely matching that of *Cetobacterium somerae* (formerly recognized
153 as “*Bacteroides* type A”) in the NCBI nucleotide database. This *Cetobacterium* species has
154 been known to produce high concentrations of vitamin B₁₂ and hence their potential
155 contributions to fish's physiology have been anticipated. Meanwhile, the *Edwardsiella* ASV
156 listed above was allied to the notorious fish pathogen *E. tarda*²⁸, illuminating paradoxical
157 relationships with eels' health. However, our supplementary phylogenetic analysis based on
158 the *sodB* gene marker²⁹ indicated that 95.1 % of *Edwardsiella* bacteria detected in the focal
159 eel aquaculture system belonged to non-pathogenic clades^{29,30} within the genus *Edwardsiella*
160 (Extended Data Fig. 6).

161 In terms of negative impacts on eels' activity level, bacteria in the genera *Aeromonas*
162 (Aeromonadaceae; Gammaproteobacteria), *Methylobacterium* (alternatively, *Methylorubrum*;
163 Beijerinckiaceae; Alphaproteobacteria), and *Acinetobacter* (Moraxellaceae;
164 Gammaproteobacteria) were highlighted (Fig. 3g). Among them, *Aeromonas* and
165 *Acinetobacter* have been known to include fish pathogens^{31,32}. At the ASV level, an ASV
166 allied to the cvE6 clade within the order Chlamydiales (Chlamydiae; Verrucomicrobiota)
167 showed strongest negative correlation with eels' activity scores (Extended Data Fig. 5d).

168 Although the above analysis controlling environmental preferences of respective
169 bacteria allows high-throughput screening for species with potential positive/negative impacts
170 on target biological functions, the simple statistical approach with partial correlation analyses
171 precludes insights into the direction of causality. Specifically, it is important to consider the
172 possibility that high/low abundance of an ASV is a consequence but not a cause of eels'
173 high/low activity. Therefore, we performed an additional analysis introducing time lags into

174 eels' activity scores throughout the time-series. We then found that the abundance of the
175 *Cetobacterium* ASV was positively correlated with eels' activity scores of the next day, while
176 correlations between *Cetobacterium* abundance and past eels' activity scores were much
177 lower than those with no time lags (Fig. 3h). Meanwhile, high correlation between 5-days-ago
178 eels' activity level and present-day *Cetobacterium* abundance was observed in some tanks
179 (Tanks 1 and 4; Fig. 3h), illuminating the importance of carefully interpreting the results of
180 the time-series analysis.

181

182 **Networks of interactions.** We then reconstructed webs of potential microbe-to-microbe
183 interactions to illuminate microbial groups or interactions positively associated with eels'
184 health. We first applied the Meinshausen-Bühlmann (MB) method, which was designed to
185 evaluate patterns of coexistence realized by the effects of microbe–microbe interactions as
186 well as those of niche sharing between microbes. For each aquaculture tank, the reconstructed
187 network of microbe–microbe coexistence (Extended Data Figs. 7–8; Supplementary Table 1)
188 was compartmentalized into several modules, which differed in mean partial correlations with
189 eels' activity scores (Fig. 4). We then found that each of the five networks included a module
190 constituted by the abovementioned *Cetobacterium* ASV and several other ASVs with
191 consistently positive associations with eels' activity level (Fig. 4; Extended Data Fig. 9). The
192 bacteria consistently formed network modules of coexistence with the *Cetobacterium* ASV
193 were *Plesiomonas* (X_0020), *Turicibacter* (X_0041), *Paraclostridium* (X_0014), *Romboutsia*
194 (X_0028), *Edwardsiella* (X_0027), and *Clostridium* (X_0029) (Fig. 4).

195 To infer the presence/absence of direct interactions between these bacteria with positive
196 relationship with eels' activity level, we conducted an additional network analysis based on
197 the sparse and low-rank (SLR) decomposition method, which allowed us to remove latent
198 effects of environmental conditions. In the networks reconstructed with the SLR method (Fig.
199 5), potential effects of niche sharing were controlled and hence the links between bacterial
200 ASVs were expected to represent potential positive interactions. The estimated interaction
201 coefficients were highly correlated between the MB and SLR methods (Extended Data Fig.
202 10). Meanwhile, in the SLR-based network, removing the effects of potential niche sharing
203 (sharing of environmental preference) resulted in the simplification of network structure, in
204 which estimated direct interactions between microbes were focused (Fig. 5). Despite the
205 considerable difference between MB- and SLR-based network topology, the *Cetobacterium*
206 ASV with the strongest associations with eels' activity level was, again, linked with the

207 *Plesiomonas*, *Turicibacter*, *Paraclostridium*, *Romboutsia*, *Edwardsiella*, and *Clostridium*
208 ASVs within the SLR network (Fig. 5), suggesting positive interactions with these bacteria.

209

210 **Potential metabolic interactions.** To estimate functional interactions between microbes, we
211 focused on genomic compositions of respective microbes within the aquaculture
212 microbiomes. After retrieving the information of genomic compositions from reference
213 databases, we analyzed the inferred gene repertoires (KEGG metabolic pathway/process
214 profiles) of the microbial ASVs based on multivariate analysis. Along the principal
215 component axes, *Cetobacterium*, which showed consistent correlations with eels' activity
216 (Fig. 3g; Extended Data Fig. 5d), was located distantly from *Edwardsiella*, *Plesiomonas*, and
217 *Turicibacter* (Fig. 6a). In contrast, *Romboutsia*, *Paraclostridium*, and *Clostridium* displayed
218 similar metabolic gene repertoires with *Cetobacterium* (Fig. 6a).

219 We next evaluated potential competitive and facilitative interactions between microbes
220 based on a genome-scale metabolic modeling approach. In the analysis, reference genomic
221 information was used to infer competition for available resources and exchanges of
222 metabolites, yielding metabolic resource overlap and metabolic interaction potential scores
223 for each pair of microbial ASVs. We then found that the *Romboutsia*, *Edwardsiella*, and
224 *Plesiomonas* ASVs had relatively low metabolic resource overlap and relatively high
225 metabolic interaction potential with the *Cetobacterium* ASV among the prokaryotes examined
226 (Fig. 6b).

227

228 **Discussion**

229 Through the 128-day monitoring of thousands of microbial species/strains, we here found that
230 aquatic microbiomes associated with fish could show drastic shifts of community structure
231 through time. Such dynamical nature of community processes has been intensively
232 investigated in human-associated microbiomes in light of potential influence on host
233 status^{17,18}. In particular, shifts (collapse) of microbial community structure to disease-related
234 states (i.e., dysbiosis) have been considered as essential mechanisms determining human
235 health^{33,34}. Given the growing literature on microbiome dynamics in medical science,
236 knowledge of shifts between alternative states of fish-related microbiomes¹⁴ is expected to
237 shed new lights on physiological and ecological processes of vertebrates.

238 The aquaculture microbiome dynamics were described as shifts among
239 Fusobacteriaceae-abundant states, Flavobacteriaceae-dominated states, and
240 Chitinophagaceae-dominated states, although intermediate states existed through the time-
241 series (Fig. 1). Among them, Fusobacteriaceae-abundant states, which were characterized by
242 high abundance of *Cetobacterium*, were designated as microbiome compositions positively
243 associated with eels' activity level (Fig. 2; Extended Data Fig. 4). In fact, among the 9,908
244 microbial ASVs examined, the ASV representing *Cetobacterium somerae* showed the
245 strongest associations with eel's activity level through the time-series even after controlling
246 effects of environmental preference (Fig. 3; Extended Data Fig. 5). This *Cetobacterium*
247 species has been reported from a broad taxonomic range of freshwater fish³⁵⁻³⁷, especially
248 from intestines of species that do not require dietary vitamin B₁₂²⁴. Although vitamin B₁₂
249 (cobalamin) plays essential roles in animal physiology (e.g., normal functioning of nervous
250 systems and the maturation of red blood cells), they can be synthesized only by specific
251 clades of bacteria and archaea^{38,39}. Genomic studies have shown that *C. somerae* has a series
252 of genes for anaerobic vitamin B₁₂ biosynthesis⁴⁰. Indeed, the bacterium produces highest
253 concentrations of vitamin B₁₂ compared to other culturable bacteria within freshwater fish-
254 associated microbiomes^{24,41}. Given the prevalence of *Cetobacterium* in freshwater fish
255 species³⁵⁻³⁷, our results suggest that maintaining microbiomes at *Cetobacterium*-abundant
256 states is the key to build general platforms for stably keeping freshwater
257 aquaculture/aquarium systems.

258 Further analyses based on network theory and metabolic modeling indicated the
259 possibility that the *Cetobacterium* species form facilitative interactions with some other
260 microbial species/ASVs (Figs. 4-6). Among the bacteria for which interactions with
261 *Cetobacterium* were inferred from multiple analyses, *Edwardsiella tarda* has been known to
262 include notorious pathogens of broad taxonomic groups of fish including eels^{28,29,42}. However,
263 we found that the *E. tarda* population of the investigated aquaculture system was dominated
264 by non-pathogenic strains^{29,30} of the species (Extended Data Fig. 6). Thus, the presence of
265 microbial species/strains belonging to broadly-known taxa of pathogens do not necessarily
266 result in negative impacts on fish. Rather, our analyses suggested that “seemingly pathogenic”
267 microbes could be involved in core microbiome components (network modules) constituted
268 by microbes contributing to the maintenance of fish health. Further studies are awaited to
269 explore potential mechanisms such as competitive exclusion of pathogenic strains by non-
270 pathogenic strains^{43,44} or indirect negative impacts on pathogenic strains through the
271 activation of fish immune systems^{6,45} by non-pathogenic strains. In contrast to *E. tarda*,
272 *Romboutsia* and *Plesiomonas*, which were inferred as microbes with facilitative interactions

273 with *C. somerae*, too (Figs. 4-6), have been poorly investigated in terms of their functions.
274 Their potential roles in competitive exclusion of pathogens or activation of host immune
275 systems deserve further investigations.

276 While the time-series dataset allowed us to highlight core species and interactions
277 within microbial communities, more sophisticated statistical platforms beyond simple
278 correlational approaches are necessary for confirming causative relationships between
279 microbiome dynamics and vertebrate health/performance. In this respect, methods based on
280 nonlinear mechanics, such as transfer entropy and empirical dynamic modeling^{46,47}, are
281 expected to help us infer causative interactions among microbial population dynamics,
282 environmental factors, and vertebrate performance. Albeit promising, these methods require
283 substantial computational resources when we try to analyze microbiomes consisting of
284 thousands of ASVs. Further methodological advances will deepen our understanding of the
285 mechanisms by which microbiome dynamics and vertebrate performance are linked with each
286 other.

287 As the analyses of microbiome dynamics extend from medical science to researches
288 targeting other vertebrates, we will be more and more aware of overlooked roles of microbes
289 in both terrestrial and aquatic ecosystems. Feedback between intestine and environmental
290 microbiomes, for example, deserves future intensive research in terms of potential great
291 impacts on ecosystem processes. In particular, given that aquatic vertebrates are continuously
292 exposed to excrements of other individuals or species, their gut microbiome dynamics (and
293 related health conditions) may be more likely to be synchronized at the population or
294 community levels than those of terrestrial vertebrates. Therefore, simultaneous monitoring of
295 intestine and background environmental microbiomes will provide platforms for uncovering
296 such feedback and synchronization processes. Further insights into fish-associated
297 microbiome dynamics will reorganize our basic understanding of aquatic ecosystem
298 dynamics, advancing technologies for sustainable food production through stable aquaculture
299 systems⁴⁸⁻⁵⁰.

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472 reconstruction of genome-scale metabolic models for microbial species and
473 communities. *Nucleic Acids Res* **46**, 7542–7553 (2018).

