

1 Behavioural responses to acute warming precede critical shifts in the cellular and
2 physiological thermal stress responses in fish.

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10
11 **Abstract**

12 From a conservation perspective, it is important to identify when sub-lethal temperatures begin
13 to adversely impact an organism. However, it is unclear whether, during acute exposures, these
14 cellular thresholds occur at similar temperatures to other physiological or behavioural changes.
15 To test this, we estimated temperature preference (15.1 ± 1.1 °C) using a shuttle box, thermal
16 optima for aerobic scope (10–15 °C) using respirometry, agitation temperature (22.0 ± 1.4 °C) as
17 the point where a fish exhibits a behavioural avoidance response and the CT_{max} (28.2 ± 0.4 °C) as
18 the upper thermal limit for 1 yr old Brook Trout (*Salvelinus fontinalis*) acclimated to 10 °C. We
19 then acutely exposed a different subset of fish to these temperatures and sampled tissues when
20 they reached the target temperature or after 60 min of recovery at 10 °C. We used qPCR to
21 estimate mRNA transcript levels of genes associated with heat shock proteins, oxidative stress,
22 apoptosis, and inducible transcription factors. A major shift in the transcriptome response
23 occurred near the agitation temperature, which may identify a link between the cellular stress
24 response and the behavioural avoidance response.

25 **Keywords:** Ectotherm, cellular response, mRNA, Brook Trout, agitation temperature, Aerobic
26 scope, CT_{max}

27 **Introduction**

28 Many riverine systems are experiencing increasing water temperatures due to climate change and
29 habitat modifications (Bowerman et al., 2021; Mauger et al., 2017; Morash et al., 2020; Westley,
30 2020). Changes in temperature affect freshwater fishes, with increased summer temperatures in
31 river systems leading to increased mortality in fishes, including migratory populations
32 (Bowerman et al., 2021; Mauger et al., 2017; Morash et al., 2020; Westley, 2020) and juvenile
33 resident populations that are unable to escape unfavourable temperatures (Cairns et al., 2005;
34 Morash et al., 2020; Petty et al., 2012; Rodnick et al., 2008; White & Wagner, 2021).
35 Understanding a species' or population's temperature preference and upper thermal thresholds is
36 important for determining how increasing water temperatures will affect wild fish populations.

37 There are several common methods for estimating the thermal tolerance limits in fishes
38 that assess endpoints at different levels of biological organization. The critical thermal maximum
39 (CT_{max}) is a commonly used non-lethal estimate of its acute upper thermal limits for a species or
40 population, while the upper incipient lethal temperature (UILT) estimates a longer term upper
41 thermal threshold (Beiting et al., 2000). Acclimation to differing temperature regimes can
42 affect the thermal tolerance of fishes, with previous studies showing differences in CT_{max} within
43 the same population of fish acclimated to different temperatures (Kelly et al., 2014; Morrison et
44 al., 2020). A species' preferred temperature often coincides with their optimal growth and
45 metabolism (Macnaughton et al., 2021; Schulte et al., 2011), and these physiological parameters
46 can often be found to correspond with increased activities and behaviours, such as feeding,
47 reproduction, or avoidance responses (Crawshaw, 1977; Killen et al., 2013). Temperatures
48 around a species' preferred temperature generally provide the greatest opportunity for
49 energetically expensive activities (e.g., growth and reproduction), while temperatures towards
50 the upper and lower ends of a species' thermal range tend to reduce or prevent them (Crawshaw,
51 1977, 1984; Desforges et al., 2023). Research assessing the CT_{max} of fishes has shown an
52 activation of a behavioural response at temperatures leading up to CT_{max} , which has been
53 described as the agitation temperature (McDonnell & Chapman, 2015). Additionally, the CT_{max} -
54 agitation window, which represents the difference between the temperatures for these endpoints
55 can be used to assess how early a species' avoidance response is activated before critical
56 temperatures are reached (Wells et al., 2016). Studies that assess thermal endpoints can provide a
57 general understanding of when fish display a behavioural reaction to a thermal threshold,

58 however, such studies do not provide insight into sub-organismal responses of the fish. Despite
59 the potential utility of identifying behavioural avoidance responses to thermal stress in fishes,
60 relatively little is known about the physiological mechanisms that coincide with the agitation
61 temperature in fishes (Bouyoucos et al. 2023).

62 Measuring how temperature affects performance traits can be accomplished using
63 thermal performance curves (TPC). Performance traits, defined as biological processes that occur
64 over a period of time (e.g., growth, reproduction, metabolic rate; Schulte et al., 2011), often peak
65 at an optimal temperature and decrease as the temperature moves further away from this
66 optimum. Estimates of the amount of energy available to an organism, both for basic biological
67 functions (e.g., digestion and growth) or energy demanding life processes (e.g., reproduction),
68 can be drawn from the determination of aerobic scope (AS) (Fry, 1947). Metabolic rates of fishes
69 are often estimated using oxygen consumption over a set period of time, with standard metabolic
70 rate (SMR) representing the lowest oxygen consumption or minimum energy requirements to
71 sustain physiological function (i.e., homeostasis; Beamish & Mookherjee, 1964; Brett & Groves,
72 1979; Fry, 1971), maximum metabolic rate (MMR) being the highest rate of oxygen
73 consumption, often measured following exhaustive exercise (Brett & Groves, 1979; Treberg et
74 al., 2016), and AS signifying the difference between SMR and MMR (Fry, 1947). Requirements
75 and availability of energy for fish changes with water temperature, and an optimal temperature
76 for aerobic performance is often considered to be where AS peaks. It is important to note that a
77 defined AS peak is not always present in fishes with some species able to maintain AS over a
78 broad range of temperatures (e.g., Brook Trout (*Salvelinus fontinalis*); Durhack et al., 2021),
79 possibly right up to lethal temperatures (Clark et al., 2013; Munday et al., 2009). However, at
80 temperatures above peak AS, cardiac arrhythmia, and increased reliance on anaerobic
81 metabolism have been shown to occur in salmonid species (Anttila et al., 2013) highlighting the
82 importance of investigating responses to temperature across different levels of biological
83 organization. Combining whole-body metabolic, behavioural, and cellular responses to changes
84 in water temperature provides a full picture of the integrated organismal response to different
85 temperatures, from their preferred temperature to thermal extremes.

86 When an organism experiences changes in their thermal environment, the activation of
87 the heat shock response (HSR) via up- or downregulation of heat shock proteins (HSPs) occurs

88 (Tomanek, 2010). However, acclimation to a temperature within the thermal limits of a fish can
89 change when the HSR is induced. Within the thermal distribution of a species, moderate
90 temperature increases may be stimulatory, with a potential benefit to the organism (Jeffries et al.,
91 2016, 2018; Schreck, 2010). As temperatures increase there will be a sub-lethal threshold where
92 a thermal stress response is activated, which may lead to a reduction in organismal fitness
93 (Buckley et al., 2006; Komoroske et al., 2015; Logan & Somero, 2011; Schulte, 2014). At
94 extreme temperatures approaching CT_{max} an individual will start exhibiting a HSR (Jeffries et al.,
95 2018). Exposure to these extreme temperatures is where cellular level damage may take place
96 leading to apoptosis and there is an up- and downregulation of genes associated with cellular
97 survival mechanisms (Komoroske et al., 2015; Logan & Somero, 2011). Consequently, recovery
98 from exposure to extreme temperatures may be delayed or no longer possible. Survival at acute
99 exposures to temperatures above the onset of a thermal stress response is possible, however,
100 prolonged exposure to temperatures on the high end of a species' TPC may have impacts on
101 fitness.

102 The aim of this study was to assess whether the behavioural and whole-organism
103 responses relate to shifts in the cellular response to acute temperature increases. Brook Trout are
104 a eurythermal species that are widely distributed throughout North America. Upper thermal
105 tolerance and temperature preference studies have been previously conducted on other
106 populations of this species and temperature preference, peak AS, UILT, and CT_{max} have been
107 estimated to be 15.9 ± 1.4 °C, 15 °C, 25.1 ± 0.2 °C, and 29.9 ± 0.6 °C respectively (values from:
108 Durhack et al., 2021; Hasnain, 2012; Macnaughton et al., 2021; Morrison et al., 2020; reviewed
109 by Smith & Ridgway, 2019). We first conducted intermittent respirometry, shuttle box, and
110 CT_{max} experiments to identify key behavioural and physiological endpoints commonly used to
111 study temperatures impacts on fishes. Following a CT_{max} experimental protocol, we acutely
112 exposed a different subset of fish from the same cohort to different target temperatures ranging
113 from the preferred temperature to the CT_{max} . Gill, liver, and blood samples were then taken from
114 fish to quantify the changes in mRNA transcript levels and plasma indices. The combination of
115 experiments to assess responses to changes in temperature at a whole body, cellular, and
116 behavioural level tested the hypothesis that there is a sub-lethal threshold at a temperature
117 approaching the acute upper thermal limit (i.e., CT_{max}) characterized by both behavioural and

118 cellular responses. Defining critical physiological thresholds could be used as an upper
119 temperature limit for designating suitable habitat for a species in the wild.

120

121 **Methods**

122 *Animal care and holding*

123 Young-of-the-year Brook Trout were obtained from the Whiteshell Fish Hatchery in
124 southeastern Manitoba, Canada. The strain originated from Lake Nipigon, Ontario, Canada and
125 was brought to the hatchery in the summer of 2019 (P1), at which point a batch was obtained and
126 held at Fisheries and Oceans Canada's Freshwater Institute fish holding facility in 600 L tanks
127 that were maintained on a flow-through of de-chlorinated City of Winnipeg tap water and with
128 independent aeration. Fish were maintained under a 12-12 photoperiod (65 min of dawn/dusk,
129 full light starting at 07:05, and full dark at 19:05) at $10^{\circ}\text{C} \pm 1^{\circ}\text{C}$. Fish were fed daily on a diet
130 of commercial food pellets (EWOS Pacific: Complete Fish Feed for Salmonids, Cargill,
131 Minneapolis, MN, USA). All procedures were conducted under Animal Use Protocols approved
132 by the Fisheries and Oceans Canada Freshwater Institute Animal Care Committee (FWI-ACC
133 AUP-2020-05 & FWI-ACC AUP-2020-07) under the standards set by the Canadian Council for
134 Animal Care.

