

1 **Extinction cascades, community collapse, and recovery across a Mesozoic**  
2 **hyperthermal event**

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17 **Biotic interactions and community structure are seldom examined in mass extinction**  
18 **studies but must be considered if we are to truly understand extinction and recovery**  
19 **dynamics at the ecosystem scale. Here, we model shallow marine food web structure**  
20 **across the Toarcian extinction event in the Cleveland Basin, UK using a trait-based**  
21 **inferential modelling framework. First, we subjected our pre-extinction community to**  
22 **extinction cascade simulations in order to identify the nature of extinction selectivity**  
23 **and dynamics. Second, we tracked the pattern and duration of the recovery of**  
24 **ecosystem structure and function following the extinction event. In agreement with**  
25 **postulated scenarios, we found that primary extinctions targeted towards infaunal and**  
26 **epifaunal benthic guilds reproduced the empirical post-extinction community. These**  
27 **results are consistent with geochemical and lithological evidence of an anoxia/dysoxia**  
28 **kill mechanism for this extinction event. Structural and functional metrics show that**

29 **the extinction event caused a switch from a diverse, stable community with high levels**  
30 **of functional redundancy to a less diverse, more densely connected, and less stable**  
31 **community of generalists. Ecological recovery appears to have lagged behind the**  
32 **recovery of biodiversity, with most metrics only beginning to return to pre-extinction**  
33 **levels ~7 million years after the extinction event. This protracted pattern supports the**  
34 **theory of delayed benthic ecosystem recovery following mass extinctions even in the**  
35 **face of seemingly recovering taxonomic diversity.**

36

37 Earth has experienced a number of mass extinction events that have shaped evolutionary  
38 history, not only by the dramatic loss of species over relatively short periods of time but also  
39 by repeatedly restructuring ecosystems<sup>1</sup>. Many mass extinctions have been linked to rapid  
40 releases of greenhouse gases into the atmosphere by large igneous province (LIP) volcanism,  
41 which led to a cascade of environmental effects, including rapid global warming, ocean anoxia,  
42 and ocean acidification<sup>2</sup>. Thus, palaeobiologists have long viewed mass extinctions as the  
43 quintessential example of a Court Jester driver of macroevolution, whereby external abiotic  
44 environmental pressures act as the predominant driver of macroevolution across long  
45 geological timescales<sup>3</sup>. However, community ecological theory suggests that biotic  
46 interactions, and thus Red Queen processes<sup>4</sup>, would have also played a major role during  
47 mass extinction events. Many victims of past extinction events were unlikely to have gone  
48 extinct as a direct effect of abiotic stress, but probably did so in response to cascading  
49 secondary effects via the loss of key prey sources<sup>5</sup>.

50 Macroecological studies of selectivity<sup>6,7</sup>, functional diversity loss<sup>8,9</sup>, and recovery<sup>10,11</sup>  
51 indicate that warming-related mass extinctions are associated with latitudinal extinction  
52 selectivity<sup>6,12,13</sup> as well as preferential loss of taxa vulnerable to hypercapnia, anoxia and  
53 acidification, and photosymbiotic taxa<sup>6,13-15</sup>. All of these patterns can be qualitatively linked to  
54 one of the aforementioned abiotic effects of LIP eruptions, but it remains incredibly difficult to  
55 ascertain the full range of, or relative impact of, abiotic stressors that contributed to a particular  
56 mass extinction event.

57       Ecological theory suggests that extinction dynamics are most effectively studied within  
58 a community framework where details about interactions among taxa may accentuate or buffer  
59 the responses of individual taxa from direct (i.e. primary) and cascading 'secondary'  
60 extinctions<sup>16</sup>. In fact, many victims of mass extinctions are unlikely to have become extinct as  
61 a direct effect of abiotic stress, but probably did so in response to cascading secondary  
62 effects<sup>5</sup>. Biotic interactions are seldom considered in mass extinction studies (but see<sup>17-19</sup>) as  
63 it is very difficult to ascertain consumer-resource and other interactions between long extinct  
64 organisms. However, if we are to truly understand mass extinction dynamics, we must quantify  
65 such interactions as many prominent extinction patterns, such as the high levels of extinction  
66 amongst pelagic predators<sup>6</sup>, are difficult to explain in the absence of extinction cascades  
67 through communities<sup>20</sup>.

68       In addition to the study of mass extinction causality, magnitude, and selectivity,  
69 extinction recovery has also become an intensely studied topic over the past decade<sup>21</sup>, driven  
70 in part by the desire to understand how long the planet may take to recovery from the current  
71 "Sixth Mass Extinction"<sup>22</sup>. Studies based on taxonomic and functional diversity suggest full  
72 ecosystem recovery can take anywhere between 0.7 to 50 million years from the largest mass  
73 extinctions<sup>10,11,23</sup>, but it is possible that ecosystem function could recover despite persistent  
74 low levels of biodiversity. Thus, studies of extinction recovery could be greatly improved by  
75 adopting a community ecology approach that integrates across biodiversity, community  
76 structure and ecosystem function.

77       Here, we utilise a community ecology food web approach to model primary and  
78 secondary extinction dynamics and community recovery across the early Toarcian extinction  
79 event (ETE; ~183 Ma) from the Cleveland Basin, Yorkshire, UK. Specifically, we use  
80 ecological trait data to reconstruct plausible food webs. We then subject these food webs to  
81 several primary extinction scenarios that link event characteristics (e.g. dysoxia, acidification,  
82 warming) to traits. We use well established ecological modelling tools to evaluate patterns of  
83 secondary extinction, ultimately identifying several target traits and species whose sensitivity  
84 to the ETE event led to the ensuing post event community structure. Finally, we also look at

85 empirical patterns of recovery from this extinction event, detailing changes in biodiversity,  
86 functional groups and community structure.

87 The ETE, is traditionally referred to as a second order extinction event<sup>24</sup> (i.e. an  
88 extinction event that caused less than 40% generic extinction<sup>25</sup>) and resulted in the loss of  
89 around 26% of marine genera globally<sup>26</sup>. It is linked to the eruption of the Karoo-Ferrar Large  
90 Igneous Province<sup>27</sup> which resulted in a globally distributed negative carbon isotope shift<sup>28,29</sup>,  
91 hyperthermal warming of up to 13°C in the midlatitudes<sup>30,31</sup>, prolonged regional ocean dysoxia  
92 and anoxia<sup>26,32-34</sup>, and ocean acidification<sup>35</sup>.