474

475

476 Methods

477 **Sampling.** Monitoring of microbiome dynamics was conducted targeting the five water tanks
478 of the eel aquaculture system of A-Zero Inc. (Nishiawakura, Okayama Prefecture, Japan). In
479 each water tank (diameter = 2.5 m; height = 1 m; volume = 20 m³), 1,400–4,300 eel
480 individuals (average weight = 80–130 g) had been kept. About 10 % of tank water was
481 replaced with warmed fresh well water every day, and the water temperature in the tanks was
482 kept at around 30 °C. The drainage from the five tanks were mixed and processed in a series
483 of filtration equipment. The filtered drainage was returned to each tank after being processed
484 in another filtration equipment adjacent to each tank. The eels were fed with mixture of
485 commercial artificial diets. The pH, dissolved oxygen (DO), and eels' activity level were
486 recorded for each tank every day. The eels' activity level was evaluated based on the sum of
487 the scores of the following eight criteria: initial responses to feeders, the proportion of eels
488 responding to feeders, sharpness of movement, the proportion of eels eating the artificial diet,
489 the level of splashes, the amount of scattered diet, the time to consume the diet, and the
490 proportion of foraging eels at the end of feeding. For each of the criteria, scoring was done on
491 a five-point scale (maximum point = 5) by an expert of eel aquaculture maintenance: thus, 40
492 (5 × 8 criteria) is the maximum point of the eels' activity score. Albeit subjective, the criteria
493 evaluated continuously by a professional provide inferences of eel's health conditions
494 throughout the time series. The water in the tanks were continuously mixed by the movement
495 of eels.

496 From each aquaculture tank (Tank 1–5), ca. 1.5 mL of water was sampled in the
497 morning every day during the 128 days from March 25 to July 30, 2020, except for 9 days
498 (Days 102, 103, 120, 121, 122, 123, 124, 125, and 126): i.e., the samples of 119 days were
499 available. In Tank 4, samples were unavailable on additional three days (Day 67–69) due to
500 the cleaning and the entire replacement of water. Consequently, the number of collected
501 samples were 592 (119 days × 5 tanks – 3 days in Tank 4). Water sample was collected in a
502 2.0 mL microtube and they were immediately stored at -20 °C in a freezer until DNA
503 extraction.

504

505 **Quantitative 16S rRNA sequencing.** To extract DNA from each sample, 250 µL of the
506 collected water was mixed with mixed with 400 µL lysis buffer (0.0025 % SDS, 20 mM Tris
507 (pH 8.0), 2.5 mM EDTA, and 0.4 M NaCl) and 250 µL 0.5 mm zirconium beads in a 2.0 mL
508 microtube. The microtubes were then shaken at 25 Hz for 5 min using TissueLyser II
509 (Qiagen, Venlo). After centrifugation, the aliquot was mized with proteinase K solution

510 ($\times 1/100$ of the total volume), being incubated at 40 °C for 60 min followed by 95 °C for 5
511 min.

512 We then performed PCR by applying a quantitative amplicon sequencing method^{22,51}.
513 Although most existing microbiome studies were designed to infer “relative” abundance of
514 microbial amplicon sequence variants (ASVs) or operational taxonomic units (OTUs),
515 information of “absolute” abundance provide additional insights into microbiome dynamics:
516 i.e., insights into increase/decrease of the population size of each prokaryote ASV/OTU
517 within a microbiome throughout a time-series²². The quantitative amplicon sequencing
518 approach is based on the addition of artificial (standard) DNA sequences with defined
519 concentrations into PCR master solutions. Therefore, even if compositions or concentrations
520 of PCR inhibitor molecules in DNA extracts vary among time-series samples, potential bias
521 caused by such inhibitors can be corrected based on the use of the internal standards (i.e.,
522 standard DNAs within PCR master solutions).

523 Prokaryote 16S rRNA region was PCR-amplified with the forward primer 515f⁵² fused
524 with 3–6-mer Ns for improved Illumina sequencing quality and the forward Illumina
525 sequencing primer (5'- TCG TCG GCA GCG TCA GAT GTG TAT AAG AGA CAG- [3–6-
526 mer Ns] – [515f] -3') and the reverse primer 806rB⁵³ fused with 3–6-mer Ns for improved
527 Illumina sequencing quality⁵⁴ and the reverse sequencing primer (5'- GTC TCG TGG GCT
528 CGG AGA TGT GTA TAA GAG ACA G [3–6-mer Ns] - [806rB] -3') (0.2 μ M each). To
529 apply the quantitative amplicon sequencing, five standard DNA sequence variants with
530 different concentrations of artificial 16S rRNA sequences (0.1, 0.05, 0.02, 0.01, and 0.005
531 nM) were added to PCR master mix solutions²². The buffer and polymerase system of KOD
532 One (Toyobo) was used with the temperature profile of 35 cycles at 98 °C for 10 s, 55 °C for
533 30 s, 68 °C for 30 s. To prevent generation of chimeric sequences, the ramp rate through the
534 thermal cycles was set to 1 °C/sec⁵⁵. Illumina sequencing adaptors were then added to
535 respective samples in the supplemental PCR using the forward fusion primers consisting of
536 the P5 Illumina adaptor, 8-mer indexes for sample identification⁵⁶ and a partial sequence of
537 the sequencing primer (5'- AAT GAT ACG GCG ACC ACC GAG ATC TAC AC - [8-mer
538 index] - TCG TCG GCA GCG TC -3') and the reverse fusion primers consisting of the P7
539 adaptor, 8-mer indexes, and a partial sequence of the sequencing primer (5'- CAA GCA GAA
540 GAC GGC ATA CGA GAT - [8-mer index] - GTC TCG TGG GCT CGG -3'). KOD One
541 was used with a temperature profile: followed by 8 cycles at 98 °C for 10 s, 55 °C for 30 s, 68
542 °C for 30 s (ramp rate = 1 °C/s). The PCR amplicons of the samples were then pooled after a
543 purification/equalization process with the AMPureXP Kit (Beckman Coulter). Primer dimers,
544 which were shorter than 200 bp, were removed from the pooled library by supplemental

545 purification with AMPureXP: the ratio of AMPureXP reagent to the pooled library was set to
546 0.6 (v/v) in this process. Because the quality of forward sequences is generally higher than
547 that of reverse sequences in Illumina sequencing, we optimized the MiSeq run setting in order
548 to use only forward sequences. Specifically, the run length was set 271 forward (R1) and 31
549 reverse (R4) cycles to enhance forward sequencing data: the reverse sequences were used
550 only for screening 16S rRNA sequences in the following bioinformatic pipeline.

551

552 **Bioinformatics.** In total, 16,298,203 sequencing reads were obtained in the Illumina
553 sequencing. The raw sequencing data were converted into FASTQ files using the program
554 bcl2fastq 1.8.4 distributed by Illumina. The raw sequencing data were converted into FASTQ
555 files using the program bcl2fastq 1.8.4 distributed by Illumina. The output FASTQ files were
556 demultiplexed using Claident v0.2. 2018.05.29⁵⁷. The removal of low-quality sequences and
557 ASV inferences were done using DADA2⁵⁸ v.1.22.0 of R 4.1.2⁵⁹(R Core Team, 2020). The
558 taxonomy of the output ASVs was inferred based on the naive Bayesian classifier method⁶⁰
559 using the SILVA v.138 database⁶¹. Based on the calibration with the concentration gradients
560 of the five standard DNAs, concentrations of respective ASVs were obtained for each sample
561 (16S rRNA copy numbers per unit volume of tank water samples; copies/µL). As the number
562 of 16S rRNA copies per genome generally varies among prokaryotic taxa⁶², 16S rRNA copy
563 concentration is not directly the optimal proxy of cell or biomass concentration. Meanwhile,
564 in this study, estimates of 16S rRNA copy concentrations were used to observe
565 increase/decrease of abundance (i.e., population dynamics) *within* the time-series of
566 respective microbial ASVs. Thus, variation in the number 16S rRNA copy numbers among
567 microbial taxa had no qualitative effects on the subsequent population- and community-
568 ecological analyses. The samples in which Pearson's coefficients of correlations between
569 sequencing read numbers and standard DNA copy numbers (i.e., correlation coefficients
570 representing calibration curves) were less than 0.8 were removed as those with unreliable
571 estimates. Samples with less than 1,000 reads were discarded as well. In total, microbiome
572 data were successfully obtained from 577 out of 592 samples. For each aquaculture tank, we
573 then obtained a sample × ASV matrix, in which a cell entry depicted the concentration of 16S
574 rRNA copies of an ASV in a sample.

575

576 **Community structure.** For each aquaculture tank, Bray-Curtis β -diversity was calculated for
577 all pairs of time points based on the matrix describing the relative abundance of prokaryote
578 families using the vegan 2.6.2 package⁶³ of R. Based on the β -diversity estimates, the

579 community structure of all the samples across the five water tanks were visualized on the
580 surface of non-metric multidimensional scaling (NMDS). The vectors representing the
581 environmental variables (pH and DO) and eels' activity level were calculated with the
582 "envfit" function of R and they were shown on the NMDS surface. The analysis was
583 conducted as well based on the matrix describing the relative abundance of genera.

584

585 **Environmental preference of ASVs.** To evaluate environmental preference of each
586 microbial ASV, Spearman's correlation between absolute abundance (in the metric of DNA
587 copy numbers of 16S rRNA) and pH was calculated. For each tank, the ASVs that appeared
588 in 30 or more samples were subjected to the analysis. For each ASV in each water tank, the
589 statistical significance of the obtained correlation coefficient was examined with a
590 randomization analysis obtained based on a twin-surrogate method for time-series data²⁷
591 (100,000 permutations). Correlation coefficients less than -0.3 and those larger than 0.3
592 tended to show statistically significant negative and positive correlations with pH,
593 respectively, after Benjamini-Hochberg adjustment of *P* values in multiple testing [i.e., false
594 discovery rate (FDR)]. Likewise, Pearson's correlation coefficients between respective ASVs'
595 absolute abundance and DO concentrations were calculated.

596

597 **ASV abundance and eel's activity.** We explored microbial ASVs that potentially have
598 profound impacts on eels' health. For each water tank, Spearman's correlation between
599 absolute abundance and eels' activity score was calculated for the ASVs that appeared in 30
600 or more samples. However, because ASV abundance could be affected by pH or dissolved
601 oxygen concentration, the use of such simple correlation coefficients might be misleading.
602 Therefore, we controlled potential effects by environmental factors/conditions based on a
603 partial correlation approach as follows:

604
$$r_{xy \cdot z} = \frac{r_{xy} - r_{xz}r_{yz}}{\sqrt{1-r_{xz}^2} \sqrt{1-r_{yz}^2}},$$

605 where r_{xy} , r_{xz} , and r_{yz} were correlation between ASV abundance and eels' activity level,
606 that between ASV abundance and an environmental factor (pH or dissolved oxygen
607 concentration), and that between eels' activity level and an environmental factor, respectively.
608 For each ASV, a randomization analysis was performed with the twin-surrogate method
609 (100,000 permutations).

610

611 **Time-lag analysis.** We extended the analysis of partial correlation between microbial
612 abundance and eels' activity level by introducing time lags between the two variables.
613 Specifically, partial correlation between an ASV's abundance on Day x and eels' activity
614 score on Day $x + l$ was calculated. The time lag l ranged from -5 to 5 in the analysis ($l = 0$
615 means no delay introduced to eels' activity level).