135 *Experimental Treatments*

136 *Temperature preference experiments*

137 A total of $n = 15$ fish were tested for temperature preference using the Shuttle Box system
138 (Loligo[®] Systems, Viborg, Denmark). Fish were haphazardly selected from the general
139 population tank after feeding and transferred into the Shuttle Box for 48 h. Fish were allowed to
140 acclimate to the shuttle box tank for ~ 15 min before the system was turned on and movement
141 tracking started. The temperature in the system started at 10°C when the fish was put in the tank
142 and was maintained between a maximum of 25°C and a minimum of 5°C for the safety of the
143 fish, as these temperatures are within the temperature tolerance of Brook Trout (reviewed by
144 Smith & Ridgway, 2019). Fish were allowed to freely swim back and forth between the
145 “heating” and “cooling” chambers on either side of the tank, and temperature difference between

146 the “heating” and “cooling” chambers was kept within 2 °C of each other, as well as having a
147 maximum temperature change of 3 °C per hour to avoid heat stressing the fish. Room lighting
148 was kept on a 12-12 cycle, the same as in the general population tank. During low light/overnight
149 times, infrared-light was used to illuminate the tank from below to allow the camera to track the
150 fish. Infrared lights were on a timer to turn on at 18:30 and turn off at 07:30, which coincided
151 with dawn and dusk periods of the room lighting cycle. It was noted that for ~15 min at these
152 times the camera was unable to track the fish, however, tracking was successful during the rest of
153 the day/night and no significant changes in temperature occurred during these brief windows.
154 Fish movement was monitored by an overhead camera and water temperature was recorded
155 using a TMP-REG instrument (Loligo® Systems, Viborg, Denmark) constantly for 48 h by the
156 Shuttlesoft software (Loligo® Systems, Viborg, Denmark). The software then automatically
157 calculated a preferred temperature based on the water temperature that the fish sought out. The
158 first 24 h that the fish was in the Shuttle Box was considered an acclimation period, and the data
159 from this time was excluded from final analysis of temperature preference.

160 *Intermittent Respirometry experiments*

161 From December 14, 2020 – March 17, 2021, we conducted intermittent respirometry
162 experiments on $n = 65$ fish to estimate metabolic rates (SMR, MMR and AS). Groups of $n = 15$
163 fish were acclimated for a minimum of three weeks at 5, 10, 15, 20 and 23 ± 0.1 °C. Intermittent
164 respirometry was conducted using AutoResp software to monitor and record oxygen and
165 temperature levels via Witrox4 instruments and control water pumps via a DAQ-M instrument
166 (Loligo® Systems, Viborg, Denmark). Fish were fasted for a minimum of 24 h prior to
167 experimentation. Following fasting, one fish at a time was haphazardly netted from the general
168 population tanks and first tested for MMR using an exhaustive chase protocol. The exhaustive
169 chase protocol entailed fish being coaxed to swim against a constant current of water until they
170 were deemed exhausted. Exhaustion was determined to be once a fish no longer responded to a
171 gentle pinch of the caudal fin (Durhack et al., 2021; Reidy et al., 1995). Fish were then
172 immediately transferred to acrylic respirometry chambers (volume 679 mL; Loligo® Systems,
173 Viborg, Denmark) where three measurements were recorded (measurement cycle = Measure –
174 180 s, Flush – 300 s, Wait – 40 s) to estimate MMR. The flush and wait periods were skipped for
175 the first measurement to ensure no recovery period was allowed for the fish before measurements

176 began. Following MMR estimates, $\square\text{O}_2$ measurements were recorded for 24 h to be used for
177 SMR estimates. Following SMR, fish were sacrificed as described below, measured for fork and
178 total length and sex determination.

179 *CT_{max} experiments*

180 Treatments of $n = 16$ fish were used in this study (total of $n = 208$). Six treatments (handling,
181 acclimation, agitation, CT_{max} , B1 and B2) were separated into two time points: T^0 – where fish
182 were sampled immediately once either a certain temperature or behaviour was reached, or T^1 –
183 where the fish were given a 1 h recovery period in a recovery bath held at 10 \square once the
184 temperature or behaviour was reached (Figure 1). A group of $n = 16$ fish was also used as a
185 control group to assess baseline stress levels. This group will here forth be termed the “baseline”
186 treatment. Fish were tested in groups of 4 in one of two 200 L green sampling tanks, each in an
187 individual acrylic chamber (volume – 1380 mL) with netting over the ends to contain and track
188 the fish (Figure 2). For treatments other than the baseline group, fish were haphazardly netted
189 from the general population tank and placed in their sampling chambers to fast and acclimate
190 overnight prior to experimental trials (Approx. 18 hours). Temperature of the tank was
191 monitored using a Witrox 4 instrument and AutoResp software (Loligo® Systems, Viborg,
192 Denmark). Both sampling tanks had air stones to maintain dissolved oxygen levels and two 5
193 L·min⁻¹ pumps (Eheim, Deizisau, Germany) to ensure homogenous temperature throughout the
194 tank. For treatments where heating was needed, a rate of approximately 0.3 °C min⁻¹ was used
195 with four 300 W heating elements (Finnex TH-0300S titanium heaters, Finnex, Chicago, USA).
196 Fish were removed from the treatment and either immediately sacrificed (T^0) or moved to the
197 recovery bath (T^1) when they either reached the specified temperature or exhibited the specified
198 behaviour. For treatments that ended with a specified temperature, fish were removed two at a
199 time, with one fish designated T^0 for sampling and one fish designated for T^1 and placed into the
200 recovery bath. The remaining two fish were held at stable temperature (± 0.2 °C) for 2 min to
201 allow time for tissue sampling from the first fish. In treatments with a behavioral endpoint, fish
202 were removed once the behaviour was noted and sampled in the same order as above. Behavioral
203 endpoint fish were not necessarily spaced out as well as temperature endpoint fish since
204 sampling depended on when the individual fish reached the endpoint (Figure 1). All fish were
205 sacrificed in Syncaine (MS-222; concentration: 450 mg L⁻¹ buffered with 900 mg L⁻¹ sodium

206 bicarbonate; Syndel Canada, Nanaimo, BC, Canada) for 3 min followed by cranial percussion.
207 Oxygen saturation was measured at the beginning, middle, and end of each trial and never fell
208 below 98%.

209 *Description of CT_{max} treatment groups*

210 The control groups are described first. The fish for the baseline treatment were haphazardly
211 netted out of the general population tanks and immediately euthanized for sampling as described
212 above. The baseline treatment fish were used to assess baseline stress levels before fish are
213 introduced to the CT_{max} chambers. Another control group, termed the “handling” control group,
214 was treated the same as all experimental treatments, but instead of the water being heated, the
215 tank was kept at a temperature of 10 °C for one hour (~ length of time for the CT_{max} treatment).
216 T⁰ fish were sampled to assess stress levels from experimental housing without temperature
217 effects, while T¹ fish represent residual stress levels from handling in fish following 1 h of
218 recovery.

219 The preferred temperature treatment for this population of fish was established with the
220 shuttle box experiments described above. The temperature preference experiments determined
221 the average preferred temperature to be 15.1 ± 1.13 °C. As such, 15.1 °C was used as the target
222 temperature for this treatment.

223 For the agitation treatment, we defined the agitation temperature as the temperature at
224 which sustained (>5 s) ‘curling’ and ‘bursting’ behaviours were displayed by the fish (Video 1).
225 The curling behaviour was defined as the fish’s body curved in a C-shape and was a result of
226 attempting to turn around in the sampling chambers to look for an escape from the water
227 temperature. Bursting behaviour was defined as bursts of energetic swimming by the fish and
228 was often observed as sustained pushing against the netting enclosing the ends of the chamber.
229 Again, this was likely an attempt to escape increasing water temperatures. We chose these two
230 behaviours because we sought to study a flight response possibly reflected in physiological
231 changes in a stressed fish, as opposed to a less extreme preference response with less drastic
232 movements (McDonnell & Chapman, 2015). Preliminary assessment of 16 fish from CT_{max} trials
233 led us to define the average agitation temperature for this treatment as 22.5 °C. The average
234 agitation temperature was used to define temperatures for treatments B1 and B2.

235 The B1 treatment was defined as the midpoint between the preferred temperature
236 treatment – 15.1 °C – and the mean agitation temperature treatment – 22.5 °C. As such, fish in
237 the B1 treatment were sampled at 18.8 °C.

238 As with the B1 treatment, the B2 treatment was defined as a group between the mean
239 agitation temperature and the average CT_{max} temperature. The average CT_{max} temperature was
240 estimated from the first $n = 16$ fish to undergo a CT_{max} trial and was found to be 28.1 °C. The B2
241 treatment temperature was thus set at 25.3 °C, halfway between the mean agitation temperature –
242 22.5 °C – and the average CT_{max} – 28.1 °C.

243 Similar to the agitation temperature treatment, we sampled fish in the CT_{max} treatment at
244 their individual CT_{max} thresholds. Fish were deemed to have reached CT_{max} once they were
245 unable to maintain equilibrium and ceased attempts to right themselves. In each trial of four, the
246 first fish to reach CT_{max} was sampled at T^0 , the second went to recovery for T^1 , the third sampled
247 for T^0 , and the fourth to recovery for T^1 so that neither the T^0 or T^1 sampling groups would be
248 biased for fish with low or high CT_{max} values.

249 *Tissue Sampling*

250 Following euthanasia, fish length and mass were recorded. Blood was then collected using
251 ammonium-heparinized capillary tubes (Fisherbrand®, Fisher Scientific, Pittsburgh,
252 Pennsylvania, USA) following severing of the caudal fin. Gill tissue from the ventral side of the
253 second gill arch, cut to the base of the gill filaments was taken for each fish. Liver was excised,
254 weighed, and the distal end of the lobe was taken. Each tissue was immediately stored in
255 RNAlater™ (Invitrogen™, Carlsbad, California, USA) kept on ice, and stored at 4 °C overnight
256 prior to storage at -80 °C following RNAlater™ best practices (Life Technologies, 2011).
257 Following this, the sex of the animal was determined visually (male, female, or indeterminant).