93 In the Cleveland Basin and in much of the NW Tethyan basins of NW Europe the ETE  
94 is coincident with the deposition of finely laminated, organic-rich, black shales which signify  
95 persistent dysoxia/anoxia, at shallow depths on the continental shelf, termed the Toarcian  
96 Oceanic Anoxic Event (TOAE or Jenkins Event)<sup>33,36</sup>. The ETE resulted in the loss of around  
97 60% of marine species within the Cleveland Basin (87% benthic species extinction)<sup>24,37</sup>, with  
98 post-extinction benthic communities made up of low diversity/high abundance assemblages  
99 of hypoxia-tolerant and opportunistic species<sup>26,34,38,39</sup>. Recovery seems to occur in two pulses  
100 and, in total, takes as long as 7 million years in terms of both taxonomic and functional  
101 diversity<sup>37,40</sup>.

102 We use a data set of 38,670 occurrences of 162 species of marine invertebrates, fish,  
103 and trace fossils derived from years of detailed field studies<sup>24,26,41,42</sup> of one of the most  
104 expanded Pliensbachian to Toarcian sections in the world to produce a series of community  
105 trophic networks (i.e. food webs) (Fig. 1). We aim to ascertain: (i) the most plausible set of  
106 traits and species impacted by primary extinction events and the nature of any secondary  
107 extinction cascades, that led to the post extinction community structure and diversity; (ii) the  
108 likely environmental trigger of the extinction cascades; and (iii) the pattern and duration of  
109 ecological recovery in the aftermath of the extinction event.

110

## 111 **Results and discussion**

### 112 **The Early Toarcian extinction event**

113 *Empirical data: pre-extinction.*

114 The pre-extinction community is characterised by a diverse assemblage of benthic and pelagic  
115 taxa belonging to 48 different trophic guilds (Fig. 1; Fig 2A; see supplementary data for  
116 network metrics values for all communities). The community consists of a range of primary  
117 consumers (including suspension feeders, deposit feeders, miners, and grazers), intermediate  
118 predators (cephalopods, crustaceans, fish, gastropods, scaphopods), and top predators  
119 (larger cephalopods). The pre-extinction community is characterised by values of many  
120 common network metrics that are well within the bounds for typical modern day marine  
121 communities (e.g. connectance, generality, vulnerability, mean/max. trophic level etc.)<sup>5,39,43</sup>  
122 (Fig 2B).

123

124 *Empirical data: post-extinction.*

125 The ETE causes a reduction in richness from 48 to 21 trophic guilds (Fig. 1; Fig 2A) and drove  
126 disproportionate losses amongst infaunal, large, highly motile, and predatory benthic guilds  
127 (~80% benthic species extinction<sup>24,37</sup>). Post-extinction benthic assemblages were dominated  
128 by low-diversity/high-abundance communities of very small, epifaunal, suspension-feeding  
129 bivalves, most notably the presumably low-oxygen-tolerant opportunistic species *Bositra*  
130 *buchii* and *Pseudomytiloides dubius*<sup>34</sup>. Although previous research has suggested that pelagic  
131 taxa were also affected by the ETE<sup>44</sup> (including high mortality amongst pelagic faunas and  
132 cephalopods migrating away from the Cleveland Basin in response to high temperatures and  
133 low food supply), at the structural level, pelagic elements of the community appear much less  
134 affected.

135 The severe benthic losses at the guild level are in contrast to global studies of mass  
136 extinctions which have postulated that, although species losses are typically severe, mass  
137 extinctions rarely cause global guild-level extinction<sup>8,45</sup>. However, the loss of >50% of guilds  
138 in this local, mid-latitude setting is in keeping with regional studies of extinction across the  
139 ETE<sup>24,26,37,46-49</sup>, that have shown species and guild-level losses to be much higher in the mid  
140 latitudes (i.e. NW Tethys and NE Panthalassan margin) than globally<sup>6,45</sup>. The post-extinction

141 benthic assemblage is dominated by a low-diversity/high-abundance community of small-  
142 bodied, epifaunal, suspension feeders with apparent selectivity against larger benthic taxa  
143 with more active modes of life. Together with lithological and geochemical evidence, this  
144 strongly supports the theory of a dysoxia/anoxia driven extinction in the Cleveland Basin<sup>41</sup>.

145 Overall community connectivity (i.e. connectance) increased after the ETE (Fig. 2B)  
146 which also corresponds with an increase in generality, vulnerability, and maximum trophic  
147 level within the community (Fig. 2B-E). Together with a reduction in the number of linear chains  
148 within the food web (S1), the levels of omnivory (S2) and both apparent (S4) and direct  
149 competition (S5) (Fig. 3A-D) this suggests that the post-extinction community consists of fewer  
150 guilds that are more generalistic in their feeding habits and thus taxa are more closely linked  
151 to one another via consumer-resource interactions than taxa in the pre-extinction community.  
152 Selective extinction of benthic taxa, which are predominantly lower- and intermediate-level  
153 consumers, led to the food web becoming taller and thinner (i.e. fewer linear chains) with more  
154 restricted energy flows, fewer lower-level consumers, and increased predation pressure on  
155 the remaining lower trophic level taxa (i.e. increased vulnerability). This also led to reduced  
156 direct completion (S5) as benthic predators disappear from the community and reduced  
157 apparent competition (i.e. predator choice; S4) as the extinction wiped out the majority of the  
158 benthic guilds such that pelagic predators had fewer prey options.

159 The post-extinction community also displays lower levels of omnivory as the  
160 intermediate consumers in the benthic realm go extinct, meaning that top predators are now  
161 feeding across fewer trophic levels. This pattern of densely connected, species-poor  
162 communities of opportunists/generalists is consistent with evidence from palaeoecological  
163 interpretations of the fossil record (i.e. low-diversity/high-abundance communities of  
164 opportunistic species)<sup>34,37</sup> and other instable post-mass extinction food webs reconstructed  
165 from the fossil record<sup>18,19,50</sup>.