616

617 **Pathogenic and non-pathogenic *Edwardsiella*.** We performed an additional analysis to infer
618 the proportion of pathogenic and non-pathogenic clades^{29,30} of *Edwardsiella* bacteria in the
619 aquaculture system. In a previous phylogenetic study based on an internal fragment of iron-
620 cofactored superoxide dismutase gene (*sodB*), *Edwardsiella* species and strains have been
621 classified into two major clades, which differ in the presence of pathogenicity to fish
622 (hereafter, "pathogenic" and "non-pathogenic" clades). Therefore, we characterized
623 *Edwardsiella* bacteria in the aquaculture tanks based on the illumina sequencing of the
624 *Edwardsiella sodB* gene sequences. The fragment of the *sodB* region was PCR-amplified with
625 the forward primer E1F²⁹ fused with 3–6-mer Ns for improved Illumina sequencing quality
626 and the forward Illumina sequencing primer (5' - TCG TCG GCA GCG TCA GAT GTG TAT
627 AAG AGA CAG- [3–6-mer Ns] – [E1F] -3') and the reverse primer 497R²⁹ fused with 3–6-
628 mer Ns for improved Illumina sequencing quality⁵⁴ and the reverse sequencing primer (5'-
629 GTC TCG TGG GCT CGG AGA TGT GTA TAA GAG ACA G [3–6-mer Ns] - [497R] -3')
630 (0.2 μ M each). The buffer and polymerase system of KOD One (Toyobo) was used with the
631 temperature profile of 35 cycles at 98 °C for 10 s, 55 °C for 5 s, 68 °C for 30 s (ramp rate = 1
632 °C/sec). The sequencing adaptors and sample identifier indexed were added to the amplicons,
633 and the purification of the library and sequencing was performed as detailed above.

634 The output sequencing data were demultiplexed and processed with DADA2. The
635 ASVs that were not aligned to the *sodB* sequences of *Edwardsiella*²⁹ were discarded. The
636 neighbor-joining tree of the remaining ASVs and previously reported *Edwardsiella* sequences
637 was reconstructed based on the maximum composite likelihood method with a bootstrap test
638 (1,000 permutations). The ASVs belonging to the pathogenic clade and those belonging to the
639 non-pathogenic clade of *Edwardsiella* were distinguished within the phylogeny.

640

641 **Microbe–microbe interactions.** Potential positive/negative interactions between microbial
642 ASVs were inferred based on the framework of sparse inverse covariance estimation for

643 ecological associations (SPIEC-EASI⁶⁴). For each water tank, patterns in the coexistence (co-
644 occurrence) were examined with the Meinshausen-Bühlmann (MB) method as implemented
645 in the SpiecEasi package⁶⁴ of R. The network inference based on coexistence patterns allowed
646 us to detect pairs of microbial ASVs that potentially interact with each other in facilitative
647 ways and/or those potentially sharing environmental preference. Because estimation of
648 coexistence patterns was not feasible for rare nodes, the microbial ASVs that appeared in less
649 than 30 samples were excluded from the input matrices of the network analysis. Network
650 modules, within which closely associated ASVs were interlinked with each other, were
651 identified with the algorithm based on edge betweenness using the igraph package⁶⁵ of R. For
652 each network module in each water tank, mean partial correlation with eels' activity level was
653 calculated across ASVs constituting the module.

654 In addition to the networks representing whole coexistence patterns, we reconstructed
655 networks depicting direct interactions between microbial ASVs. To separate effects of direct
656 microbe–microbe interactions from those of shared environmental preferences between
657 microbes (i.e., shared niches), 10 latent components (latent variables) were included in the
658 analysis based on the “sparse and low-rank” (SLR) model⁶⁶.

659

660 **KEGG pathway/process profiles.** To infer metabolic interactions between microbial ASVs,
661 we performed a series of analysis based on reference genome information. We performed
662 phylogenetic prediction of gene repertoires using PICRUSt2 v2.3.0-b⁶⁷ in order to gain the
663 overview of the niche space defined with metagenomic information^{68,69}. ASVs that appeared
664 in 30 or more sample across the five tanks were subjected to the analysis. Based on the
665 inferred KEGG metabolic pathway/process profiles⁷⁰, microbial ASVs were plotted on a two-
666 dimensional surface of a principal coordinate analysis (PCoA) based on Bray-Curtis β -
667 diversity of KEGG metabolic pathway/process profiles.

668

669 **Metabolic modeling.** To infer potential metabolic interactions between microbes, we
670 performed the species metabolic interaction analysis⁷¹. For the ASVs that appeared in 30 or
671 more samples (day) in at least one aquaculture tank, we explored NCBI RefSeq genome
672 sequences whose 16S rRNA sequences matched those of query ASVs with $\geq 99\%$ identity. In
673 the database exploration, reference genome information was available for 181 out of 417
674 ASVs examined. The reference genome information was subjected to genome-scale metabolic
675 modeling as implemented in CarveMe⁷² 1.5.0. Metabolic resource overlap (MRO) and

676 metabolic interaction potential (MIP) were then estimated for each pair of microbial ASVs as
677 implemented in SMETANA⁷¹ 1.0.0.

678

679 **Data availability**

680 The 16S rRNA sequencing data are available from the DNA Data Bank of Japan (DDBJ) with
681 the accession number PRJDB14313. The microbial community data are deposited at the
682 GitHub repository (<https://github.com/hiro-toju/EelMicrobiome128>).

683

684 **Code availability**

685 All the R scripts used to analyze the data are available at the GitHub repository
686 (<https://github.com/hiro-toju/EelMicrobiome128>).

687

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692

693 **Author Contributions.** H.T. designed the work with D.Y.. D.Y., I.H., G.S., and H.F.
694 performed experiments. D.Y., H.F., K.S., and H.T. analyzed the data. H.T. and D.Y. wrote the
695 paper with all the authors.

696

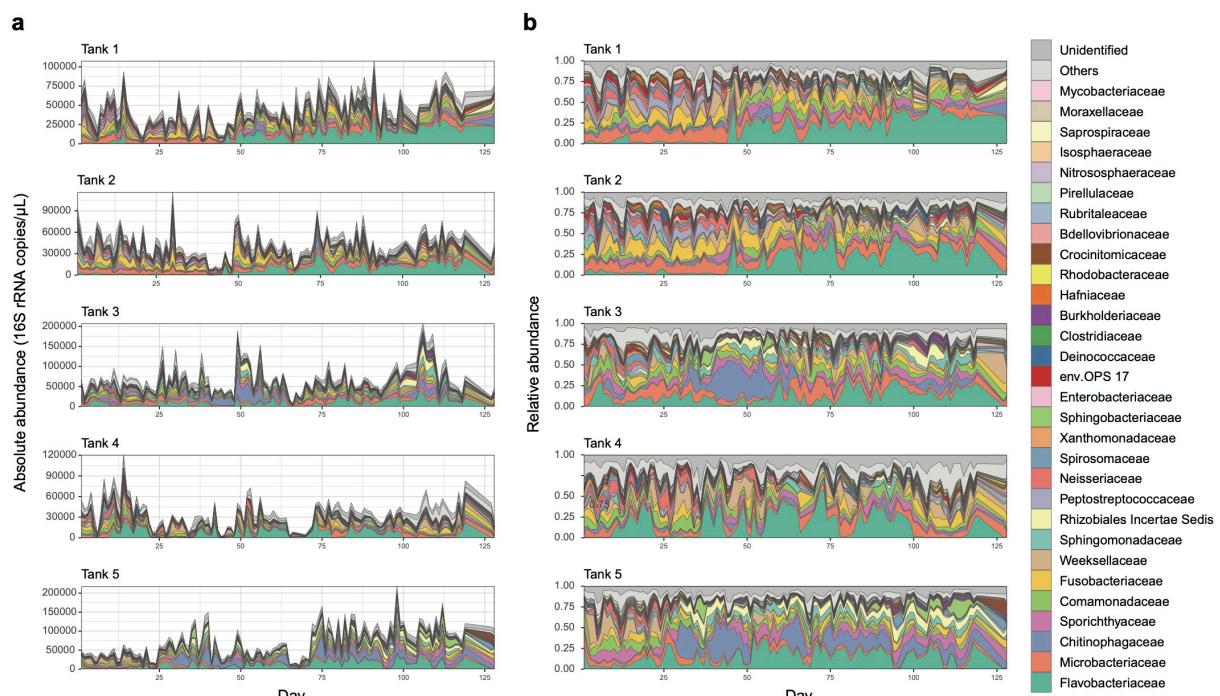
697 **Competing Interests.** The authors declare no competing interests.

698

699 **Correspondence and requests for materials** should be addressed to
700 toju.hirokazu.4c@kyoto-u.ac.jp.

701

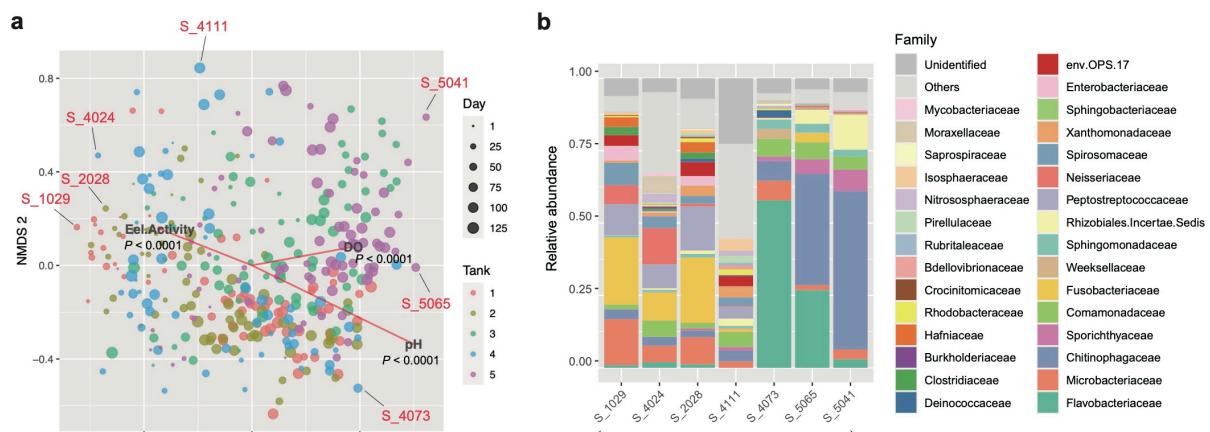
702



703

704 **Fig. 1 | Microbiome dynamics in the eel aquaculture system. a**, Dynamics of absolute
705 abundance. For each water sample of each aquaculture tank, absolute abundance of
706 prokaryotes was inferred as 16S rRNA gene copy concentration based on the quantitative
707 amplicon sequencing approach with standard DNA gradients. **b**, Dynamics of relative
708 abundance. The time-series of the family-level taxonomic compositions are shown for each
709 aquaculture tank. See Extended Data Figures 1–3 for phylum-, order-, and genus-level
710 taxonomic compositions.