258 *Quantitative PCR*

259 Total RNA was extracted from the gill and liver tissues using a Qiagen RNeasy Plus Mini Kit
260 (Qiagen, Toronto, ON, CA) following manufacturer's protocols. The RNA samples were
261 checked for purity (A260/A280, A260/A230) and concentration using a NanoDrop One
262 Spectrophotometer (Thermo Fisher Scientific, Wilmington, DE, USA). The integrity of the RNA
263 was assessed by electrophoresis on a 1% agarose gel. 1 µg of total RNA was reverse transcribed

264 into cDNA using the Invitrogen™ SuperScript™ IV First-Strand Synthesis System (Thermo Fisher
265 Scientific, Waltham, MA, USA) following the manufacturer's protocols.

266 Forward and reverse quantitative PCR (qPCR) primers and probes (*cat*, *cirbp*, *efl**a*, *fos*,
267 *gpx1a*, *hsf1*, *hsp70a*, *hsp90aa*, *hsp90ba*, *ier2*, *jun*, *junb*, *jund*, *mycb*, *rpl7*, *rpl13a*, *rps9*, *sod1*,
268 *sod2*; Table 1) were designed in Geneious Prime software version 2021.2.2 (Biomatters Ltd,
269 Auckland, New Zealand) based off multiple salmonid species using sequences from GenBank®
270 and PhyloFish (Sayers et al., 2019; Sutherland et al., 2019). The remaining primers and probes
271 (*casp9*, *cs*, *ldh*, *hsp90ab1*; Supplementary Table 1) were designed by Jourdain-Bonneau et al.
272 (2023) using Primer Express software version 3.0 (Applied Biosystems, ThermoFisher
273 Scientific, Wilmington, DE, USA).

274 Primers and probes were designed for 18 target genes (Supplementary Table 1) that
275 represented transcriptomic responses consistent with high temperature (*cold-inducible RNA-*
276 *binding protein (cirbp)*, *heat shock transcription factor-1 (hsf1)*, *heat shock protein 70-alpha*
277 (*hsp70a*), *cytosolic heat shock protein 90-alpha (hsp90aa)*, *heat shock protein 90-beta-1*
278 (*hsp90ab1*), *heat shock protein 90-beta-alpha (hsp90ba)*), cell cycle and transcription (*caspase-9*
279 (*casp9*), *catalase (cat)*, *citrate synthase (cs)*, *protein c-fos (fos)*, *immediate-early response gene 2*
280 (*ier2*), *transcription factor ap-1 (jun)*, *transcription factor jun-b (junb)*, *transcription factor jun-*
281 *d (jund)*, *transcriptional regulator myc-2 (mycb)*), and general cellular function (*glutathione*
282 *peroxidase 1a (gpx1a)*, *lactate dehydrogenase (ldh)*). Primers were designed for four reference
283 genes including *elongation factor 1-alpha (efl**a**)*, *60s ribosomal protein L13a* and *L7 (rpl13a*
284 and *rpl7*), and *40s ribosomal protein S9 (rps9)* (Supplementary Table 1). Primer and probe
285 efficiencies were tested by generating standard curves using cDNA synthesized from the RNA
286 pooled from 7 individuals from the treatment groups. Each 12 µL qPCR reaction consisted of 1
287 µL of a 1:10 dilution of cDNA, 500 nM forward and reverse primer, 6 µL of Applied
288 Biosystems™ Taqman FAST Advanced Master Mix (Thermo Fisher Scientific, Waltham, MA,
289 USA) 0.3 µL of 10 µM Taqman probe, and 0.34 µL RNase-free water. The qPCR reactions were
290 run on a QuantStudio 5 Real-Time PCR System (Thermo Fisher Scientific, Life Technologies
291 Corporation, Carlsbad, California USA) in 384 well plates.

292

293 **Analysis**

294 Temperature preference for each fish was determined by taking the final value of the recorded
295 temperature preference values from the Shuttlesoft software (Loligo[®] Systems, Viborg,
296 Denmark). Overall temperature preference was determined by taking the mean \pm S.D. of all fish
297 tested. We tested for differences in temperature preference of juvenile Brook Trout between
298 daytime and nighttime hours, as previous studies had seen differences in temperature preference
299 between daytime and nighttime using the same system for another salmonid (Westslope
300 Cutthroat Trout, *Oncorhynchus clarkii lewisi*; Macnaughton et al., 2018). No differences were
301 found, as such we included all 24 h of data to estimate our final temperature preference.

302 *Metabolic Rate*

303 Variation in fish body mass ranging from 27.8–171.1 g – was observed within the fish used in
304 the respirometry experiment. To account for the range of fish mass across treatments and the
305 known effect of mass on metabolic rate, whole body SMR, MMR, and AS estimates were mass
306 corrected to the mean mass of all fish in the study (79.08 g) using multivariate polynomial
307 predictive equations (similar to Durhack et al., 2021; Poletto et al., 2017). Sex was also
308 determined for these fish to allow for assessment of metabolic differences between male ($n = 43$)
309 and female fish ($n = 32$), respectively, and all fish were deemed to be immature. The inter-
310 individual variability within our sampled fish allowed for testing of possible differences and
311 interactions between the variables mass, sex and time to exhaustion and their possible effects on
312 metabolic rate estimates. Tukey's honest significant difference test (Tukey HSD) was used for
313 post-hoc testing on any significant variables found to identify differences in temperature, mass,
314 time to exhaustion and sex within and across treatments, with a $P < 0.05$ deemed significant.
315 Mass-corrected data are presented in mass specific values ($\text{mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$) for comparison to
316 previous studies. Statistical analysis was conducted using R and R Studio (RStudio version
317 1.3.1056, Posit Team, 2020; R version 4.0.2, R Core Team, 2020) with the following packages
318 ‘car’ (Fox & S., 2019), ‘caret’ (Kuhn, 2008), ‘dplyr’ (Wickham et al., 2020), ‘fishMO2’ (Chabot
319 et al., 2016), ‘MASS’ (Venables & Ripley, 2002), ‘multcomp’ (Hothorn et al., 2008), ‘MuMin’
320 (Barton, 2020), ‘plotrix’ (J, 2006), and ‘tidyverse’ (Wickham et al., 2019).

321 *CT_{max}*

322 *CT_{max}* and agitation temperatures were determined by taking the mean \pm standard deviation
323 values of all treatments tested at least up that temperature. Estimation of *CT_{max}* was only taken

324 from fish fully heated to CT_{max} , which included our original 16 fish used to define the other
325 treatment temperatures and the CT_{max} treatment. As such, $n = 32$ fish were used to estimate
326 CT_{max} . Agitation temperature was assessed using all fish from the CT_{max} , Agitation and B₂
327 treatments, resulting in $n = 96$ fish. The agitation- CT_{max} window was determined as the
328 difference between the overall means of CT_{max} and agitation temperatures.

329 *Plasma Lactate*

330 A linear model was used to assess the extent to which plasma lactate concentration in nmol· μ L⁻¹
331 changed across experimental treatments and between the T⁰ and T¹ groups. The model consisted
332 of lactate concentration dependent on time in minutes from the start of experimental
333 manipulations to blood draw (sampling time), Fulton's condition factor, sex, and the interaction
334 of experimental treatment and recovery time. Model fit was assessed with the *check_model()*
335 function from the R package performance v0.8.0 (Lüdecke et al., 2021). Effect size statistics
336 were calculated with the *anova_stats()* function of the R package sjstats v0.18.1 (Lüdecke, 2018)
337 on a type II ANOVA performed on the linear model with the R package car v3.0-12 (Fox &
338 Weisberg, 2019). 95% confidence intervals given the interaction between experimental treatment
339 and recovery time were calculated with predicted marginal effects for each experimental
340 treatment except baseline with sjPlot v2.8.10 (Lüdecke, 2021). Fish in the baseline treatment
341 were omitted because the interaction was rank deficient since no T¹ fish were sampled for that
342 group.

343 *qPCR Data*

344 The R package MCMC.qpcr v1.2.4 with MCMCglmm v2.33 was used as a Bayesian approach
345 designed for qPCR data (Hadfield, 2010; Matz et al., 2013). Four reference genes were used in
346 all models: *rpl13a*, *rps9*, *rpl17*, and *ef1a*. Stability of the reference genes across treatments was
347 confirmed visually prior to including them in the analyses. Separated liver and gill models were
348 used because 20 genes total were analyzed in liver and 22 total in gill. Transcript abundance of
349 *jund*, *junb*, and *mycb*, were analyzed in the gill but not liver, while *hsp90aa* was analyzed in the
350 liver but not gill. Within each tissue, treatment and recovery conditions were treated as
351 interacting variables. The baseline treatment was omitted because the interaction was rank
352 deficient, as no fish from baseline were in the T¹ recovery group. The T⁰ and the acclimation
353 groups were used as references against which other treatments were compared.

354 In all models, sex, sampling time, and Fulton's condition factor were included as fixed
355 effects that may alter transcript abundance estimates or gene expression during sampling. Each
356 of the models was run with Markov Chain Monte Carlo settings of 110,000 iterations run total, a
357 burn-in period of 10,000 iterations discarded, and a thinning interval of parameters sampled once
358 every 100 iterations. Model convergence was assessed by observing trace plots and overall
359 diagnostic plots of residuals and predicted values, variance of the residuals and predicted values
360 (as a test of homoscedasticity), and normality of residuals. Differences in transcript abundance
361 between treatments were assessed with 95% credible intervals. The package tidybayes v3.0.2
362 was used to visualize posterior estimates of transcript abundance (Kay, 2021). Separate linear
363 models were used to assess the possibility that changes in mRNA abundance may reflect
364 sampling time (Supplemental Methods).