166 This post extinction community bears a broadly similar structure to that of modern low-  
167 diversity communities dominated by generalists<sup>43</sup>. Although there is some empirical evidence  
168 that higher connectance leads to higher community robustness (i.e. more stable communities

169 that are less likely to collapse)<sup>51,52</sup>, the taller, thinner web with reduced omnivory values of the  
170 post-extinction web suggests instability as energy flows are contingent on very specific  
171 pathways and presents a rivet-hypothesis scenario where the removal of a few well-connected  
172 guilds could lead to wholesale ecosystem collapse<sup>52,53</sup>. The removal of intermediate  
173 consumers (i.e. benthic crustaceans, gastropods, scaphopods) increases the vulnerability of  
174 the few remaining species of primary consumers in the benthic realm as they continue to be  
175 predated by a similar number of pelagic higher level consumers (i.e. cephalopods) as seen in  
176 the pre-extinction community. This change in the tiering “pyramid” suggest community-level  
177 patterns reflect patterns seen in the global ecosystem in the aftermath of mass extinction  
178 events that have strong bottom-water dyoxia/anoxia drivers<sup>11</sup> as well as communities in  
179 present-day deoxygenated areas<sup>54</sup>.

180

#### 181 *Extinction dynamics*

182 We simulated, with replication, 13 different primary extinction scenarios (see Methods). Each  
183 of these scenarios generated unique signatures of primary and secondary extinction. All  
184 scenarios with identical starting guild richness (i.e. 48) were terminated when the simulations  
185 reached the post-extinction richness of 21. We calculated multiple structural metrics and a  
186 True Skill Statistic (TSS)<sup>55</sup> to compare the simulated post-extinction events to the actual post-  
187 extinction event.

188 Three of the simulated extinction scenarios produced communities that were much  
189 more similar to the empirical post-extinction community than the random primary extinction  
190 scenario (Fig. 4A). According to the TSS scores, primary extinction selectivity based on  
191 tiering, with strongest extinction selectivity against infaunal taxa, gave by far the closest  
192 replication of the empirical post-extinction community (Fig. 4A). Targeting generality was the  
193 second most plausible scenario. The other 10 were largely indistinguishable from random (Fig  
194 4a).

195 *Focus: tiering.* The similarities between the empirical post-extinction community and  
196 the simulated extinction scenario based on primary extinction selectivity against infaunal and

197 benthic taxa suggests that the strongest extinction selectivity was placed upon taxa living  
198 within the sediment or on the seabed in the benthic realm. This pattern matches (i) the  
199 empirical data where the majority of extinction occurs within the benthic realm (with almost all  
200 infaunal guilds disappearing) and (ii) an anoxia/dysoxia kill mechanism dictating that bottom  
201 waters would contain less oxygen than sea surface layers with the capacity for continuous gas  
202 exchange with the atmosphere. This result is in line the lithological evidence for anoxia/dysoxia  
203 (i.e. organic-rich black shales) within the Cleveland Basin<sup>38</sup> and provides confidence that our  
204 method is performing well.

205 Extinction simulations based on tiering (with extinction selectivity infaunal>pelagic)  
206 also produced the closest matches to the empirical post extinction community in terms of  
207 community structure and dynamics in 3 out of 13 of the metrics used (i.e. vulnerability, mean  
208 trophic level, S1: number of linear chains). Extinction selectivity based on tiering also produced  
209 the closest to the post-extinction community in terms of vulnerability when extinction selectivity  
210 was reversed (i.e. pelagic>infaunal), but this was only marginally better than extinction  
211 selectivity based on tiering (infaunal>pelagic) (Fig 4). Simulations based on extinction  
212 selectivity against tiering from infaunal to pelagic also produced relatively close matches to  
213 the empirical community in all other metrics apart from generality (Fig. 4).

214 *Focus: generality.* Scenarios with primary extinction selection based on generality (i.e.  
215 the number of resource connections per guild) more closely replicate the post-extinction  
216 community than random primary extinction selection with selectivity based both ways (i.e. high  
217 to low and low to high) producing almost equally good matches to the empirical data (Fig. 4A).  
218 This is likely because extinction selectivity was strongest in the benthic realm, which was made  
219 up of taxa that either (i) were primary consumers (i.e. bivalves, brachiopods, crinoids) and only  
220 fed upon the basal node of the food web (i.e. low generality) or (ii) were intermediate  
221 consumers (i.e. crustaceans, gastropods) that had a broad trophic niche and were feeding  
222 upon multiple primary and secondary consumer nodes (i.e. high generality).

223 Simulations based on selection on generality produced less accurate results than  
224 tiering in general, but provided the closest match to the empirical post-extinction community

225 in terms of generality, maximum trophic level (when extinction selectivity is high>low) and  
226 diameter, betweenness, S2: omnivory (when extinction selectivity is low>high) (Fig. 4). These  
227 results add confidence to the TSS results which show that extinction selectivity was strongest  
228 in the benthic realm (where taxa of lowest and highest generality are found) in which the effects  
229 of dysoxia/anoxia would be strongest felt.

230 *Focus: motility.* The only other simulated extinction scenarios that produced close  
231 matches to the empirical post-extinction community in terms of structure and dynamics were  
232 based on motility (with extinction selectivity non-motile>fast; which produced the closest match  
233 for connectance, mean degree, system omnivory index, S4: apparent competition, and S5:  
234 direct competition; Fig. 4). This suggests that motility also had a strong bearing on extinction  
235 selectivity across the ETE, with non-motile taxa being more highly prone to extinction than  
236 motile taxa. Whilst it is evident that the ETE sees the elimination of almost all the motile  
237 benthos in the Cleveland Basin, there are also catastrophic losses amongst the non-motile to  
238 facultatively motile benthos with the post-extinction community being dominated by 2 to 3 non-  
239 motile and facultatively motile taxa with low-oxygen requirements, similarly to modern day low-  
240 oxygen ecosystems<sup>38</sup>. However, simulation scenarios of primary extinction selection based on  
241 motility perform no better than random primary selection scenarios in regard to guild  
242 occupancy in the post-extinction community, thus suggesting that other traits (i.e. tiering) were  
243 more important in determining primary extinction vulnerability.

244 *Focus: other possibilities.* Interestingly, some traits that have previously been identified  
245 as key determinants of extinction across hyperthermal events, i.e. body size and  
246 calcification<sup>14,56</sup>, did not produce simulated post-extinction communities that were a closer  
247 match to the empirical post-extinction communities than random selection (Fig. 4). This  
248 suggests that the main extinction driver (i.e. dysoxia/anoxia) was not selective based on body  
249 size to the same degree as selectivity based on tiering and/or motility. The lack of apparent  
250 primary extinction selection based on calcification also suggests that ocean acidification was  
251 not a major extinction driver across the ETE in the Cleveland Basin.