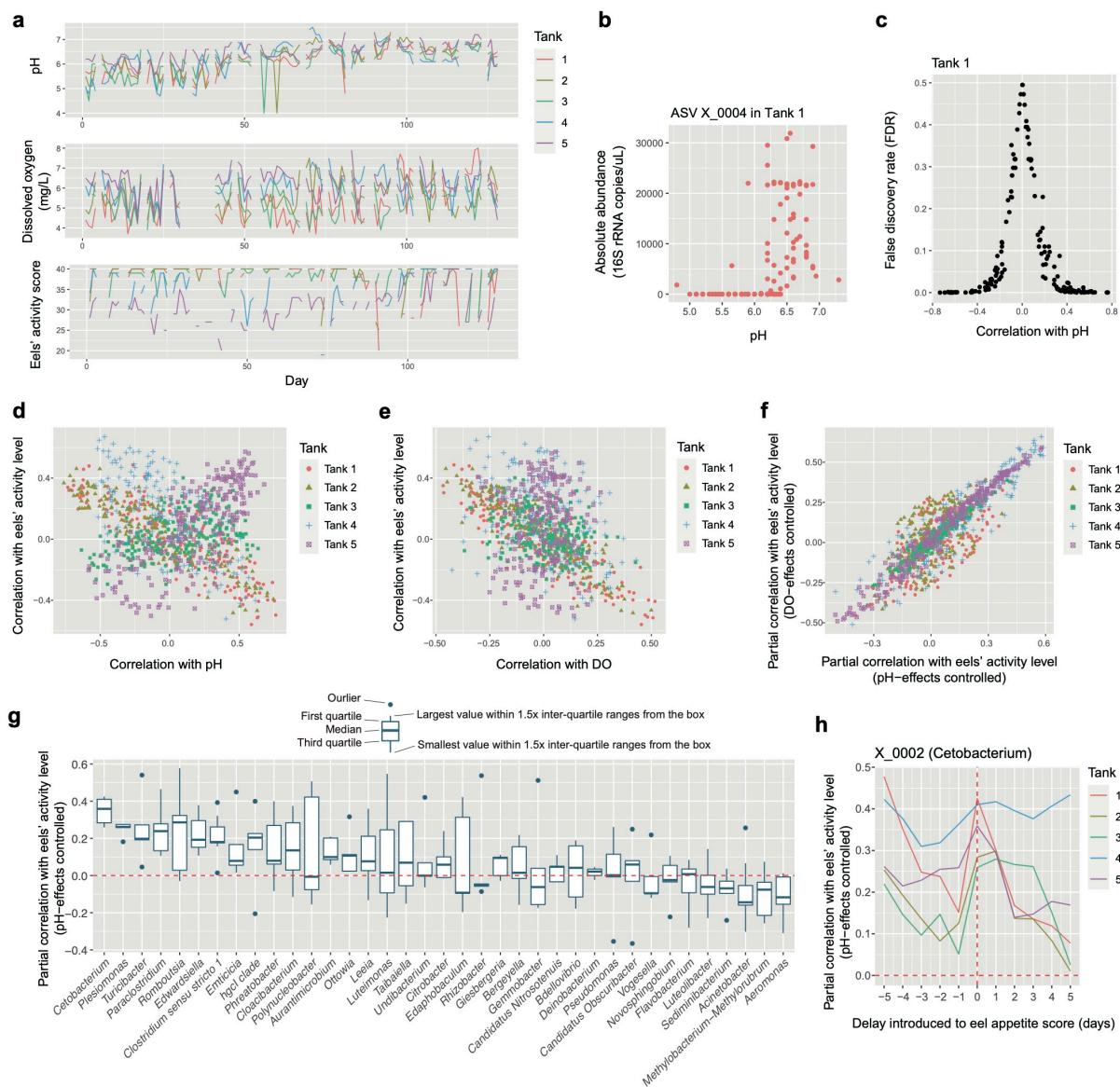
711



712

713 **Fig. 2 | Multivariate analysis of community structure.** **a**, Community state space.
714 Community compositions of the samples are plotted on the two-dimensional surface defined
715 with non-metric multidimensional scaling (NMDS). The NMDS was performed based on the
716 Bray-Curtis β -diversity of family-level taxonomic compositions. The projections of the data
717 points onto the vectors have maximum correlation with the variables examined (pH, DO, and
718 eels' activity level). See Extended Data Figure 4 for an additional analysis based on genus-
719 level taxonomic compositions. **b**, Examples of community structure in the NMDS surface.
720 For several points within the NMDS surface (panel **a**), family-level taxonomic compositions
721 are shown. The example points are ordered along the vector representing high eels' activity
722 level.

723

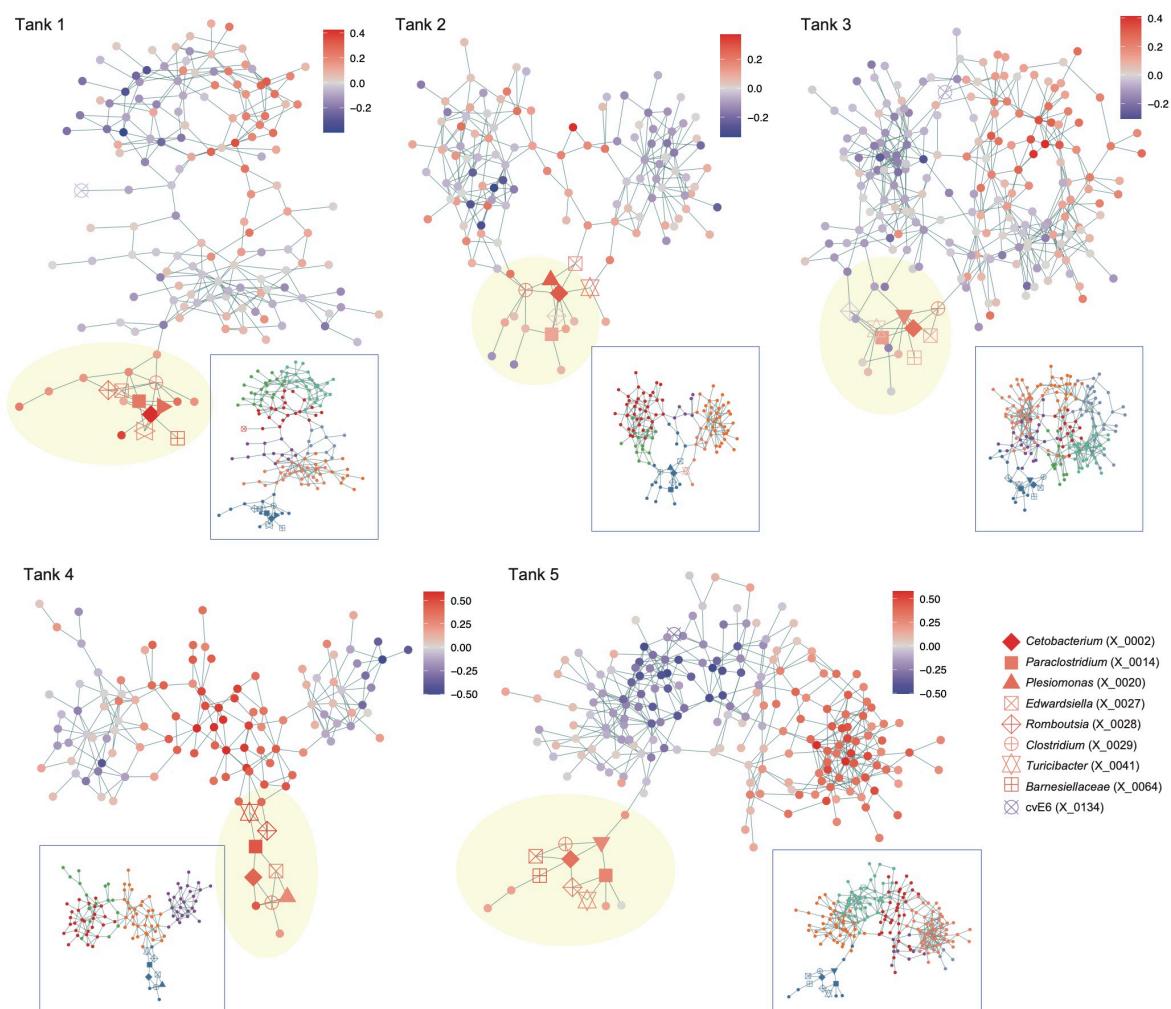


724

725 **Fig. 3 | Microbes associated with eels' activity.** **a**, Timeseries of pH, dissolved oxygen (DO)
 726 level, and eels' activity score are shown for each aquaculture tank. **b**, Example of the
 727 correlation analysis. For each variable shown in the panel **b**, Spearman's correlation with the
 728 absolute abundance of each ASV in each aquaculture tank was examined. **c**, Randomization
 729 analysis of correlation. Significance of correlation coefficients was examined based on a twin-
 730 surrogate randomization analysis of time-series data (100,000 permutations). Coefficients less
 731 than -0.3 and those larger than 0.3 roughly represent significant negative and positive
 732 correlations, respectively. **d**, Each ASV's correlation with pH and eels' activity level. **e**, Each
 733 ASV's correlation with DO and eels' activity level. **f**, Partial correlation with eels' activity
 734 level. To control the effects of pH or DO, partial correlation between absolute abundance and
 735 eels' activity scores was calculated for each ASV in each tank. **g**, Taxonomic comparison of
 736 relationship with eels' activity level. Partial correlation with eels' activity level is shown for

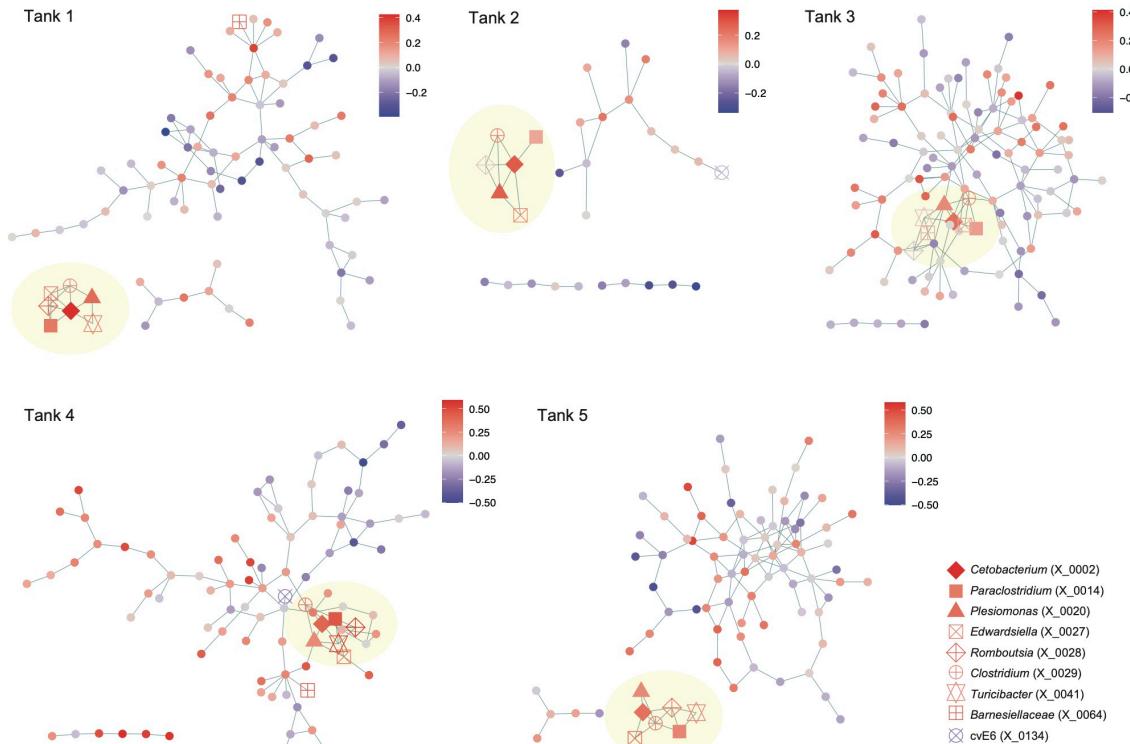
737 the genera that appeared in all the aquaculture tanks (shown in the decreasing order of mean
738 values). **h**, Time-lag analysis of correlations. In calculating partial correlation between eels'
739 activity level and the absolute abundance of the *Cetobacterium* ASV (X_0002), defined time-
740 lag was introduced to the eels' activity variable.