365

366 **Results**

367 *Intermittent Respirometry*

368 Both AS and MMR peaked between 10–15 °C (Figure 3-A & -B), with 10 and 15 °C values being
369 significantly different from all other treatment temperatures. However, no differences were
370 found between temperature treatments for either AS or MMR at 10 and 15 °C (Tukey, $p = 1.0$, p
371 = 0.95, respectively). The following data are presented as mean (\pm standard error) mass specific
372 corrected values. Mean AS at 10 and 15 °C was $396.30 \pm 18.00 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$ and $387.54 \pm$
373 $12.76 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$, respectively. At 23 °C, AS was significantly lower than any other treatment,
374 $138.93 \pm 42.70 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$. Mean mass specific MMR was $477.71 \pm 17.19 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$ at
375 10 °C and $498.66 \pm 13.01 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$ at 15 °C. Contrarily, SMR was found to increase with
376 acclimation temperature (Figure 3-B), with mean SMR at 5 °C found to be $31.022 \pm 5.034 \text{ mg}$
377 $\text{O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$ and at its highest at 23 °C, $174.23 \pm 25.64 \text{ mg O}^2 \cdot \text{kg}^{-1} \cdot \text{h}^{-1}$.

378 *Temperature Preference*

379 Mean temperature preference of juvenile Brook Trout (mean mass $50.97 \pm 14.0\text{g}$) estimated with
380 Shuttle box experiments was 15.10 ± 1.13 °C. No difference in temperature preference between
381 day (~7:30–18:30 h) and night were found (day: 15.09 ± 1.11 °C and night: 15.11 ± 1.17 °C,
382 respectively, $p = 0.48$).

383 CT_{max}

384 CT_{max} of juvenile Brook Trout held at 10 °C was 28.2 ± 0.4 °C. The mean \pm standard deviation
385 agitation temperature of fish tested in the agitation, B₂ and CT_{max} treatments was 22.0 ± 1.4 °C
386 and the mean agitation- CT_{max} window 6.2 °C.

387 *Cellular level responses*

388 *Plasma Lactate*

389 The linear model of lactate concentration dependent sampling time, Fulton's condition factor,
390 sex and the interaction of experimental treatment and recovery time was significant ($F = 19.77, p$
391 < 0.001 , adjusted $R^2 = 0.70$; Figure 4). There was reasonable normality among model residuals,
392 but variance tended to be lower at lower fitted values, and there was high collinearity between
393 sampling time, treatment, and recovery time. Nevertheless, experimental treatment had the
394 greatest effect size on lactate concentration (η squared = 0.43), while the interaction of treatment
395 and recovery time (η squared = 0.19) and recovery time on its own (η squared = 0.053) were
396 each smaller in effect size (Supplementary Table 2). Sampling time was not significant ($p =$
397 0.48) and its effect size was low (η squared = 0.001). Lactate concentrations were not
398 significantly different from baseline fish at any experimental treatment except at agitation- CT_{max}
399 and CT_{max} ($p = 0.05$ each), and the T¹ was associated with higher lactate concentrations than the
400 T⁰ ($p < 0.001$) (Supplementary Figure 1). None of sampling time, Fulton's condition factor, or
401 sex were significant in the ANOVA effect size statistics ($p > 0.05$ for each variable). 95%
402 confidence intervals showed higher plasma lactate in the CT_{max} T¹ group relative to all other
403 groups, including the CT_{max} T⁰ group (Figure 4). In every other experimental treatment in the
404 analysis, there was no significant difference between the T⁰ and T¹ groups.

405 *qPCR Data*

406 In the model of mRNA abundance in gill tissue, *fos*, *hsp70a*, *ier2*, *jun*, *jund*, and *jund* showed
407 pronounced changes both between the T⁰ and T¹ groups and as the experiment progressed from
408 acclimation to CT_{max} . In the model of mRNA abundance in liver tissue, *fos*, *hsp70a*, and *ier2*
409 showed similar pronounced changes between the groups and as the experiment progressed
410 toward CT_{max} .

411 Transcript abundance changed in both liver and gill for *hsp70a*, *ier2*, and *fos* (Figure 5).
412 In *ier2* and *fos*, responses between the T⁰ and T¹ groups were more substantial in the gill than the
413 liver. In gill tissue, *hsp70a* showed a marked increase in abundance between the agitation-*CT_{max}*
414 T⁰ and T¹ groups. This increase was consistent with a threshold effect that was both prior to
415 *CT_{max}* and subsequent to a recovery period following the agitation-*CT_{max}* temperature. A similar
416 response was observed at the agitation temperature, where in the gill, *hsp70a* and *fos* increased at
417 agitation but only after the 60-min recovery. Therefore, we observed a transcriptional response at
418 the same temperature as the behavioural response of agitation.

419 Transcript abundance of *jun*, *jurb* and *jund* in gill tissue showed an increasing gradient in
420 the T¹ recovery group in all three genes (Figure 6). The T⁰ groups showed no distinguishable
421 response except in *jun* in the gill tissue between the *CT_{max}* and acclimation recovery groups. In
422 the liver, *jun* showed no significant response between any of the T⁰ versus T¹ recovery groups.
423 An analysis of sampling time revealed a limited association with mRNA abundance
424 (Supplemental Results).

425

426 **Discussion**

427 The data from this study suggest that the agitation temperature is a useful endpoint for estimating
428 when temperatures may start causing damage to the fish. In the current study, juvenile Brook
429 Trout exhibited a behavioural response to temperature, declining AS, increases in plasma lactate
430 levels and increases in stress-related mRNA transcript abundance when acutely exposed to
431 temperatures around 22 °C. The estimated realized thermal niche of Brook Trout in the wild is
432 predicted to be 10–20 °C (reviewed in Smith & Ridgway, 2019) and they are thought to be
433 restricted by an upper temperature limit of ~24 °C in nature (MacCrimmon & Campbell, 1969;
434 Meisner, 1990; Ricker, 1934). Studies that estimated the onset of thermal avoidance and
435 physiological stress in Brook Trout estimate an upper temperature threshold of 21–23.5 °C
436 (Chadwick et al., 2015; Chadwick & McCormick, 2017; Goyer et al., 2014; Lund et al., 2003;
437 reviewed in Smith and Ridgeway, 2019). The agitation temperature of 22.0 ± 1.4 °C from this
438 study falls within this expected range, and the activation of several genes at the agitation
439 treatment temperature, suggests that a shift in the cellular response to temperature occurs at the
440 onset of thermal avoidance behaviour. Similar results were observed in Pacific spiny dogfish

441 (Squalus suckleyi) where they showed a cellular stress response occurred at the agitation
442 temperature (Bouyoucos et al., 2023). Morrison et al. (2020) showed that CT_{max} generally
443 increased with acclimation temperature in Brook Trout, however, CT_{max} plateaued at acclimation
444 temperatures above 20 °C that coincided with a decrease in hepatosomatic index, and an increase
445 in plasma lactate. The results by Morrison et al. (2020) agree with the 22 °C agitation
446 temperature and plasma lactate responses from this study despite the current study being at an
447 acute level of exposure. One caveat on the differences seen across treatments for the changes in
448 the mRNA transcripts in this study, is that the higher a temperature of a treatment, the longer a
449 fish was exposed to increasing water temperatures. Therefore, there is a chance some of the
450 responses seen may be a result of longer acute exposure to elevated temperature in the extreme
451 temperature treatments.

452 A neuroendocrine response is a likely precursor to the onset of behavioural and cellular
453 responses to high temperatures observed in this study. A neuroendocrine stress response may
454 lead to direct tissue stimulation by the nervous system or to an increase in catecholamines
455 released into general circulation when a stress threshold is hit, which may be followed by a
456 glucocorticoid response (reviewed by Fabbri et al., 1998; Molinoff & Axelrod, 1971; Reid et al.,
457 1998). A Rainbow Trout (*Onchorhynchus mykiss*) strain selected for a high cortisol response
458 showed an increase in circulating plasma catecholamines after an acute thermal stress that was
459 approximately 3–4 °C lower than the CT_{max} for that strain (Leblanc et al., 2012). An increase in
460 circulating levels of catecholamines at a temperature threshold may lead to increases in heart rate
461 that might be expected to be associated with increased activity in the fish. Indeed, intraperitoneal
462 injections of pharmacological agents to stimulate adrenergic receptors results in an increase in
463 heart rate in coho salmon (*Oncorhynchus kisutch*; Casselman et al., 2012). Further, the point
464 during acute temperature increases leading to cardiac arrhythmia occurs approximately 2–3 °C
465 lower than the CT_{max} in Arctic Char (*Salvelinus alpinus*; Gilbert & Farrell, 2021). These studies
466 suggest that the effects of temperature on heart rate is under neuroendocrine control, which may
467 be consistent with the onset of avoidance behaviours at the agitation temperature. The onset of
468 the agitation response may also be linked to the start of neuronal dysfunction in the fish, where a
469 behavioural avoidance response may help avoid neural impairment of the locomotor functions at
470 higher water temperatures (Andreassen et al., 2022). Neuron activity related to increasing water
471 temperatures has been described in other fish species (i.e., Transient Receptor Potential cation

472 channel (TRPV1) in Zebrafish (*Danio rerio*); Gau et al., 2013), which may also be related to an
473 activation of neuroendocrine response that is a precursor to an avoidance response at a relevant
474 thermal limit (i.e., the agitation temperature). These neuronal responses to water temperature
475 may be a mechanistic link between the behavioural and cellular level responses seen in this
476 study.