252

253 **Ecosystem recovery following the ETE**

254 *Early recovery.* The early recovery interval sees an increase in richness as a number of guilds  
255 return that were absent from the basin during the post-extinction interval. This return of guilds  
256 is associated with the re-oxygenation of the benthic realm (Fig. 2A). However, despite the  
257 return of some species occupying motile benthic and infaunal guilds, the majority of new  
258 species occupy guilds that were also present during the immediate post-extinction interval (i.e.  
259 surficial suspension feeders and pelagic predators). The lower stratigraphic sections from the  
260 early recovery interval are characterised by abundant *Dacromya ovum*, a shallow infaunal,  
261 low oxygen tolerant, mining bivalve. It has been hypothesised that *D. ovum* may have been  
262 able to survive epifaunally before oxygenation improved within the sediment and then  
263 subsequently acted as an ecosystem engineer that catalysed the re-oxygenation of the  
264 sediment via bioirrigation<sup>38</sup>. *D. ovum* is representative of the first infaunal guild to reappear,  
265 some 1 million years after the extinction event<sup>38,39</sup>, and is then followed by subsequent shallow  
266 and deep infaunal taxa by the end of the early recovery interval. Whilst the epifaunal  
267 community is still dominated by sessile or facultatively mobile taxa, oxygenation of bottom  
268 waters is further indicated by the return of motile predatory and grazing guilds (i.e. gastropods  
269 and echinoids) as well as the establishment of a greater diversity of soft-bodied epifaunal  
270 grazing and infaunal mining guilds (i.e. trace fossils).

271 Metrics of food web structure suggest that ecosystem recovery is also well underway  
272 in the early recovery interval. This period sees connectance and generality returning to lower  
273 pre-extinction levels, in-line with increased guild diversity (Fig. 2B and D). Despite a significant  
274 recovery of guild diversity and some structural metrics returning towards pre-extinction levels,  
275 full ecosystem recovery does not appear to have happened in the early recovery interval. This  
276 is evidenced by a paucity of infaunal tiering and motile benthos, as compared to the pre-  
277 extinction interval and several of the structural metrics and motifs remain at similar levels to  
278 the post-extinction interval rather than starting to return to pre-extinction levels. For example,  
279 maximum trophic level and vulnerability (Fig. 2C and E) remain very high and all the food web  
280 motifs remain at levels closer to the post-extinction interval than the pre-extinction interval

281 (Fig. 3). This suggests that the early recovery community is still tall, thin and top-heavy with  
282 somewhat restricted energy flows consisting of a diverse assemblage of pelagic predators  
283 feeding on a still relatively depauperate assemblage of lower-level consumers. This pattern is  
284 in contrast to some previous models of ecosystem recovery following mass extinctions that  
285 postulate that more-basal trophic levels recovered more quickly than upper trophic levels<sup>10</sup>.  
286 Instead, this pattern supports the hypothesis of delayed benthic ecosystem recovery following  
287 mass extinctions even in the face of seemingly recovering taxonomic diversity<sup>11</sup>.

288

289 *Late recovery.* The late recovery interval witnesses a further increase in guild richness and  
290 sees all the structural metrics and motifs return, or start to return, to levels seen in the pre-  
291 extinction community (Figs. 2 and 3). Although many of the taxa are different (at species level)  
292 to those of pre-extinction community, the majority of pre-extinction guilds are re-occupied by  
293 the late recovery interval. Connectance, generality, and vulnerability (Fig. 2B, D-E) are now at  
294 levels comparable to the pre-extinction community, as are levels of omnivory (S2) and  
295 apparent (S4) and direct (S5) competition (Fig. 3B-D). This suggests that intraguild diversity  
296 and functional redundancy is recovering – the reconstructed network indicates a greater  
297 number of predators are feeding upon a greater number of prey species thus increasing  
298 competition for prey and predator choice simultaneously. The recovery of lower and  
299 intermediate-level consumers in the benthic realm drove an increase in the number of linear  
300 chains and omnivory, although these metrics are still distinctly lower than the levels seen in  
301 the pre-extinction community (Fig. 3A-B).

302 Together with a further rise in maximum trophic level, these changes in the late  
303 recovery phase suggests that the overall shape of the food web still remains much taller and  
304 thinner than in the pre-extinction community (Fig. A and Fig 3A). Although the increase in the  
305 number of linear chains in the late recovery as compared to the early recovery suggests that  
306 food web shape could be starting to return towards pre-extinction levels with the greater  
307 diversity of benthic guilds, it may also be a result of changing ecosystem structure brought  
308 about by the progression of the Mesozoic Marine Revolution (MMR)<sup>57</sup>. The late recovery

309 interval contains a much more diverse array of benthic predators from groups that were  
310 supposedly key drivers of the MMR, such as decapod crustaceans<sup>58</sup>, gastropods<sup>59</sup> and  
311 echinoderms<sup>60</sup> and such changes in benthic community composition may be driving some of  
312 the stepwise increase in maximum trophic level through the system, which deviates from the  
313 common pattern of perturbation before return to pre-extinction levels as seen in most of the  
314 other structural metrics and motifs (Figs 2 and 3).

315

### 316 **Conclusions**

317 The ETE is characterised by marked changes in community structure and function from a  
318 diverse, stable community where each key ecological function is performed by a number of  
319 species/guilds to a less diverse, more densely connected and less stable community of  
320 generalistic “disaster taxa”<sup>39</sup> in which nodes (i.e., species/guilds) are more interdependent  
321 than in the pre-extinction community. This change from a diverse pre-extinction ecosystem  
322 with high degrees of functional redundancy to a contrasting post-extinction community where  
323 key functions are performed by single guilds is representative of a “rivet hypothesis”<sup>61</sup> or a  
324 “skeleton crew hypothesis”<sup>8,45</sup> model in which the subsequent loss of any “rivet” or “crew  
325 member” may cause the system to collapse.

326 Our extinction cascade simulations suggest that primary extinction targeted towards  
327 infaunal and epifaunal benthic taxa as well as less motile guilds and extreme generalists or  
328 specialists best explain post-extinction community structure and function. These conclusions  
329 agree with lithological and geochemical evidence for an anoxia/dysoxia kill mechanism which  
330 would primarily target benthic organisms (i.e. sessile suspension feeders or generalist  
331 predators) as well as taxa classified as specialists (i.e. those only feeding on a single prey  
332 source which is predominantly the basal node of the food web) which are mostly benthic  
333 suspension feeders. Despite significant increases in biodiversity during the early recovery  
334 interval, most structural and functional metrics suggest ecosystem recovery to pre-extinction  
335 levels took at least 7 million years (i.e. until the late recovery interval). However, some metrics  
336 suggest that either full recovery had not happened even by the late recovery interval or

337 ecosystem structure and function had re-equilibrated to a different state in the Middle Jurassic  
338 and perhaps represents ecological regime shifts associated with the MMR<sup>57,62</sup>.