741



742

743 **Fig. 4 | Microbe-to-microbe coexistence networks.** For each aquaculture tank, patterns of
744 coexistence were analyzed based on the sparse inverse covariance estimation for ecological
745 associations with the Meinshausen-Bühlmann (MB) model. Only the ASVs that appeared in
746 30 or more samples were targeted in the analysis of each tank. Within the networks, pairs of
747 microbial ASVs that may interact with each other in facilitative ways and/or those potentially
748 sharing environmental preference are linked with each other. Network modules, which
749 represent groups of densely linked ASVs, are shown for each network. The color of nodes
750 indicates partial correlation between ASV abundance and eels' activity level (controlled
751 variable = pH). The inferred network modules are shown by colors for each tank in a box. The
752 ASVs that consistently displayed positive or negative correlation with eels' activity level
753 (Extended Data Fig. 5) are highlighted with the defined symbols. See Extended Data Figures
754 6–8 for additional information of the nodes (ASVs) and modules within the network. ASVs
755 included in minor sub-networks (number of nodes < 5) are not shown.
756

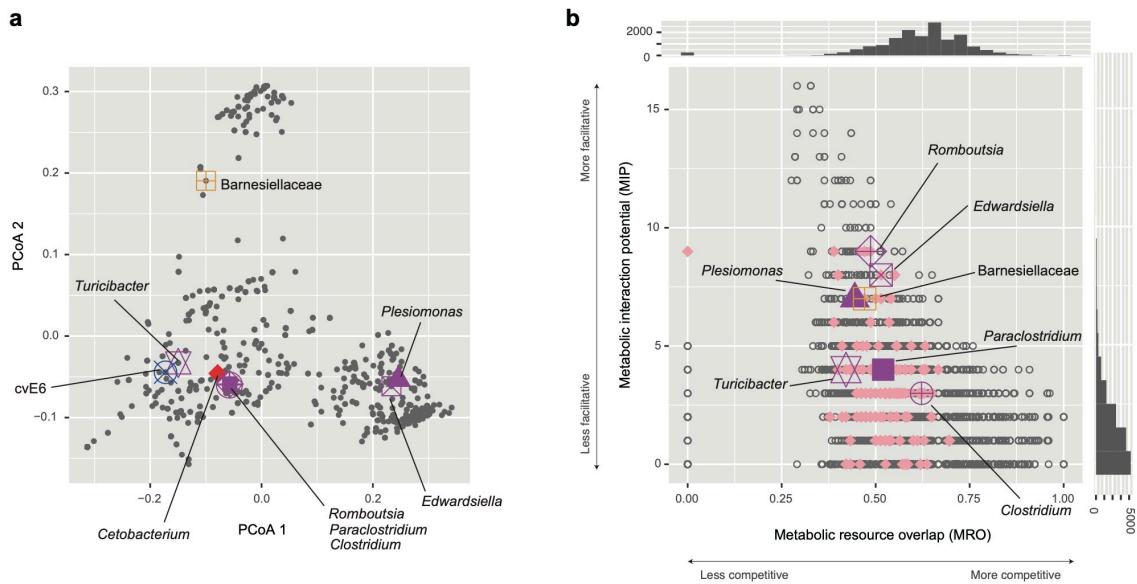


757

758 **Fig. 5 | Inferred direct interactions between microbes.** Based on the “sparse and low-rank”
759 (SLR) model, direct interactions between microbial ASVs were inferred by controlling the
760 effects of shared environmental preference. Only the ASVs that appeared in 30 or more
761 samples were targeted in the analysis of each tank. The links between nodes represent
762 potentially positive interactions between ASVs. The color of nodes indicates partial
763 correlation between ASV abundance and eels’ activity level (controlled variable = pH). ASVs
764 included in minor sub-networks (number of nodes < 5) are not shown.

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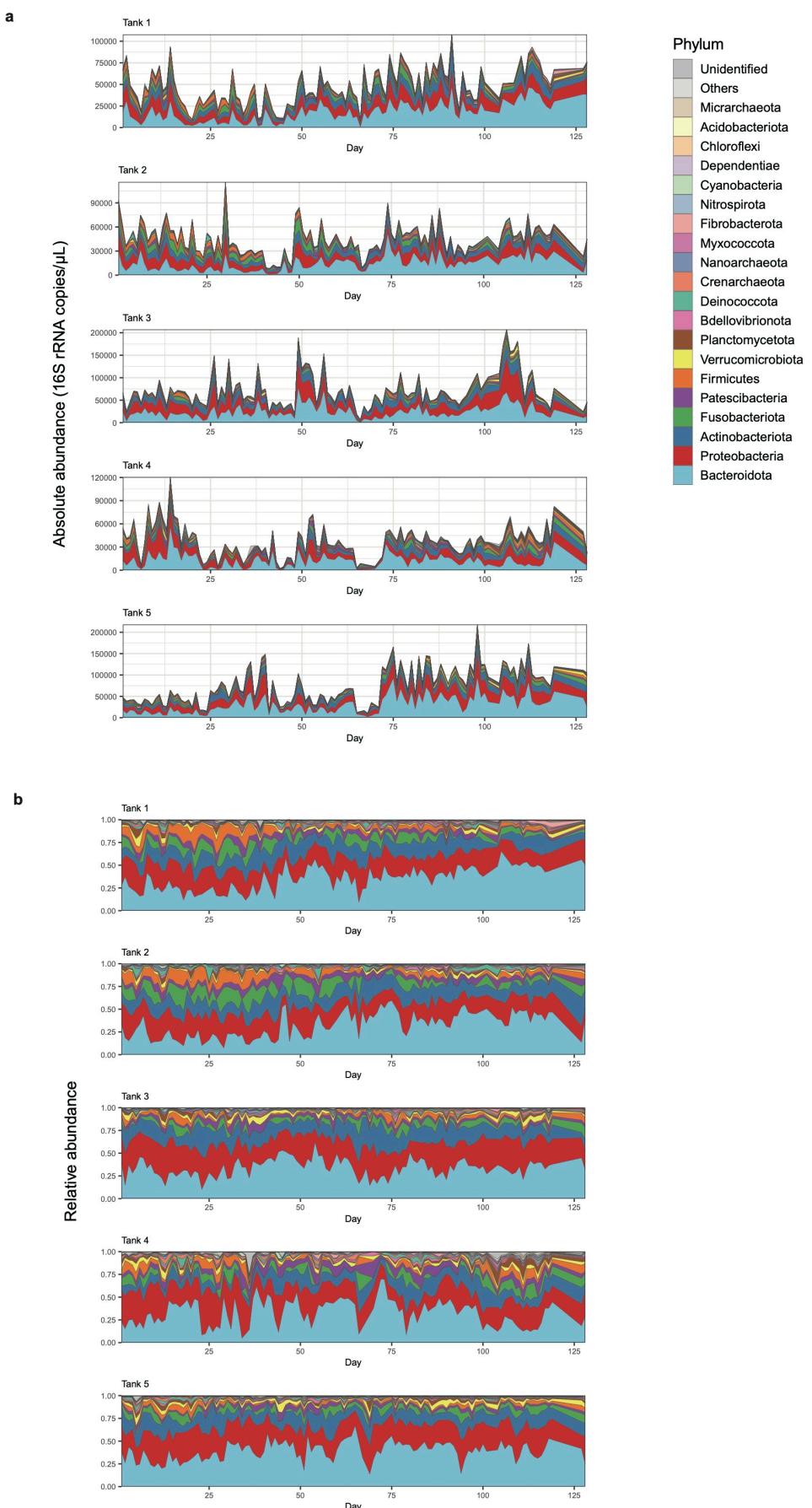
767

768 **Fig. 6 | Metabolic interactions between microbes.** **a**, Metagenomic niche space. Microbial
769 ASVs are plotted on a two-dimensional surface of PCoA based on their KEGG metabolic
770 pathway/process profiles inferred with a phylogenetic prediction of genomes. Microbial
771 ASVs plotted closely within the surface are expected to have similar gene repertoires. The
772 ASVs highlighted in Figures 4 and 5 are shown with large symbols. **b**, Potential competitive
773 and facilitative interactions. Based on the NCBI RefSeq genome information, potential
774 metabolic interactions between each pair of ASVs were inferred in terms of metabolic
775 resource overlap (MRO) and metabolic interaction potential (MIP). Histograms of MRO and
776 MIP are shown on the horizontal and vertical axes, respectively. ASV pairs including the
777 *Cetobacterium* ASV, whose abundance were positively associated with eels' activity level in
778 all the five water tanks (Fig. 3g; Extended Data Fig. 5), are shown in pink. Relationships
779 between the *Cetobacterium* ASV and the ASVs highlighted in Figures 4 and 5 are indicated
780 as well.