477 We observed a delay in the cellular response in the agitation treatment where there was
478 significantly altered mRNA transcript levels for *hsp70a* and stress inducible transcription factors
479 after 60 min of recovery, but not at 0 min when the temperature was reached. This is in contrast
480 to the higher temperature treatments that showed altered mRNA transcript levels when the
481 temperature was reached (i.e., 0 min) and after 60 min of recovery. The observation of a delay in
482 the cellular response may be consistent with the activation of a neuroendocrine response around
483 the agitation temperature for these fish. If there was an increase in circulating cortisol, there may
484 be a delay in when it binds to glucocorticoid receptors in target tissues leading to a cellular
485 response that is staggered from the more rapid catecholamine response. This is consistent with
486 previous work on pink salmon (*Oncorhynchus gorbuscha*) and sockeye salmon (*Oncorhynchus*
487 *nerka*) that showed a peak in mRNA transcript abundance occurring after the peak in blood
488 circulating cortisol after a handling stressor (Donaldson et al., 2014). Interestingly, the cortisol
489 response has been shown to dampen the HSP response in fishes (Basu et al., 2001). A delay in
490 the HSP70 production is consistent with a peak in circulating cortisol levels in rainbow trout
491 when temperatures reached 25 °C (3–4 °C lower than the CT_{max}) during acute warming from 13
492 °C over 1.5 h, however, peak HSP70 protein levels in the red blood cells were detected after 8 h
493 of recovery (LeBlanc et al. 2012). Collectively, these results and the results from the present
494 study suggest that there may be a delay between the initiation of a neuroendocrine response and
495 peak cellular transcriptome response during an acute temperature stress, which is consistent with
496 the cellular response being detected only after 60 min of recovery after the Brook Trout were
497 acutely exposed to their agitation temperature.

498 An increase in plasma lactate turnover rates at a sub-lethal temperature threshold may
499 explain the decrease in plasma lactate observed in the agitation temperature 60 min samples.
500 Mackey et al. (2021) observed a 70% decrease in white muscle lactate below baseline levels in
501 Brook Trout 24 h following exhaustive exercise at 23 °C, suggesting an increase in lactate

502 clearance at 23 °C. Results from this study were similar, with plasma lactate levels in the 22 °C
503 60 min recovery treatment below baseline treatment values. Optimal clearance of plasma lactate
504 may occur around 22 °C (i.e., agitation temperature) in this study due to a combination of an
505 increase in plasma catecholamine related to the thermal stress response (reviewed by Fabbri et
506 al., 1998; Molinoff & Axelrod, 1971; Reid et al., 1998), along with *in situ* glycogenesis in the
507 white muscle, which accounts for 80–85% of the lactate produced (Milligan, 1996; reviewed by
508 Warren & Jackson, 2008). Above the agitation temperature, lactate production due to tissue
509 oxygen demands (Gilbert & Farrell, 2021; Leblanc et al., 2012) may exceed clearance rates,
510 causing an accumulation within the blood and muscle tissue.

511 The agitation temperature appears to be a useful threshold for assessing thermal ecological
512 limits, however, other physiological factors will likely impact fish populations over chronic time
513 scales before this temperature is reached. Chadwick and McCormick (2017) observed that
514 growth rates begin decreasing above 16 °C and negative growth at temperatures ~23.4 °C.
515 Interestingly, Chadwick and McCormick (2017) also found 22 °C to coincide with drastic
516 increases in plasma cortisol and HSP70 levels in the gill, 12- and 11-fold higher than at 16 °C,
517 respectively. Mackey et al. (2021) also detected physiological responses such as increased
518 plasma cortisol, glucose, and muscle lactate levels in fish exposed to temperatures above 20 °C.
519 The Brook Trout used in this study appeared to have a relatively large agitation- CT_{max} window
520 (6.2 °C), which could afford them time to seek thermal refuge from extreme temperatures, unlike
521 other salmonids such as the Westslope Cutthroat Trout (*Oncorhynchus clarkii lewisi*), whose
522 agitation- CT_{max} window is smaller (1.8 °C ; Enders & Durhack, 2022). The activation of a
523 neuroendocrine and behavioural response, combined with a major shift in the transcriptome
524 response occurred near the agitation temperature, and may be a potential link between the
525 cellular stress response and the behavioural avoidance response in the fish.

526

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550

551 **Data Availability:** Data can be made available upon request.

552

553 **References**

554 Andreassen, A. H., Hall, P., Khatibzadeh, P., Jutfelt, F., & Kermen, F. (2022). Brain dysfunction
555 during warming is linked to oxygen limitation in larval zebrafish. *Proceedings of the*
556 *National Academy of Sciences of the United States of America*, 119(39), 1–10.
557 <https://doi.org/10.1073/pnas.2207052119>

558 Anttila, K., Casselman, M. T., Schulte, P. M., & Farrell, A. P. (2013). Optimum Temperature in
559 Juvenile Salmonids: Connecting Subcellular Indicators to Tissue Function and Whole-
560 Organism Thermal Optimum. *Physiological and Biochemical Zoology*, 86(2), 245–256.
561 <https://doi.org/10.1086/669265>

562 Barton, K. (2020). *MuMIn: Multi-Model Inference* (1.43.17). <https://cran.r-project.org/web/packages/MuMIn/MuMIn.pdf>

563

564 Basu, N., Nakano, T., Grau, E. G., & Iwama, G. K. (2001). The Effects of Cortisol on Heat
565 Shock Protein 70 Levels in Two Fish Species. *General and Comparative Endocrinology*,
566 124(1), 97–105. <https://doi.org/10.1006/gcen.2001.7688>

567 Beamish, F. W. H., & Mookherjee, P. S. (1964). Respiration of Fishes with Special Emphasis on
568 Standard Oxygen Consumption: I. Influence of Weight and Temperature on Respiration of
569 Goldfish, *Carassius Auratus* L. *Canadian Journal of Zoology*, 42(2), 161–175.
570 <https://doi.org/10.1139/z06-901>

571 Beitinger, T., Bennett, W., & McCauley, R. (2000). Temperature tolerances of North American
572 freshwater fishes exposed to dynamic changes in temperature. *Environmental Biology of*
573 *Fishes*, 58, 237–275. <https://doi.org/10.1023/A:1007676325825>

574 Bouyoucos, I. A., Weinrauch, A. M., Jeffries, K. M., & Anderson, W. G. (2023). Physiological
575 responses to acute warming at the agitation temperature in a temperate shark. *Journal of*
576 *Experimental Biology*. <https://doi.org/10.1242/jeb.246304>

577 Bowerman, T. E., Keefer, M. L., & Caudill, C. C. (2021). Elevated stream temperature, origin,
578 and individual size influence Chinook salmon prespawn mortality across the Columbia
579 River Basin. *Fisheries Research*, 237(January), 105874.
580 <https://doi.org/10.1016/j.fishres.2021.105874>

581 Brett, J. R., & Groves, T. D. D. (1979). Physiological energetics. In *Fish Physiology* (pp. 280–
582 352).

583 Buckley, B. A., Gracey, A. Y., & Somero, G. N. (2006). The cellular response to heat stress in
584 the goby *Gillichthys mirabilis*: A cDNA microarray and protein-level analysis. *Journal of*
585 *Experimental Biology*, 209(14), 2660–2677. <https://doi.org/10.1242/jeb.02292>

586 Cairns, M. A., Ebersole, J. L., Baker, J. P., Wigington, P. J., Lavigne, H. R., & Davis, S. M.
587 (2005). Influence of Summer Stream Temperatures on Black Spot Infestation of Juvenile
588 Coho Salmon in the Oregon Coast Range. *Transactions of the American Fisheries Society*,
589 134(6), 1471–1479. <https://doi.org/10.1577/t04-151.1>

590 Casselman, M. T., Anttila, K., & Farrell, A. P. (2012). Using maximum heart rate as a rapid
591 screening tool to determine optimum temperature for aerobic scope in Pacific salmon
592 *Oncorhynchus* spp. *Journal of Fish Biology*, 80(2), 358–377.

593 <https://doi.org/10.1111/j.1095-8649.2011.03182.x>

594 Chabot, D., Steffensen, J. F., & Farrell, A. P. (2016). The determination of standard metabolic
595 rate in fishes. *Journal of Fish Biology*, 88(1), 81–121. <https://doi.org/10.1111/jfb.12845>

596 Chadwick, J. G., & McCormick, S. D. (2017). Upper thermal limits of growth in brook trout and
597 their relationship to stress physiology. *Journal of Experimental Biology*, 220(21), 3976–
598 3987. <https://doi.org/10.1242/jeb.161224>

599 Chadwick, J. G., Nislow, K. H., & McCormick, S. D. (2015). Thermal onset of cellular and
600 endocrine stress responses correspond to ecological limits in brook trout, an iconic cold-
601 water fish. *Conservation Physiology*, 3(1), 1–12. <https://doi.org/10.1093/conphys/cov017>

602 Clark, T. D., Sandblom, E., & Jutfelt, F. (2013). Aerobic scope measurements of fishes in an era
603 of climate change: respirometry, relevance and recommendations. *Journal of Experimental
604 Biology*, 216(15), 2771–2782. <https://doi.org/10.1242/jeb.084251>

605 Crawshaw, L. I. (1977). Physiological and Behavioral Reactions of Fishes to Temperature
606 Change. *Journal of Fisheries Research Board of Canada*, 34, 730–734.
607 <http://www.nrcresearchpress.com/doi/pdf/10.1139/f77-113>

608 Crawshaw, L. I. (1984). Low-temperature dormancy in fish. *American Journal of Physiology -
609 Regulatory Integrative and Comparative Physiology*, 15(4), 479–486.
610 <https://doi.org/10.1152/ajpregu.1984.246.4.r479>

611 Desforges, J. E., Birnie-Gauvin, K., Jutfelt, F., Gilmour, K. M., Eliason, E. J., Dressler, T. L.,
612 McKenzie, D. J., Bates, A. E., Lawrence, M. J., Fangue, N., & Cooke, S. J. (2023). The
613 Ecological Relevance of Critical Thermal Maxima Methodology (CTM) for Fishes.
614 *Journal of Fish Biology*, February, 1–17. <https://doi.org/10.1111/jfb.15368>

615 Donaldson, M. R., Hinch, S. G., Jeffries, K. M., Patterson, D. A., Cooke, S. J., Farrell, A. P., &
616 Miller, K. M. (2014). Comparative Biochemistry and Physiology , Part A Species- and sex-
617 specific responses and recovery of wild , mature pacific salmon to an exhaustive exercise
618 and air exposure stressor. *Comparative Biochemistry and Physiology, Part A*, 173, 7–16.
619 <https://doi.org/10.1016/j.cbpa.2014.02.019>

620 Durhack, T. C., Mochnacz, N. J., Macnaughton, C. J., Enders, E. C., & Treberg, J. R. (2021).
621 Life through a wider scope: Brook Trout (*Salvelinus fontinalis*) exhibit similar aerobic
622 scope across a broad temperature range. *Journal of Thermal Biology*, 99(February), 102929.
623 <https://doi.org/10.1016/j.jtherbio.2021.102929>

624 Enders, E. C., & Durhack, T. C. (2022). Metabolic rate and critical thermal maximum CTmax
625 estimates for westslope cutthroat trout, *Oncorhynchus clarkii lewisi*. *Conservation
626 Physiology*, 10(1), 1–9. <https://doi.org/10.1093/conphys/coac071>

627 Fabbri, E., Capuzzo, A., & Moon, T. W. (1998). The role of circulating catecholamines in the
628 regulation of fish metabolism: An overview. *Comparative Biochemistry and Physiology
629 Part C: Pharmacology, Toxicology and Endocrinology*, 120(2), 177–192.