339

## 340 **Methods**

341 **Dataset.** Fossil occurrence data is obtained from a compilation of field data sets<sup>24,26,42,63-65</sup>.  
342 The study interval extends from the upper Pliensbachian to the upper Toarcian of the  
343 Cleveland Basin (North Yorkshire, UK; Fig 1.) and provides a high resolution data set across  
344 the ETE. The data set consists of 38,670 specimens of 162 pelagic and benthic  
345 macroinvertebrate species together with occurrences of fish and trace fossils. The data set  
346 was subset into four broad time periods; pre-extinction (*margaritatus-tenuicostatum* zones of  
347 the Staithes Sandstone Formation, Penny Nab and Kettleness Members of the Cleveland  
348 Ironstone Formation and Grey Shales Member of the Whitby Mudstone Formation), post-  
349 extinction (*serpentinum-commune* subzones of the Mulgrave Shale and Alum Shale Members  
350 of the Whitby Mudstone Formation), early recovery (upper *bifrons*-lower *levesquei* zones of  
351 the Alum Shale, Peak Mudstone and Fox Cliff Siltstone Members of the Whitby Mudstone  
352 Formation), and late recovery (upper *levesquei* zone of the Grey and Yellow Sandstone  
353 Members of the Blea Wyke Sandstone Formation) (Fig. 1).

354

355 **Defining organism ecologies, feeding interactions and trophic guilds.** Modes of life were  
356 defined for each fossil species based on the ecological traits defined in the Bambach ecospace  
357 model<sup>66</sup> (i.e. motility, tiering, and feeding). Ecological traits were assigned based on  
358 interpretations in the published literature which are largely based on functional morphology  
359 and information from extant relatives. Information on the body size of each species was also  
360 recorded by summarising mean specimen sizes from the section into a categorical  
361 classification. The following ecological characteristics were recorded for each fossil species;  
362 motility (fast, slow, facultative, non-motile), tiering (pelagic, erect, surficial, semi-infaunal,  
363 shallow infaunal, deep infaunal), feeding (predator, suspension feeder, deposit feeder, mining,  
364 grazer), and size: gigantic (>500 mm), very large (>300-500mm), large (>100-300mm),

365 medium (>50-100mm), small (>10-50mm), tiny ( $\leq$ 10mm). Size categories are defined by the  
366 longest axis of the fossil, estimates of tracemaker size from trace fossils based on literature  
367 accounts, or by extrapolating the total length for belemnites from the preserved guard using  
368 established approaches<sup>67,68</sup>. A single node for primary producers was added to each food web  
369 to ensure that primary consumers were not considered as primary producers in the  
370 reconstructions. Feeding interactions were then modelled between organisms based on an  
371 inferential model which assigns the possibility of encounter and consumption of prey items  
372 using rules defined by inferred ecological foraging traits (i.e., motility, feeding, tiering, and size;  
373 Fig. 5). Trophic guilds were defined by unique combinations of ecological and foraging traits  
374 (see Supplementary Materials for a full list of trophic guilds and their defining characteristics)  
375 which correspond to groups of organisms that have the same predation/prey rules dictating  
376 their interactions in the model and are thus akin to trophic species often used in the  
377 reconstruction of modern food webs<sup>19,69</sup>. Food webs were produced for each broad time  
378 interval (i.e. pre-extinction, post-extinction, early recovery, and late recovery) at both the  
379 species and trophic guild level. Further palaeoecological data, which is used to inform the  
380 extinction cascade simulations, was also assigned to each trophic species/guild in the food  
381 web. This data included motility (fast, slow, facultative, non-motile), tiering (pelagic, epifaunal,  
382 infaunal), size (gigantic, very large, large, medium, small, tiny), and calcification (heavy,  
383 moderate, light).

384 **Quantifying community structure and function.** Community network structural metrics of  
385 size (i.e., richness), connectance (C), maximum trophic level, generality (i.e., in-degree, or  
386 number of prey) and vulnerability (i.e., out-degree, or number of predators) as well as the  
387 network motifs S1 (i.e. number of linear chains), S2 (i.e. omnivory), S4 (i.e. apparent  
388 competition), and S5 (i.e. direct competition) were calculated to track changes in community  
389 structure and function across the extinction and through the recovery interval. Food web  
390 communities were built, visualised and analysed using the R package *PFIM*.

391 **Simulating extinction cascades.** Extinction cascades were simulated by subjecting guilds in  
392 the pre-extinction community to primary extinction scenarios based on ecological and trophic

393 traits that correspond to known vulnerabilities linking the traits to hypothesised mass extinction  
394 drivers of anoxia, thermal stress, acidification. For each replicate, we catalogued the timing  
395 and identity of all primary extinctions and any secondary extinctions arising when a guild lost  
396 all of its resources. The extinction cascades were stopped when the diversity of the simulated  
397 post-extinction community reached 21 species and thus equalled that of the empirical post-  
398 extinction community.

399 We explored 13 different scenarios. Simulations were run with primary extinctions  
400 selected (i) randomly, (ii/iii) by body size (large to small/small to large), (iv/v) by tiering (infaunal  
401 to pelagic/pelagic to infaunal), (vi/vii) by motility (fast to non-motile/non-motile to fast), (viii/ix)  
402 calcification (heavy to light/light to heavy), (x/xi) generality (low to high/high to low), and (xii/xiii)  
403 vulnerability (low to high/high to low).

404 We implemented the modelling using the cheddar package in R<sup>70</sup> using the  
405 *RemoveNodes()* function with the 'cascade' method of for secondary extinctions. We  
406 generated 50 replicates for each scenario by sampling among guilds from within each traits'  
407 levels in the sequence. For example, tiering has six levels (see above) and we randomised  
408 the primary extinction sequence of each guild within each of these levels.