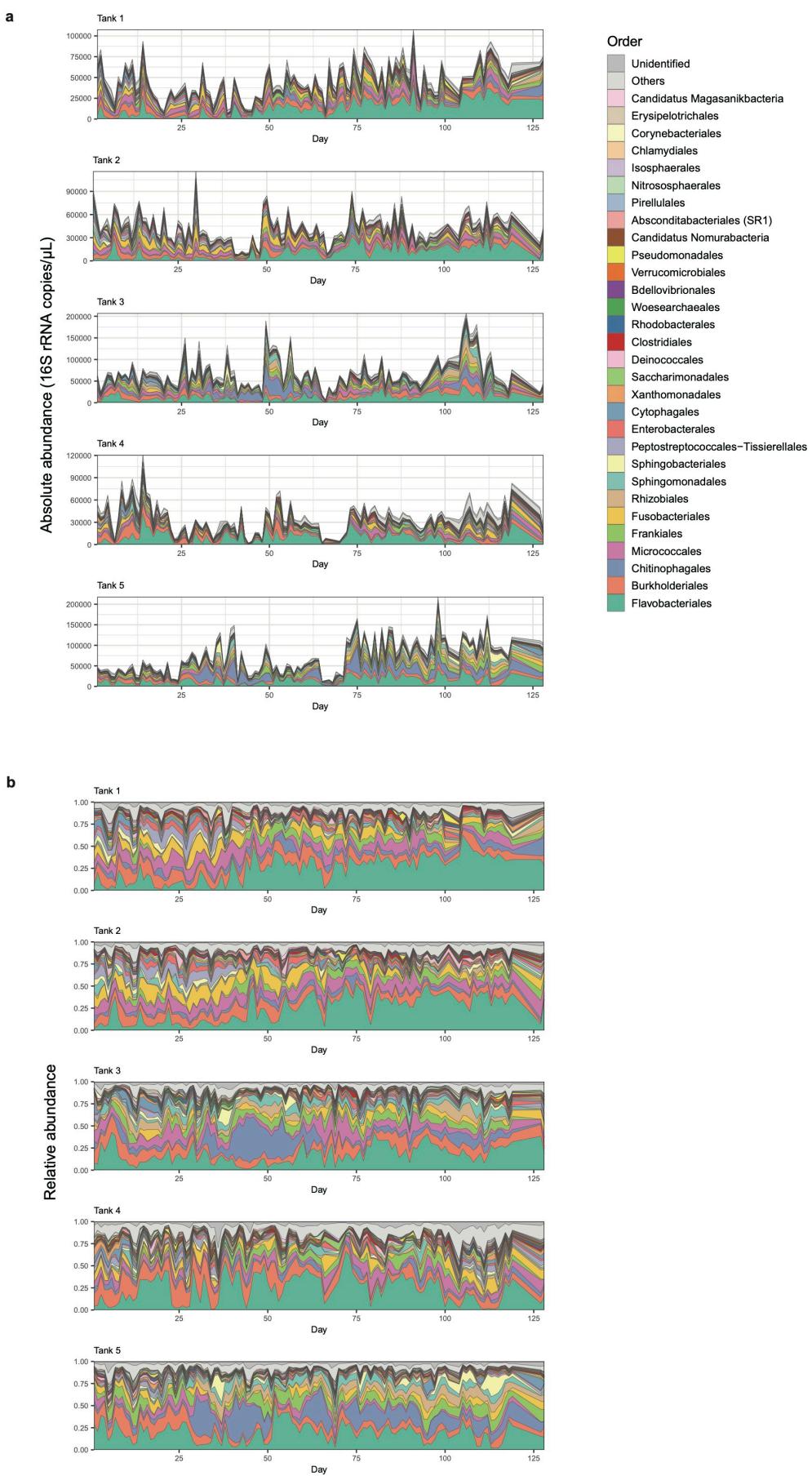
781

782 **Extended Data Figures**

783

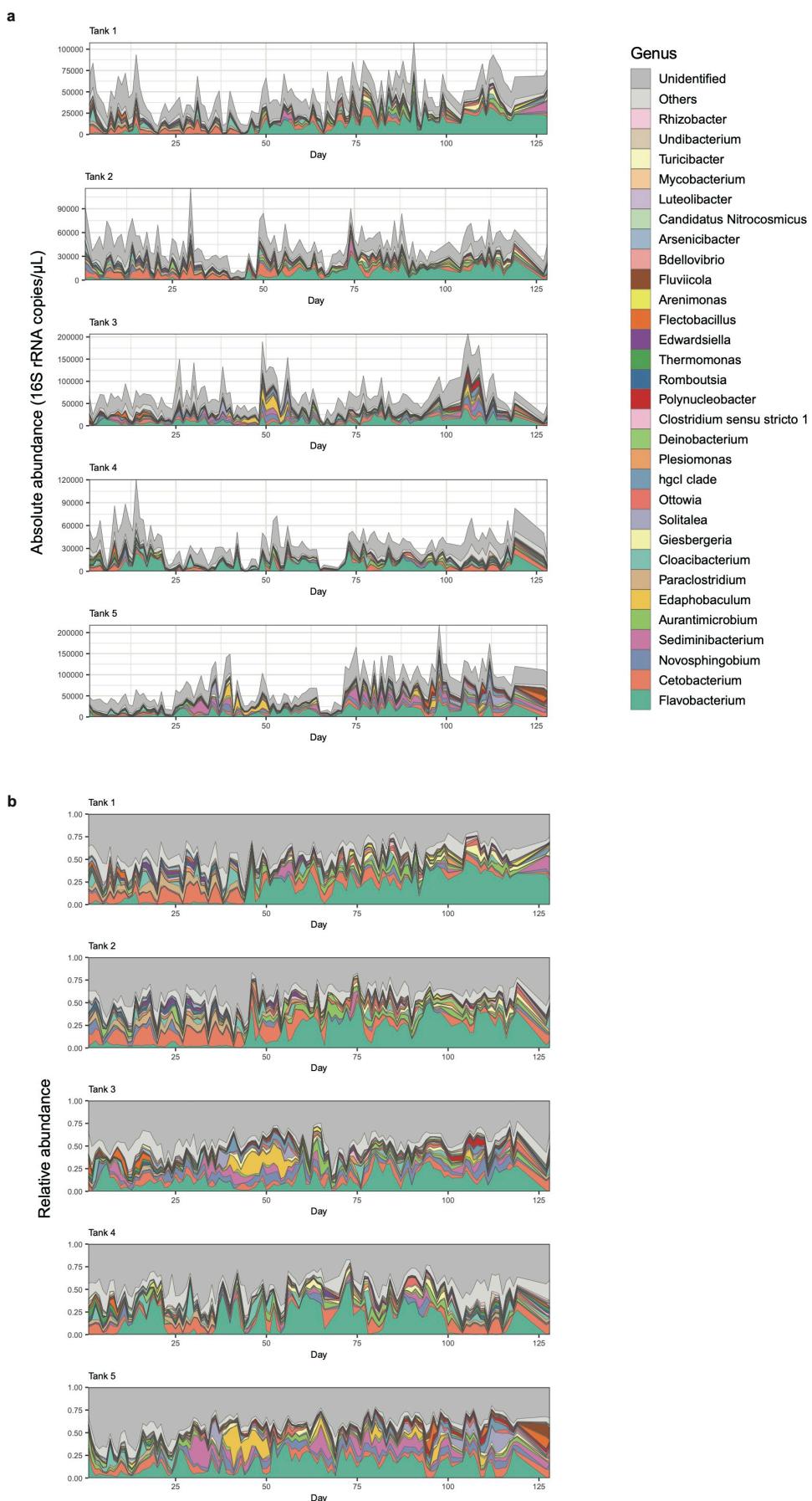


785 **Extended Data Fig. 1 | Phylum-level community structure.** **a**, Dynamics of absolute
786 abundance. For each water sample of each aquaculture tank, absolute abundance of
787 prokaryotes was inferred as 16S rRNA gene copy concentration based on the quantitative
788 amplicon sequencing approach with standard DNA gradients. **b**, Dynamics of relative
789 abundance. The time-series of the phylum-level taxonomic compositions are shown for each
790 aquaculture tank.
791



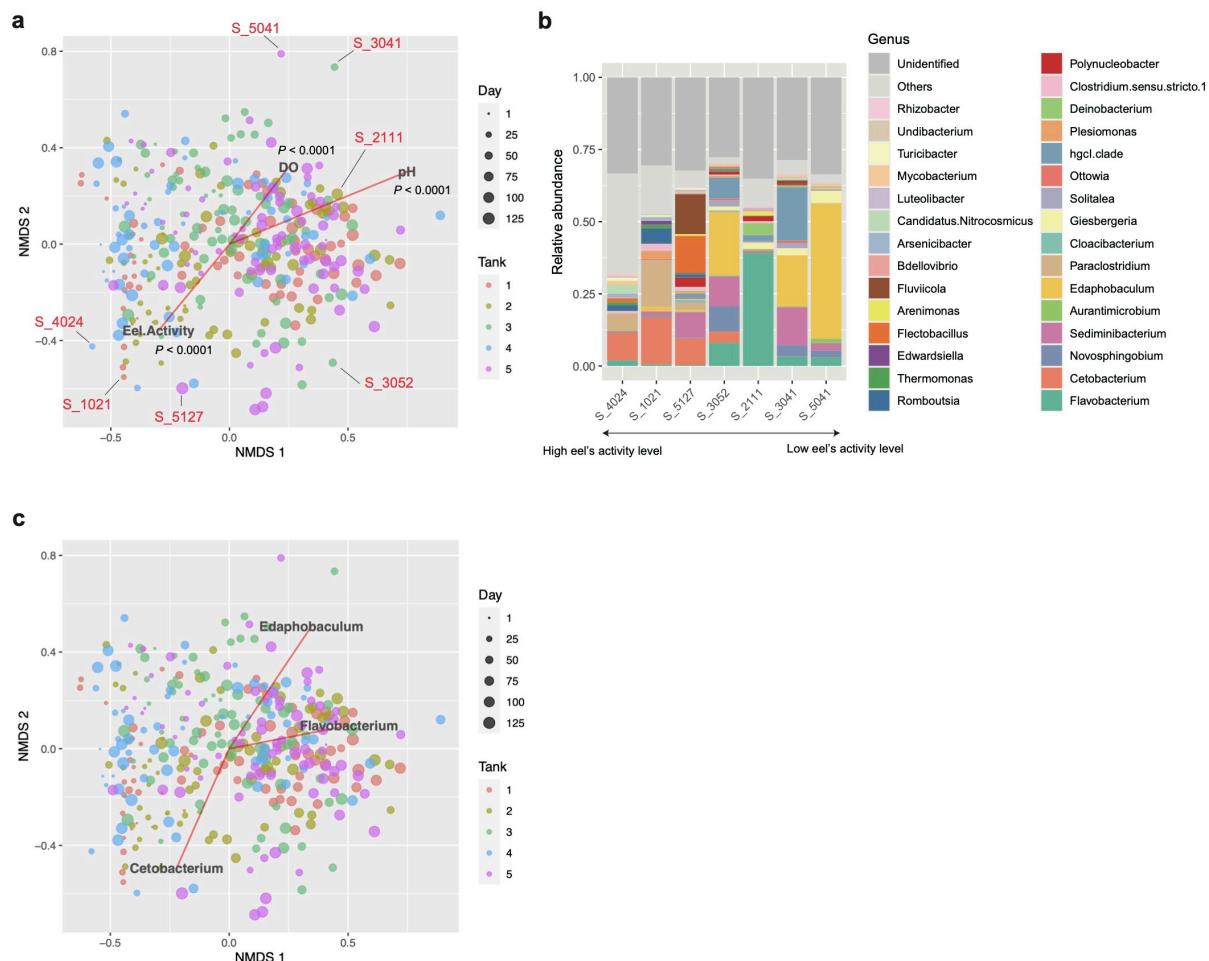
793 **Extended Data Fig. 2 | Order-level community structure.** **a**, Dynamics of absolute
794 abundance. For each water sample of each aquaculture tank, absolute abundance of
795 prokaryotes was inferred as 16S rRNA gene copy concentration based on the quantitative
796 amplicon sequencing approach with standard DNA gradients. **b**, Dynamics of relative
797 abundance. The time-series of the order-level taxonomic compositions are shown for each
798 aquaculture tank.

799



801 **Extended Data Fig. 3 | Genus-level community structure.** **a**, Dynamics of absolute
802 abundance. For each water sample of each aquaculture tank, absolute abundance of
803 prokaryotes was inferred as 16S rRNA gene copy concentration based on the quantitative
804 amplicon sequencing approach with standard DNA gradients. **b**, Dynamics of relative
805 abundance. The time-series of the genus-level taxonomic compositions are shown for each
806 aquaculture tank.

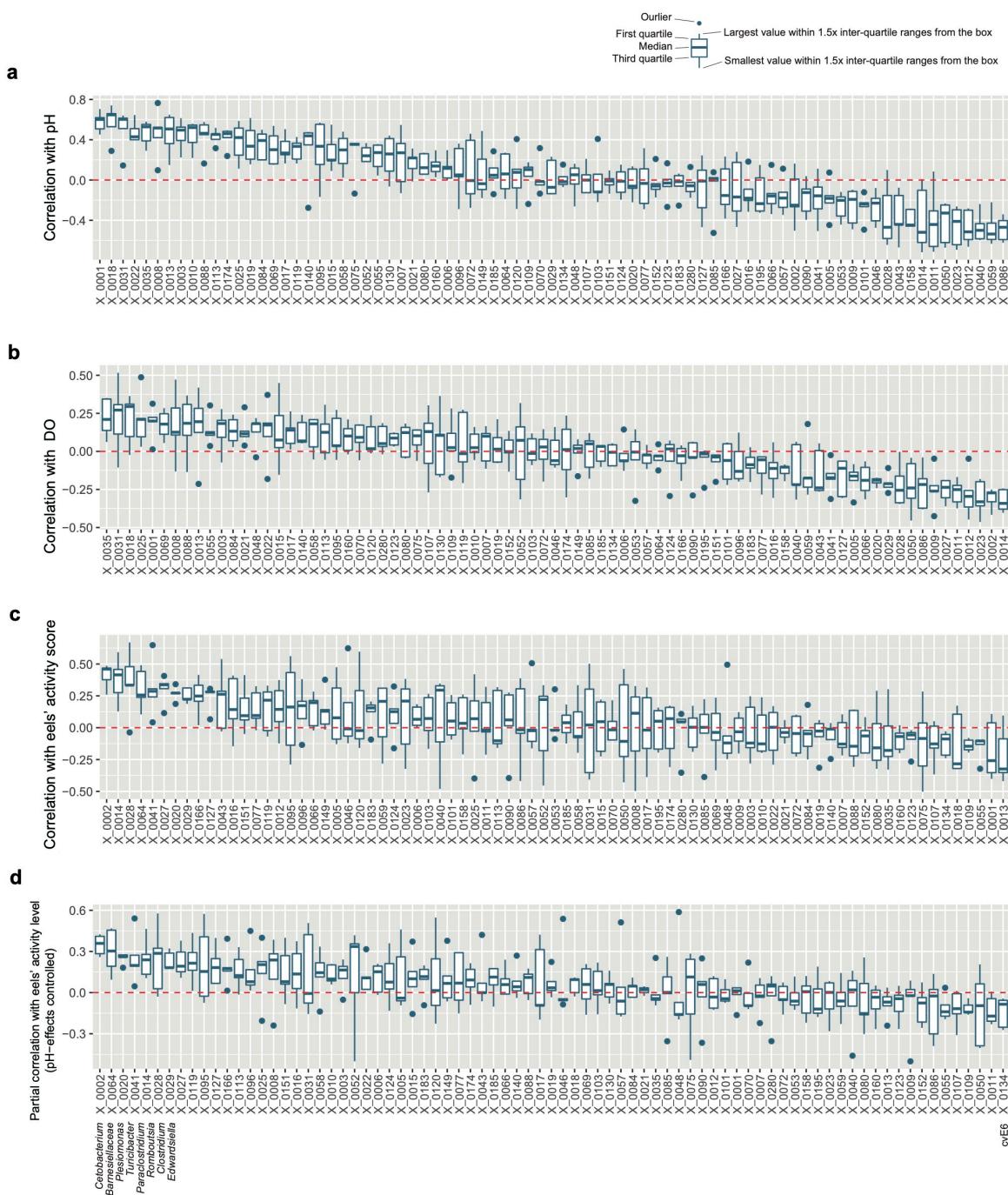
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809 **Extended Data Fig. 4 | Multivariate analysis of community structure (genus level).** **a**,
810 Community state space. Community compositions of the samples are plotted on the two-
811 dimensional surface defined with non-metric multidimensional scaling (NMDS). The NMDS
812 was performed based on the Bray-Curtis β -diversity of genus-level taxonomic compositions.
813 The projections of the data points onto the vectors have maximum correlation with the
814 variables examined (pH, DO, and eels' activity level). Examples of community structure in
815 the NMDS surface. For several points within the NMDS surface (panel **a**), genus-level
816 taxonomic compositions are shown. The example points are ordered along the vector
817 representing high eels' activity level. **c**, Indicator genera. The vectors representing the relative
818 abundance of *Cetobacterium*, *Flavobacterium*, and *Edaphobaculum*, which were highlighted
819 in the main text, are shown.