630 Fox, J., & Weisberg, S. (2019). *An R Companion to Applied Regression* (3rd).
631 <https://socialsciences.mcmaster.ca/jfox/Books/Companion/>

632 Fry, F. E. J. (1947). Effects of the Environment on Animal Activity. *Publications of the Ontario*
633 *Fisheries Research Laboratory*, 55(LXVIII), 1–62.

634 Fry, F. E. J. (1971). The effect of environmental factors on the physiology of fish. In W. Hoar &
635 D. Randall (Eds.), *Fish Physiology: Environmental relations and behavior* (pp. 1–98).
636 Academic Press.

637 Gau, P., Poon, J., Ufret-Vincenty, C., Snelson, C. D., Gordon, S. E., Raible, D. W., & Dhaka, A.
638 (2013). The zebrafish ortholog of TRPV1 is required for heat-induced locomotion. *Annals*
639 *of Internal Medicine*, 158(6), 5249–5260. <https://doi.org/10.1523/JNEUROSCI.5403-12.2013>

641 Gilbert, M. J. H., & Farrell, A. P. (2021). The thermal acclimation potential of maximum heart
642 rate and cardiac heat tolerance in arctic char (*Salvelinus alpinus*), a northern cold-water
643 specialist. *Journal of Thermal Biology*, 95(October 2020), 102816.
644 <https://doi.org/10.1016/j.jtherbio.2020.102816>

645 Goyer, K., Bertolo, A., Pépino, M., & Magnan, P. (2014). Effects of lake warming on
646 behavioural thermoregulatory tactics in a cold-water stenothermic fish. *PLoS ONE*, 9(3).
647 <https://doi.org/10.1371/journal.pone.0092514>

648 Hadfield, J. D. (2010). MCMCglmm: MCMC Methods for Multi-Response GLMMs in R.
649 *Journal of Statistical Software*, 33(2), 1–22. <https://www.jstatsoft.org/v33/i02/>

650 Hasnain, S. (2012). *Factors influencing ecological metrics of thermal response in North*
651 *American freshwater fish*.

652 Hothorn, T., Bretz, F., & Westfall, P. (2008). Simultaneous Inference in General Parametric
653 Models. *Biometrical Journal*, 50(3), 346–363.

654 J, L. (2006). Plotrix: a package in the red light district of R. *R-News*, 6(4), 8–12.

655 Jeffries, K. M., Connon, R. E., Davis, B. E., Komoroske, L. M., Britton, M. T., Sommer, T.,
656 Todgham, A. E., & Fangue, N. A. (2016). Effects of high temperatures on threatened
657 estuarine fishes during periods of extreme drought. *Journal of Experimental Biology*,
658 219(11), 1705–1716. <https://doi.org/10.1242/jeb.134528>

659 Jeffries, K. M., Fangue, N. A., & Connon, R. E. (2018). Multiple sub-lethal thresholds for
660 cellular responses to thermal stressors in an estuarine fish. *Comparative Biochemistry and*
661 *Physiology -Part A□: Molecular and Integrative Physiology*, 225(June), 33–45.
662 <https://doi.org/10.1016/j.cbpa.2018.06.020>

663 Jourdain-Bonneau, C., Deslauriers, D., Gourtay, C., Jeffries, K. M., & Audet, C. (2023).
664 Metabolic and transcriptomic response of two juvenile anadromous brook charr (. *Canadian*
665 *Journal of Zoology*.

666 Kay, M. (2021). *tidybayes: Tidy data and geoms for bayesian models* (3.0.2).
667 <https://doi.org/10.5281/zenodo.1308151>

668 Kelly, N. I., Burness, G., McDermid, J. L., & Wilson, C. C. (2014). Ice age fish in a warming
669 world: Minimal variation in thermal acclimation capacity among lake trout (*Salvelinus*
670 *namaycush*) populations. *Conservation Physiology*, 2(1), 1–14.

671 https://doi.org/10.1093/conphys/cou025

672 Killen, S. S., Marras, S., Metcalfe, N. B., McKenzie, D. J., & Domenici, P. (2013).
673 Environmental stressors alter relationships between physiology and behaviour. *Trends in*
674 *Ecology and Evolution*, 28(11), 651–658. <https://doi.org/10.1016/j.tree.2013.05.005>

675 Komoroske, L. M., Connell, R. E., Jeffries, K. M., & Fangue, N. A. (2015). Linking
676 transcriptional responses to organismal tolerance reveals mechanisms of thermal sensitivity
677 in a mesothermal endangered fish. *Molecular Ecology*, 24(19), 4960–4981.
678 <https://doi.org/10.1111/mec.13373>

679 Leblanc, S., Höglund, E., Gilmour, K. M., & Currie, S. (2012). Hormonal modulation of the heat
680 shock response: insights from fish with divergent cortisol stress responses. *American*
681 *Journal of Physiology - Regulatory Integrative and Comparative Physiology*, 302(1), R184–
682 R192. <https://doi.org/10.1152/ajpregu.00196.2011>

683 Life Technologies. (2011). *RNAlater Tissue Collection: RNA Stabilization Solution*.
684 https://tools.thermofisher.com/content/sfs/manuals/cms_056069.pdf

685 Logan, C. A., & Somero, G. N. (2011). Effects of thermal acclimation on transcriptional
686 responses to acute heat stress in the eurythermal fish *Gillichthys mirabilis* (Cooper).
687 *American Journal of Physiology - Regulatory Integrative and Comparative Physiology*,
688 300(6), 1373–1383. <https://doi.org/10.1152/ajpregu.00689.2010>

689 Lüdecke, D. (2018). *sjstats: Statistical functions for regression models* (0.18.1).
690 <https://doi.org/10.5281/zenodo.1284472>

691 Lüdecke, D. (2021). *sjPlot: Data visualization for statistics in social science* (2.8.10).
692 <https://cran.r-project.org/package=sjPlot>

693 Lüdecke, D., Ben-Shachar, M. S., Indrajeet, P., Waggoner, P., & Makowski, D. (2021).
694 performance: An R package for assessment, comparison and testing statistical models. *The*
695 *Journal of Open Source Software*, 6(60), 3139.
696 <https://doi.org/https://doi.org/10.21105/joss.03139>

697 Lund, E., Olsen, E. M., & Vøllestad, L. A. (2003). First-year survival of brown trout in three
698 Norwegian streams. *Journal of Fish Biology*, 62(2), 323–340.
699 <https://doi.org/10.1046/j.1095-8649.2003.00025.x>

700 MacCrimmon, H. R., & Campbell, S. J. (1969). World Distribution of Brook Trout, *Salvelinus*
701 *fontinalis*. *Journal of Fisheries Research Board of Canada*, 26(7), 1699–1725.

702 Mackey, T. E., Hasler, C. T., Durhack, T. C., Jeffrey, J. D., Macnaughton, C. J., Ta, K., Enders,
703 E. C., & Jeffries, K. M. (2021). Molecular and physiological responses predict acclimation
704 limits in juvenile brook trout (*Salvelinus fontinalis*). *Journal of Experimental Biology*,
705 224(16). <https://doi.org/10.1242/jeb.241885>

706 Macnaughton, C. J., Durhack, T. C., Mochnacz, N. J., & Enders, E. C. (2021). Metabolic
707 performance and thermal preference of westslope cutthroat trout (*Oncorhynchus clarkii*
708 *lewisii*) and non-native trout across an ecologically relevant range of temperatures.
709 *Canadian Journal of Fisheries and Aquatic Sciences*, 78, 1247–1256.
710 <https://doi.org/10.1139/cjfas-2020-0173>

711 Macnaughton, C. J., Kovachik, C., Charles, C., & Enders, E. C. (2018). Using the shuttlebox
712 experimental design to determine temperature preference for juvenile Westslope Cutthroat
713 Trout (*Oncorhynchus clarkii lewisi*). *Conservation Physiology*, 6(1), 1–10.
714 <https://doi.org/10.1093/conphys/coy018>

715 Matz, M. V., Wright, R. M., & Scott, J. G. (2013). No control genes required: Bayesian analysis
716 of qRT-PCR data. *PloS One*, 8(8), 1–12. <https://doi.org/10.1371/journal.pone.0071448>

717 Mauger, S., Shaftel, R., Leppla, J. C., & Rinella, D. J. (2017). Summer temperature regimes in
718 southcentral Alaska streams: Watershed drivers of variation and potential implications for
719 Pacific salmon. *Canadian Journal of Fisheries and Aquatic Sciences*, 74(5), 702–715.
720 <https://doi.org/10.1139/cjfas-2016-0076>

721 McDonnell, L. H., & Chapman, L. J. (2015). At the edge of the thermal window: Effects of
722 elevated temperature on the resting metabolism, hypoxia tolerance and upper critical
723 thermal limit of a widespread African cichlid. *Conservation Physiology*, 3(1), 1–13.
724 <https://doi.org/10.1093/conphys/cov050>

725 Meisner, J. D. (1990). Effect of Climatic Warming on the Southern Margins of the Native Range
726 of Brook Trout, *Salvelinus fontinalis*. *Canadian Journal of Fisheries & Aquatic Sciences*,
727 47, 1065–1070.