409 Simulated post-extinction food webs were then compared to the empirical post-  
410 extinction community using three approaches. First, we compared nine structural metrics  
411 between the empirical post-extinction web and the simulated networks. Second, we compared  
412 the frequency of four motifs (S1: number of linear chains; S2: number of omnivory motifs; S4:  
413 number of apparent competition motifs; S5: number of direct competition motifs) between the  
414 empirical post-extinction web and the simulated networks. Third, we used a True Skill  
415 Statistic<sup>55</sup> (TSS/classification-misclassification table/confusion matrix: true positive, true  
416 negative, false positive, false negative) to compare the guild-node level similarities of  
417 position/identity between the empirical post-extinction web and the simulated networks. All  
418 metric calculations were made with functions coded using the *PFIM* package for R.

419 We combined the inference from all three of these comparisons to identify the most  
420 plausible set of primary extinction and associated secondary extinction scenarios (e.g. which

421 trait sequence) that could deliver a community that most closely resembles the post-extinction  
422 community.

423

424 **Data availability**

425 Contact authors for access to data.

426

427 **Code availability**

428 PFIM is currently under publication, contact authors for access to code.

429

430 **References**

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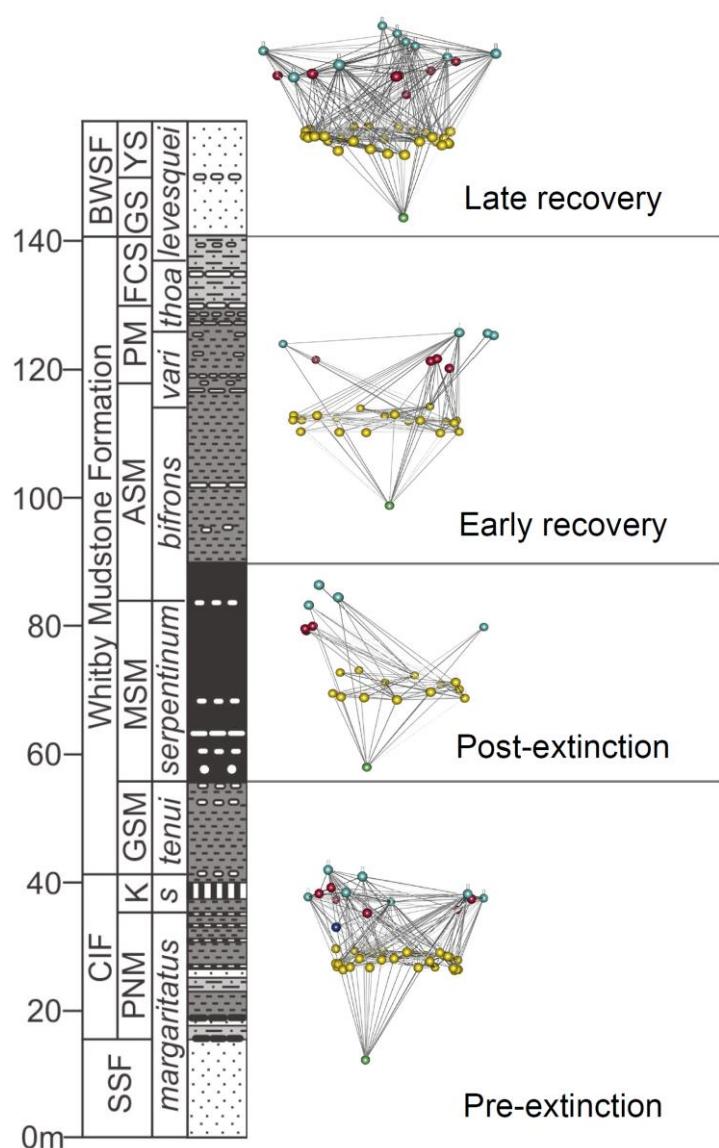
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650

#### 651 **Author information**

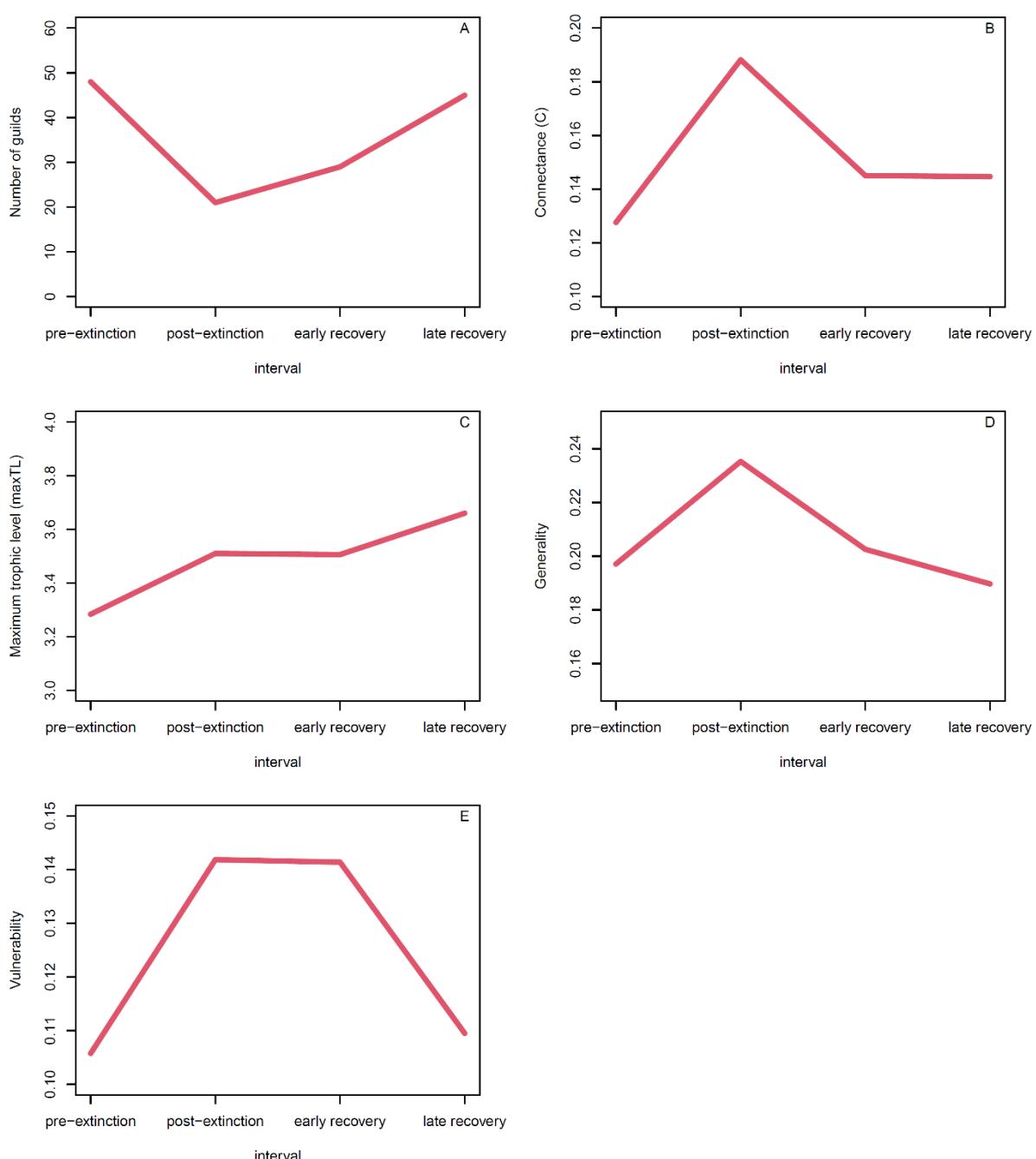
652 A. Dunhill devised the project. K. Zarzyczny, J. Atkinson, C. Little and A. Dunhill collected the  
653 data. A. Dunhill, A. Beckerman, J. Shaw and K. Zarzyczny performed the analysis. A. Dunhill  
654 lead the write up and all authors contributed to editing and improving the manuscript.