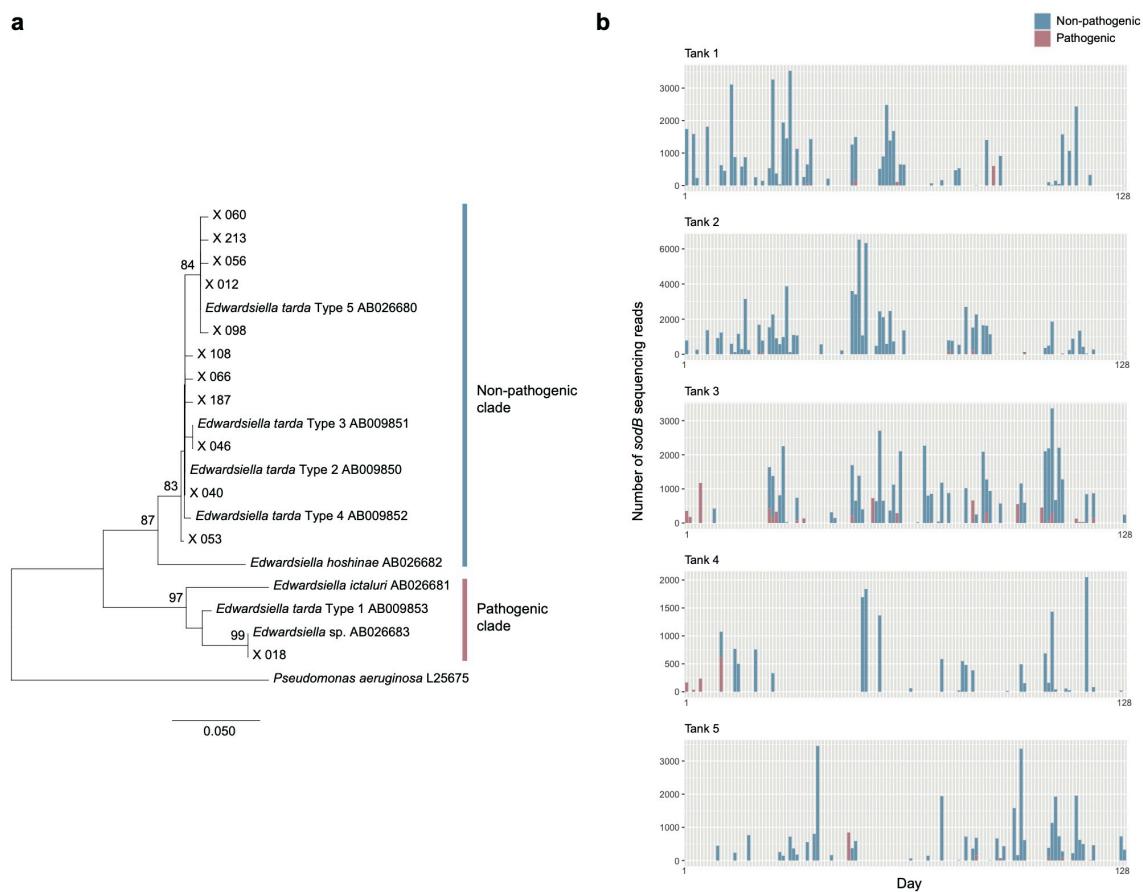
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821

822 **Extended Data Fig. 5 | ASV-level comparison of correlation with environmental**
 823 **variables and eels' activity level.** **a**, Correlation with pH. Correlation with eels' activity level
 824 is shown for the ASVs that appeared in all the aquaculture tanks (shown in the decreasing
 825 order of mean values). The boxes and bars represent variation across tanks. **b**, Correlation
 826 with DO. **c**, Correlation with eels' activity level. **d**, Partial correlation with eels' activity level
 827 (controlled variable = pH). Taxonomic information is shown for the ASVs discussed in the
 828 main text.

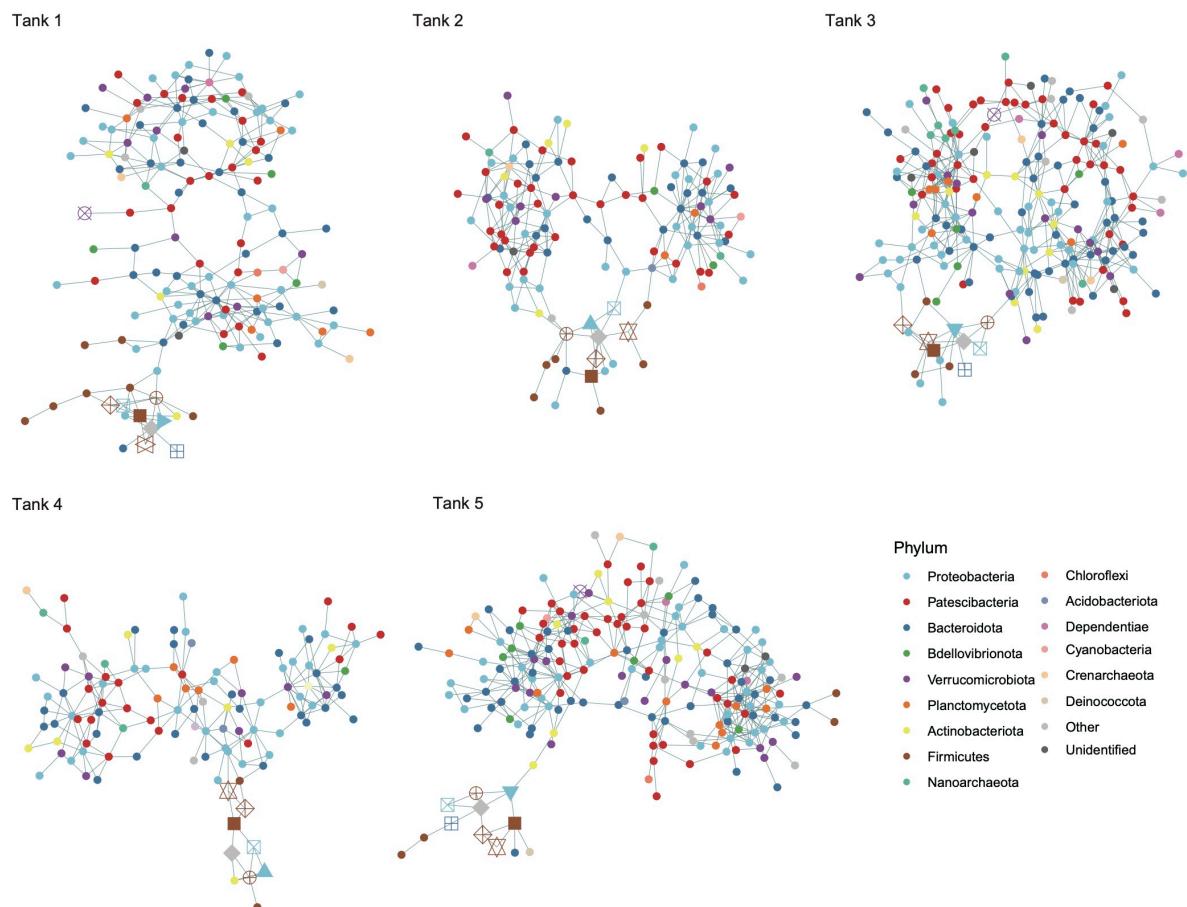
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830

831 **Extended Data Fig. 6 | Phylogenetic analysis of *Edwardsiella*.** **a**, Phylogeny of
832 *Edwardsiella*. In an additional amplicon sequencing of the *sodB* gene, the neighbor-joining
833 tree of the *Edwardsiella* bacteria was reconstructed with the maximum composite likelihood
834 method. Bootstrap values larger than 70 % are shown on the nodes (1,000 permutations). The
835 pathogenic and non-pathogenic clades identified in a previous study²⁹ are indicated. **b**, Time-
836 series of pathogenic and non-pathogenic *Edwardsiella*. The number of detected sequencing
837 reads of the *sodB* fragment is across the time-series of each aquaculture tank.