728 Milligan, C. L. (1996). Metabolic Recovery from Exhaustive Exercise in Rainbow Trout.
729 *Comparative Biochemistry and Physiology Part A: Physiology*, 113(1), 51–60.
730 [https://doi.org/https://doi.org/10.1016/0300-9629\(95\)02060-8](https://doi.org/https://doi.org/10.1016/0300-9629(95)02060-8)

731 Molinoff, P. B., & Axelrod, J. (1971). Biochemistry of catecholamines. *Annual Review of
732 Biochemistry*, 40(1), 465–500.

733 Morash, A. J., Speers-Roesch, B., Andrew, S., & Currie, S. (2020). The physiological ups and
734 downs of thermal variability in temperate freshwater ecosystems. *Journal of Fish Biology*.
735 <https://doi.org/10.1111/jfb.14655>

736 Morrison, S. M., Mackey, T. E., Durhack, T. C., Jeffrey, J. D., Wiens, L. M., Mochnacz, N. J.,
737 Hasler, C. T., Enders, E. C., Treberg, J. R., & Jeffries, K. M. (2020). Sub-lethal temperature
738 thresholds indicate acclimation and physiological limits in brook trout *Salvelinus fontinalis*.
739 *Journal of Fish Biology*, 97(2), 583–587. <https://doi.org/10.1111/jfb.14411>

740 Munday, P. L., Crawley, N. E., & Nilsson, G. E. (2009). Interacting effects of elevated
741 temperature and ocean acidification on the aerobic performance of coral reef fishes. *Marine
742 Ecology Progress Series*, 388, 235–242. <https://doi.org/10.3354/meps08137>

743 Petty, J. T., Hansbarger, J. L., Huntsman, B. M., & Mazik, P. M. (2012). Brook trout movement
744 in response to temperature, flow, and thermal refugia within a complex Appalachian
745 riverscape. *Transactions of the American Fisheries Society*, 141(4), 1060–1073.
746 <https://doi.org/10.1080/00028487.2012.681102>

747 Poletto, J. B., Cocherell, D. E., Baird, S. E., Nguyen, T. X., Cabrera-stagno, V., Farrell, A. P., &
748 Fangue, N. A. (2017). Unusual aerobic performance at high temperatures in juvenile
749 Chinook salmon, *Oncorhynchus tshawytscha*. *Conservation Physiology*, 5, 1–13.
750 <https://doi.org/10.1093/conphys/cow067>

751 Posit Team. (2020). *RStudio: Integrated Development Environment for R* (1.3.1056). Posit
752 Software, PBC. <http://posit.co/>

753 R Core Team. (2020). *R: A Language and Environment for Statistical Computing* (4.0.2). R
754 Foundation for Statistical Computing. <http://www.r-project.org>

755 Reid, S. G., Bernier, N. J., & Perry, S. F. (1998). *The adrenergic stress response in fish* □:
756 *control of catecholamine storage and release*. 120, 1–27.

757 Reidy, S. P., Nelson, J. A., Tang, Y., & Kerr, S. R. (1995). Post-exercise metabolic rate in
758 Atlantic cod and its dependence upon the method of exhaustion. *Journal of Fish Biology*,
759 47, 377–386.

760 Ricker, W. E. (1934). *An ecological classification of certain Ontario streams*. *Ontario Fisheries
761 Research Laboratory, Publication No 49, University of Toronto Studies, Biological Series,
762 No 37*. University of Toronto Press. Toronto, ON.

763 Rodnick, K. J., St.-Hilaire, S., Battiprolu, P. K., Seiler, S. M., Kent, M. L., Powell, M. S., &
764 Ebersole, J. L. (2008). Habitat Selection Influences Sex Distribution, Morphology, Tissue
765 Biochemistry, and Parasite Load of Juvenile Coho Salmon in the West Fork Smith River,
766 Oregon. *Transactions of the American Fisheries Society*, 137(6), 1571–1590.
767 <https://doi.org/10.1577/t07-138.1>

768 Schreck, C. B. (2010). Stress and fish reproduction: The roles of allostasis and hormesis.
769 *General and Comparative Endocrinology*, 165(3), 549–556.
770 <https://doi.org/10.1016/j.ygcen.2009.07.004>

771 Schulte, P. M. (2014). What is environmental stress? Insights from fish living in a variable
772 environment. *Journal of Experimental Biology*, 217, 23–34.
773 <https://doi.org/10.1242/jeb.089722>

774 Schulte, P. M., Healy, T. M., & Fangue, N. A. (2011). Thermal performance curves, phenotypic
775 plasticity, and the time scales of temperature exposure. *Integrative and Comparative
776 Biology*, 51(5), 691–702. <https://doi.org/10.1093/icb/icb097>

777 Smith, D. A., & Ridgway, M. S. (2019). Temperature selection in Brook Charr: lab experiments,
778 field studies, and matching the Fry curve. *Hydrobiologia*, 840(1), 143–156.
779 <https://doi.org/10.1007/s10750-018-3869-4>

780 Tomanek, L. (2010). Variation in the heat shock response and its implication for predicting the
781 effect of global climate change on species' biogeographical distribution ranges and
782 metabolic costs. *Journal of Experimental Biology*, 213(6), 971–979.
783 <https://doi.org/10.1242/jeb.038034>

784 Treberg, J. R., Killen, S. S., MacCormack, T. J., Lamarre, S. G., & Enders, E. C. (2016).
785 Estimates of metabolic rate and major constituents of metabolic demand in fishes under
786 field conditions: Methods, proxies, and new perspectives. *Comparative Biochemistry and
787 Physiology -Part A □: Molecular and Integrative Physiology*, 202, 10–22.
788 <https://doi.org/10.1016/j.cbpa.2016.04.022>

789 Venables, W. N., & Ripley, B. D. (2002). *Modern Applied Statistics with S* (4th ed.). Springer.
790 <http://www.stats.ox.ac.uk/pub/MASS4>

791 Warren, D. E., & Jackson, D. C. (2008). Lactate metabolism in anoxic turtles□: an integrative
792 review. *Journal of Comparative Physiology B*, 178, 133–148.
793 <https://doi.org/10.1007/s00360-007-0212-1>

794 Wells, Z. R. R., McDonnell, L. H., Chapman, L. J., & Fraser, D. J. (2016). Limited variability in
795 upper thermal tolerance among pure and hybrid populations of a cold-water fish.
796 *Conservation Physiology*, 4(1), cow063. <https://doi.org/10.1093/conphys/cow063>

797 Westley, P. A. H. (2020). Documentation of en route mortality of summer chum salmon in the
798 Koyukuk River, Alaska and its potential linkage to the heatwave of 2019. *Ecology and
799 Evolution*, 10(19), 10296–10304. <https://doi.org/10.1002/ece3.6751>

800 White, S. L., & Wagner, T. (2021). Behaviour at short temporal scales drives dispersal dynamics
801 and survival in a metapopulation of brook trout (*Salvelinus fontinalis*). *Freshwater Biology*,
802 66(2), 278–285. <https://doi.org/10.1111/fwb.13637>

803 Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L., François, R., Grolemund, G.,
804 Hayes, A., Henry, L., Hester, J., Kuhn, M., Pedersen, T. L., Miller, E., Bache, S. M.,
805 Müller, K., Ooms, J., Robinson, D., Seidel, D. P., Spinu, V., ... Yutani, H. (2019).
806 Welcome to the tidyverse. *Journal of Open Source Software*, 4(43), 1.3.0.
807 <https://doi.org/10.21105/joss.01686>

808 Wickham, H., François, R., Henry, L., & Müller, K. (2020). *A Grammar of Data Manipulation*
809 (0.8.5). <http://dplyr.tidyverse.org/>

810

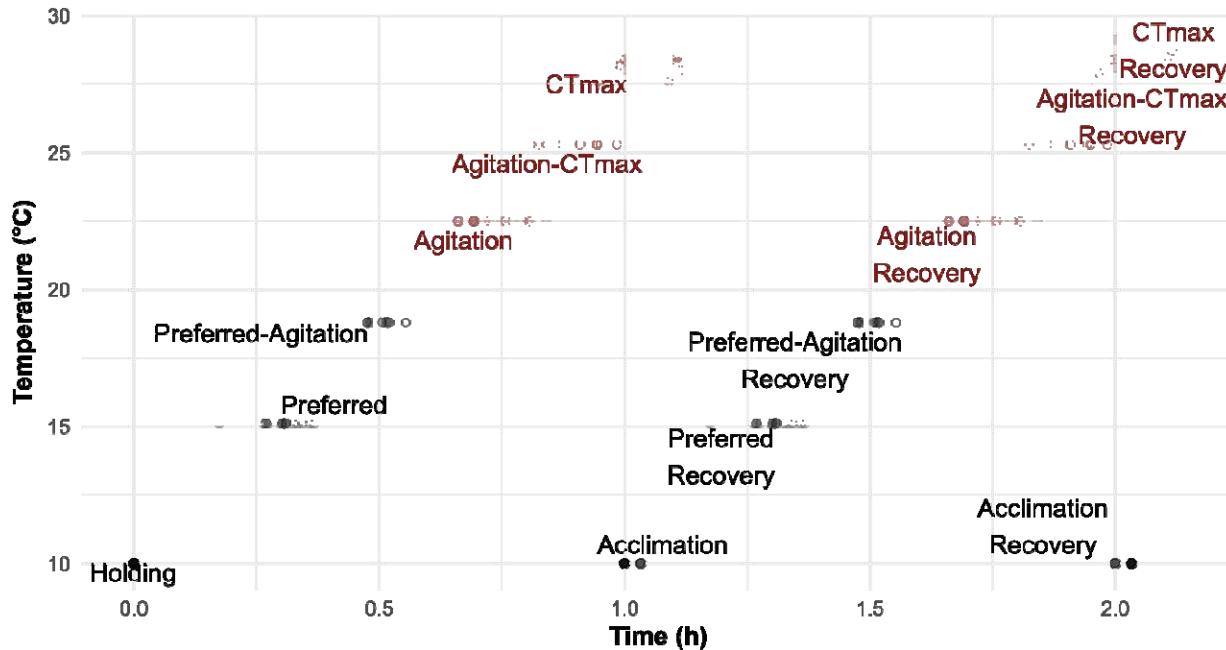
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Temperaure (°C)	Standard Metabolic Rate (mg O ² ·kg ⁻¹ ·h ⁻¹)	Maximum Metabolic Rate (mg O ² ·kg ⁻¹ ·h ⁻¹)	Aerobic Scope (mg O ² ·kg ⁻¹ ·h ⁻¹)
5	31.02 ± 5.03	358.90 ± 17.00	300.63 ± 17.37
10	38.27 ± 2.51	477.71 ± 17.19	396.30 ± 18.00
15	72.56 ± 3.92	498.66 ± 13.01	387.54 ± 12.76
20	105.40 ± 10.78	392.64 ± 19.88	242.45 ± 14.28
23	174.23 ± 25.64	325.50 ± 35.14	138.93 ± 42.70

813 **Table 1** – Mass specific metabolic rate estimates ± standard error (mg O²·kg⁻¹·h⁻¹) for juvenile Brook Trout assessed using
814 intermittent respirometry.