655

656 **Figure 1.** Stratigraphic column of the Pliensbachian-Toarcian (Lower Jurassic) of the  
657 Cleveland Basin at Ravenscar (North Yorkshire, UK) showing community food webs for pre-  
658 extinction, post-extinction, early recovery and late recovery intervals. Lithostrat column  
659 abbreviations need explanation. SSF = Staithes Sandstone Fomration; CIF = Cleveland  
660 Ironstone Formation; BWSF = Blea Wyke Sandstone Formation; PNM = Penny Nab Member;  
661 K = Kettleness Member; GSM = Grey Shales Member; MSM = Mulgrave Shales Member;  
662 ASM = Alum Shales Member; PM = Peak Mudstone Member; Fox Cliff Sandstone Member =  
663 ; GS = Grey Sandstone Member; YS = Yellow Sandstone Member; *s* = *spinatum*; *vari* =  
664 *variabilis*; *thoa* = *thouarsense*.

665

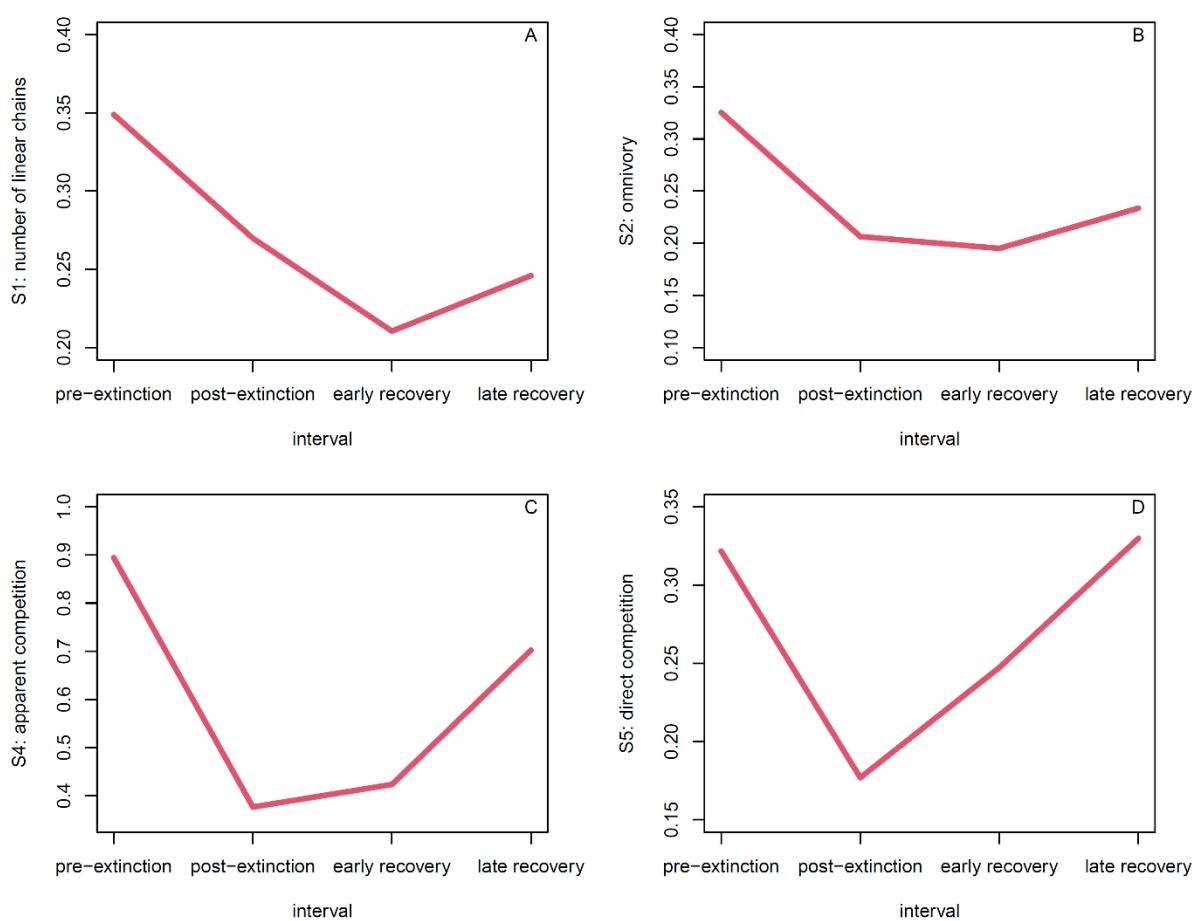


666 **Figure 2.** Structural food web metrics across the four intervals of the ETE; (A) guild richness;

667 (B) connectance (C); (C) maximum trophic level (MaxTL); (D) generality; (E) Vulnerability.