838

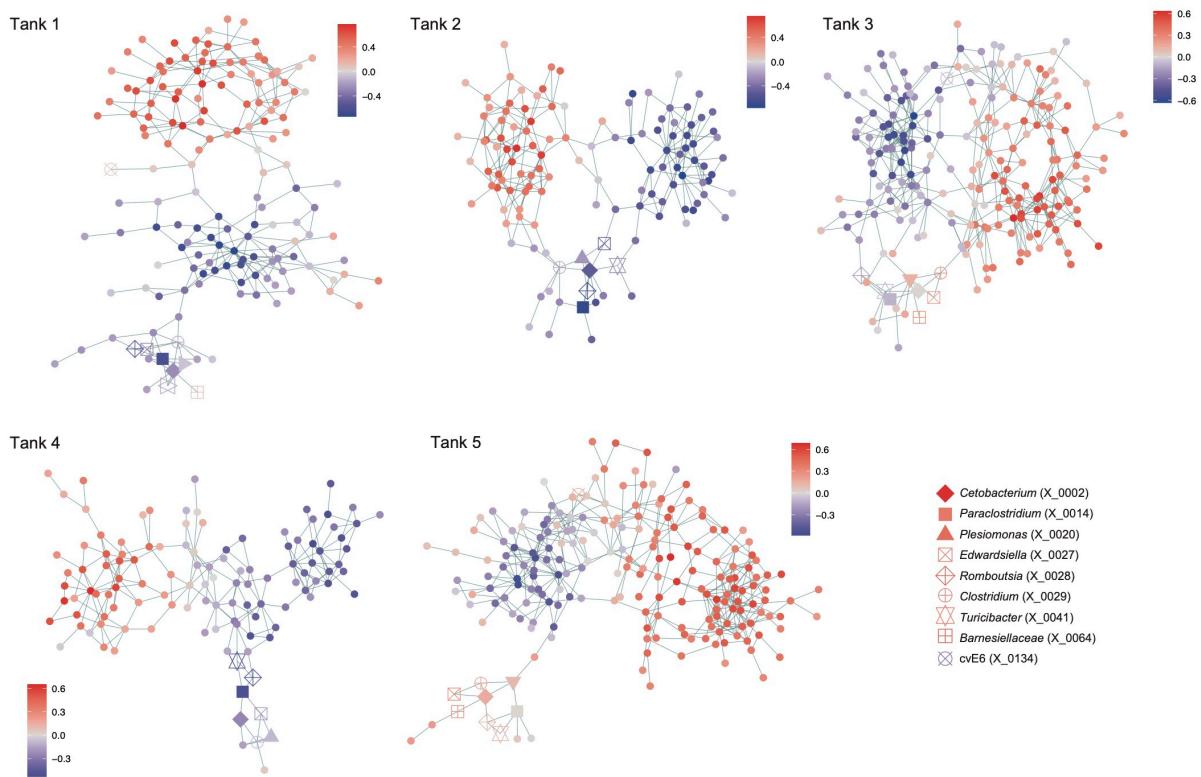


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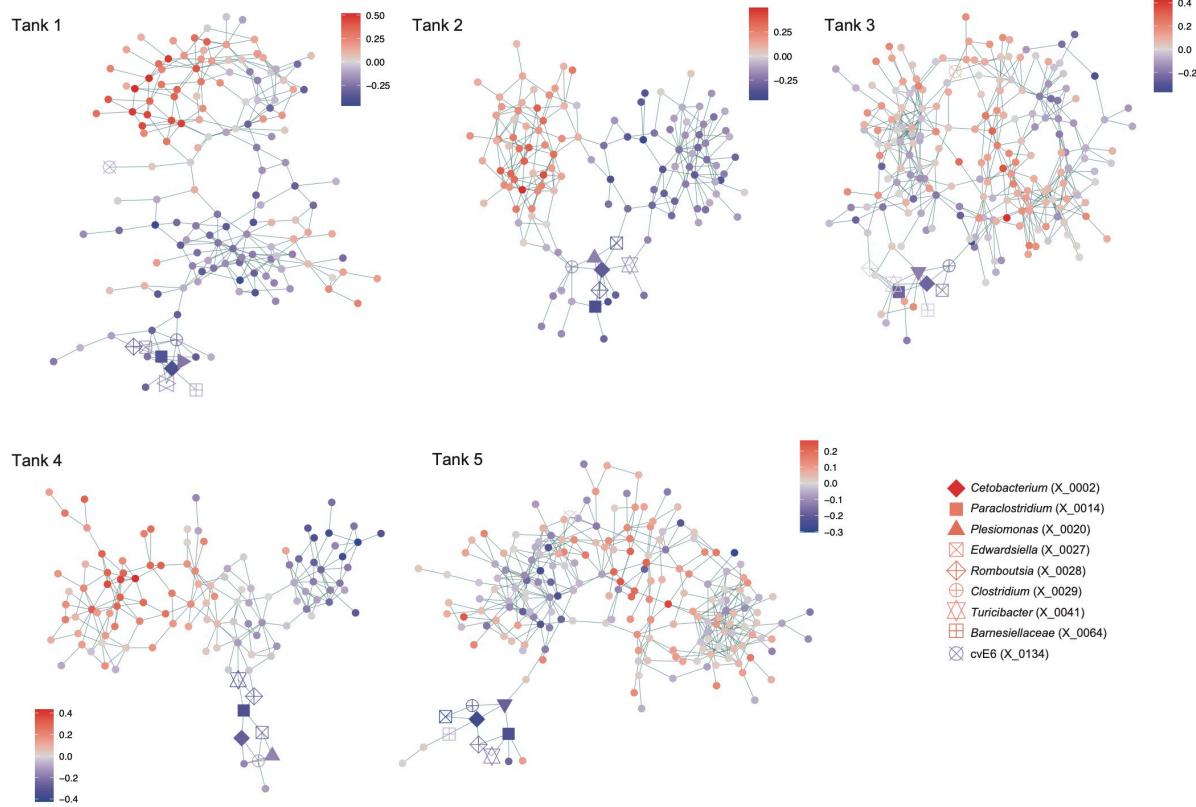
840 **Extended Data Fig. 7 | Taxonomy of the nodes within the coexistence networks.** Within
841 the coexistence networks shown in Figure 4, phylum-level taxonomy of the ASVs is shown.
842 ASVs included in minor sub-networks (number of nodes < 5) are not shown. Only the ASVs
843 that appeared in 30 or more samples were targeted in the analysis of each tank.

844

a Correlation with pH

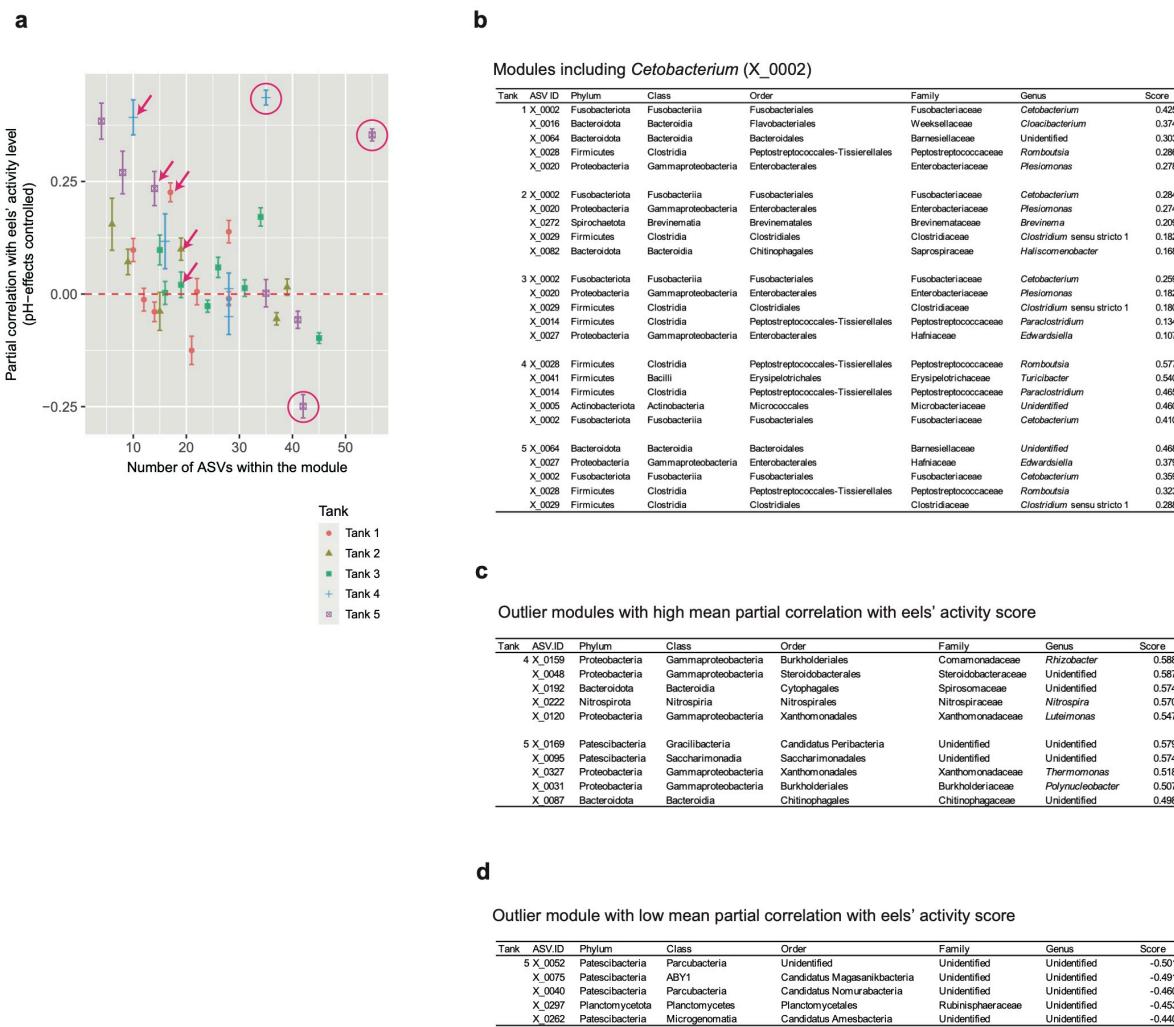


b Correlation with DO



846 **Extended Data Fig. 8 | Correlations with environmental variables.** **a**, Correlation with pH.
847 For each microbial ASV included within the coexistence network of each aquaculture tank
848 (Fig. 4), correlation between absolute abundance and pH is shown. ASVs included in minor
849 sub-networks (number of nodes < 5) are not shown. Only the ASVs that appeared in 30 or
850 more samples were targeted in the analysis of each tank. **b**, Correlation with dissolved oxygen
851 level. For each microbial ASV included within the coexistence network of each aquaculture
852 tank (Fig. 4), correlation between absolute abundance and dissolved oxygen (DO) level is
853 shown.

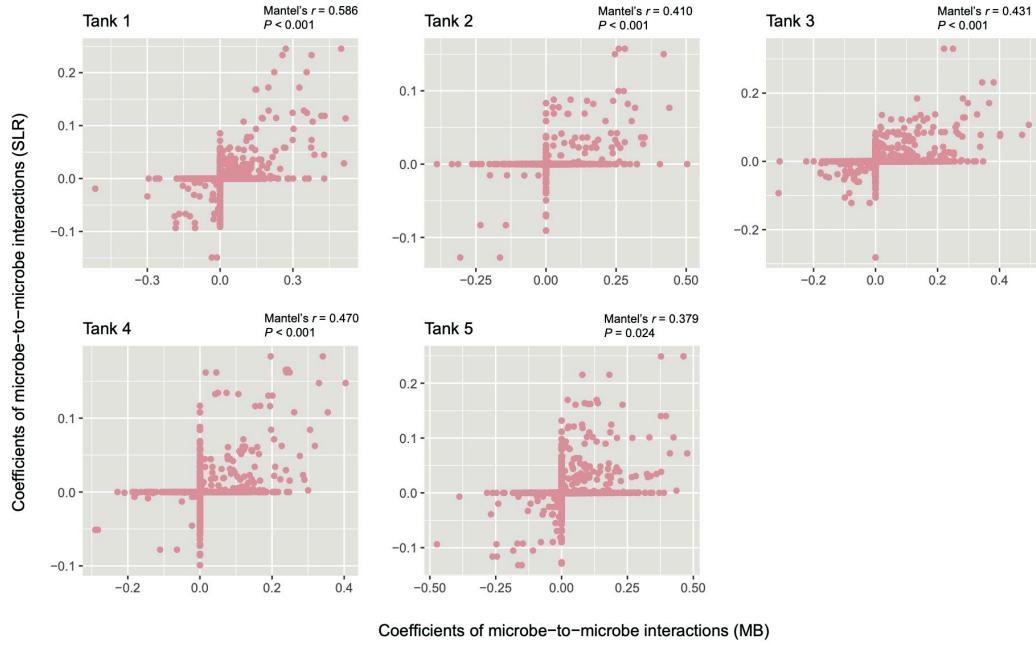
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855

856 **Extended Data Fig. 9 | Properties of network modules.** **a**, Module size and mean partial
857 correlation with eels' activity level. For each module within the coexistence network of each
858 aquaculture tank (Fig. 4), the number of ASVs and mean partial correlation with eels' activity
859 level are shown. The modules including the *Cetobacterium* ASV (X_0002) is indicated by
860 arrows. The outlier modules with large numbers of constituent ASVs and low/high mean
861 partial correlation with eels' activity level are highlighted by circles. **b**, Modules including the
862 *Cetobacterium* ASV (X_0002). The top-five ASVs with the highest partial correlation with
863 eels' activity level are shown for each module. **c**, Outlier modules with high mean partial
864 correlation with eels' activity level. **d**, Outlier module with low mean partial correlation with
865 eels' activity level.

866



867

868 **Extended Data Fig. 10 | Comparison of network reconstruction methods.** For each
869 aquaculture tank, the network links inferred with the MB method was compared with those
870 inferred with the SLR methods. The former is expected to represent interspecific interactions
871 as well as potential sharing of environmental preference (i.e., niches) between nodes (ASVs).
872 Meanwhile, the latter is expected to represent direct interactions between nodes. A
873 positive/negative value indicates a potentially positive/negative interaction between a pair of
874 microbial ASVs. The positive values (> 0) were used to draw networks of potential positive
875 interactions between microbes as shown in Figure 5.