668

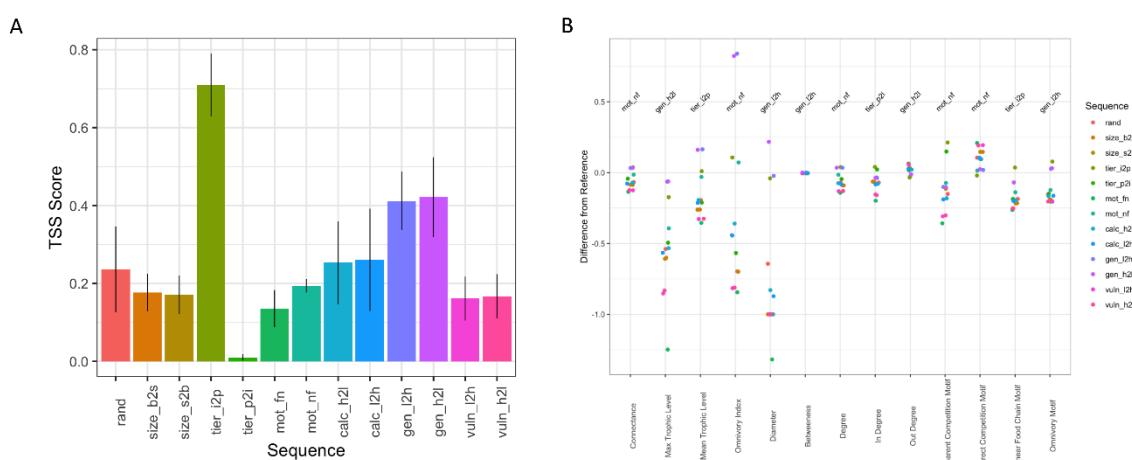
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670 **Figure 3.** Functional food web motifs across the four intervals of the ETE; (A) S1: number of  
671 linear chains; (B) S2: number of omnivory motifs; S4: number of apparent competition motifs;  
672 (D) S5: number of direct competition motifs.

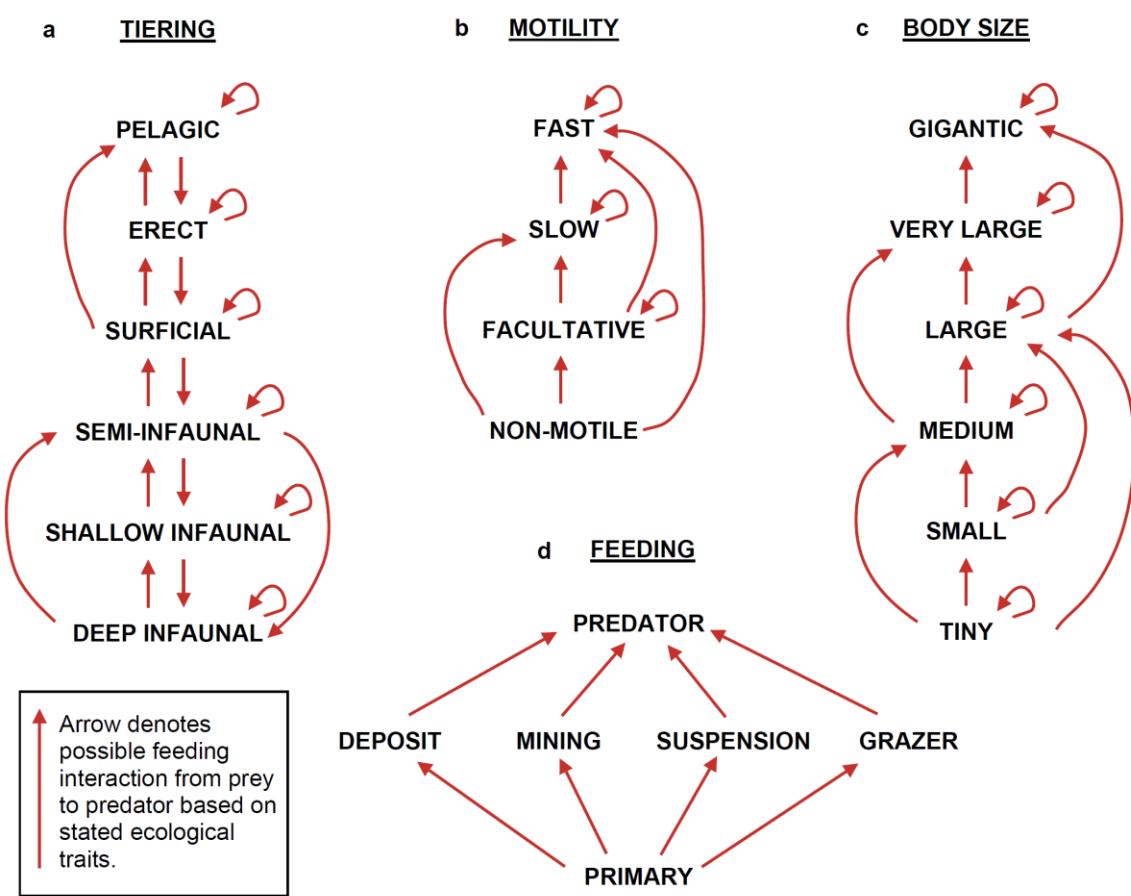
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674



675 **Figure 4.** Results of secondary extinction cascade simulations showing similarity between  
676 simulated and empirical pre-extinction food web structure and function via; (A) True Skills  
677 Statistic showing the similarity in node presence/absence between simulated scenarios and  
678 empirical post-extinction community; (B) similarity in structural food web metrics and food web  
679 motifs between simulated scenarios and empirical post-extinction community (see Table 1 for  
680 sequence definitions).

681



682

683 **Figure 5.** Trait-based feeding rules that parameterise PFIM for reconstructing empirical food  
684 webs across all intervals.

685

**Table 1.** Definitions of extinction cascade simulation sequences.

Sequence	Definition
rand	Primary extinctions targeted in a random fashion
size_b2s	Primary extinctions targeted at larger guilds
size_s2b	Primary extinctions targeted at smaller guilds
tier_i2p	Primary extinctions targeted at infaunal > epifaunal > pelagic guilds
tier_p2i	Primary extinctions targeted at pelagic > epifaunal > infaunal guilds
mot_fn	Primary extinction targeted at most motile guilds
mot_nf	Primary extinction targeted at least motile guilds
calc_h2l	Primary extinction targeted at most heavily calcified guilds
calc_l2h	Primary extinction targeted at least heavily calcified guilds
gen_l2h	Primary extinction targeted at specialist guilds
gen_h2l	Primary extinction targeted at generalist guilds
vuln_l2h	Primary extinction targeted at least heavily predated guilds
vuln_h2l	Primary extinction targeted at most heavily predated guilds