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817 **Figure 1** – Experimental design of treatments and timing for gene expression tissue sampling. Fish were sampled for white
818 muscle, liver, and gill. Each treatment was sampled either directly following heating to the treatment temperature or following a 1
819 h recovery period in a 10 °C recovery tank.

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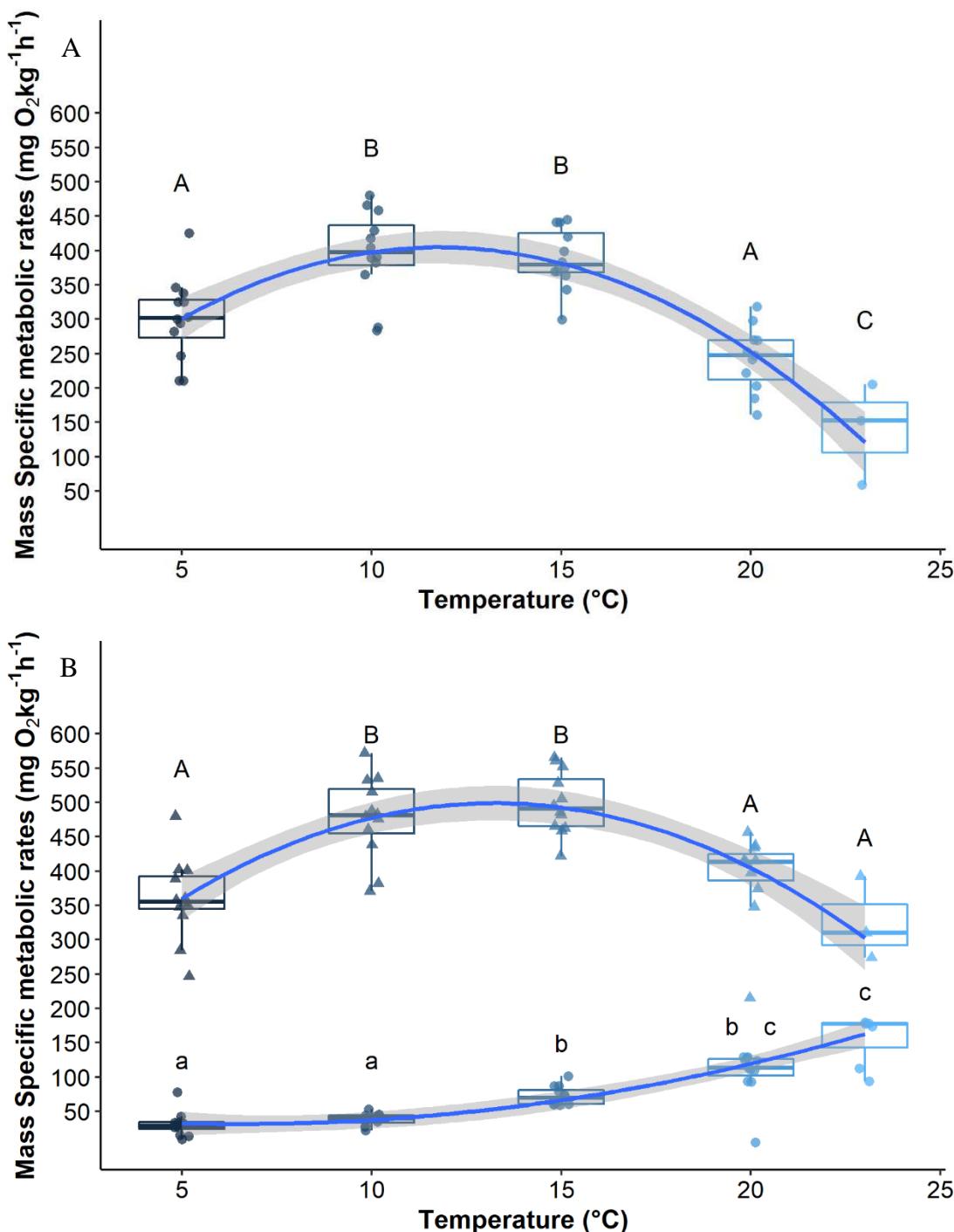
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823 **Figure 2** – Example of CT_{max} experimental setup. Treatments of $n = 16$ fish were used in this study (total of $n = 208$). Six
824 treatments (handling, acclimation, agitation, CT_{max} , B1 and B2) were separated into two time points: T^0 – where fish were
825 sampled immediately once either a certain temperature or behaviour was reached, or T^1 – where the fish were given a 1 h
826 recovery period in a recovery bath held at 10 °C once the temperature or behaviour was reached (Figure 1). A group of $n = 16$ fish
827 was also used as a control group to assess baseline stress levels. Fish were tested in groups of 4 in one of two 200 L green
828 sampling tanks, each in an individual acrylic chamber (volume – 1380 mL) with netting over the ends to contain and track the
829 fish. Each tank also contained an air stone to maintain dissolved oxygen levels above 90%, a temperature probe to monitor
830 heating rates, four 300 watt titanium heaters and two 5L/min pumps for water movement to ensure even heating throughout the
831 tank.

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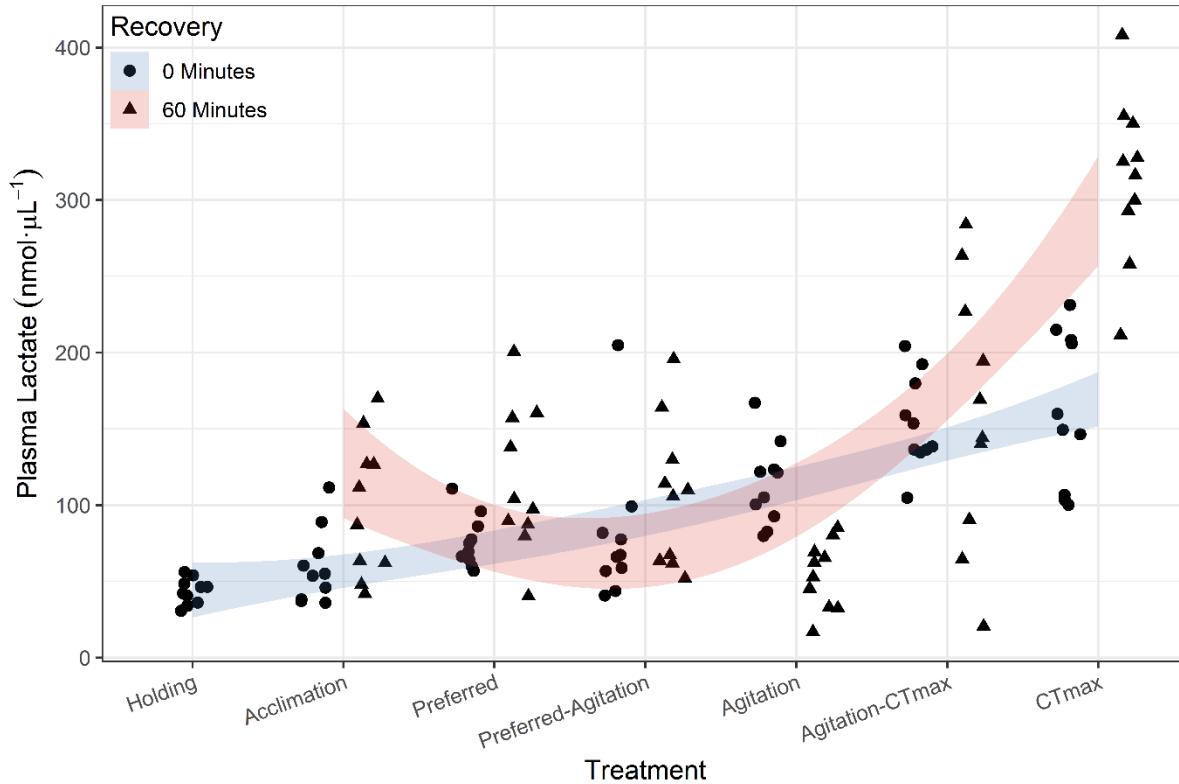
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835 **Figure 3** – Mass specific metabolic rate estimates (Plot A Aerobic Scope, Plot B Standard Metabolic Rate and Maximum
836 Metabolic Rate) for juvenile Brook Trout across an ecologically relevant range of acclimation temperatures. Temperatures that
837 share a letter are not significantly different. In Plot B triangles and uppercase letters represent maximum metabolic rate and
838 circles and lower case letters represent standard metabolic rate. Boxplots show the median, 25th and 75th percentile values, with
839 whiskers extending up to 1.5·IQR. Blue lines are LOESS smoothed regression lines to emphasize the general pattern.

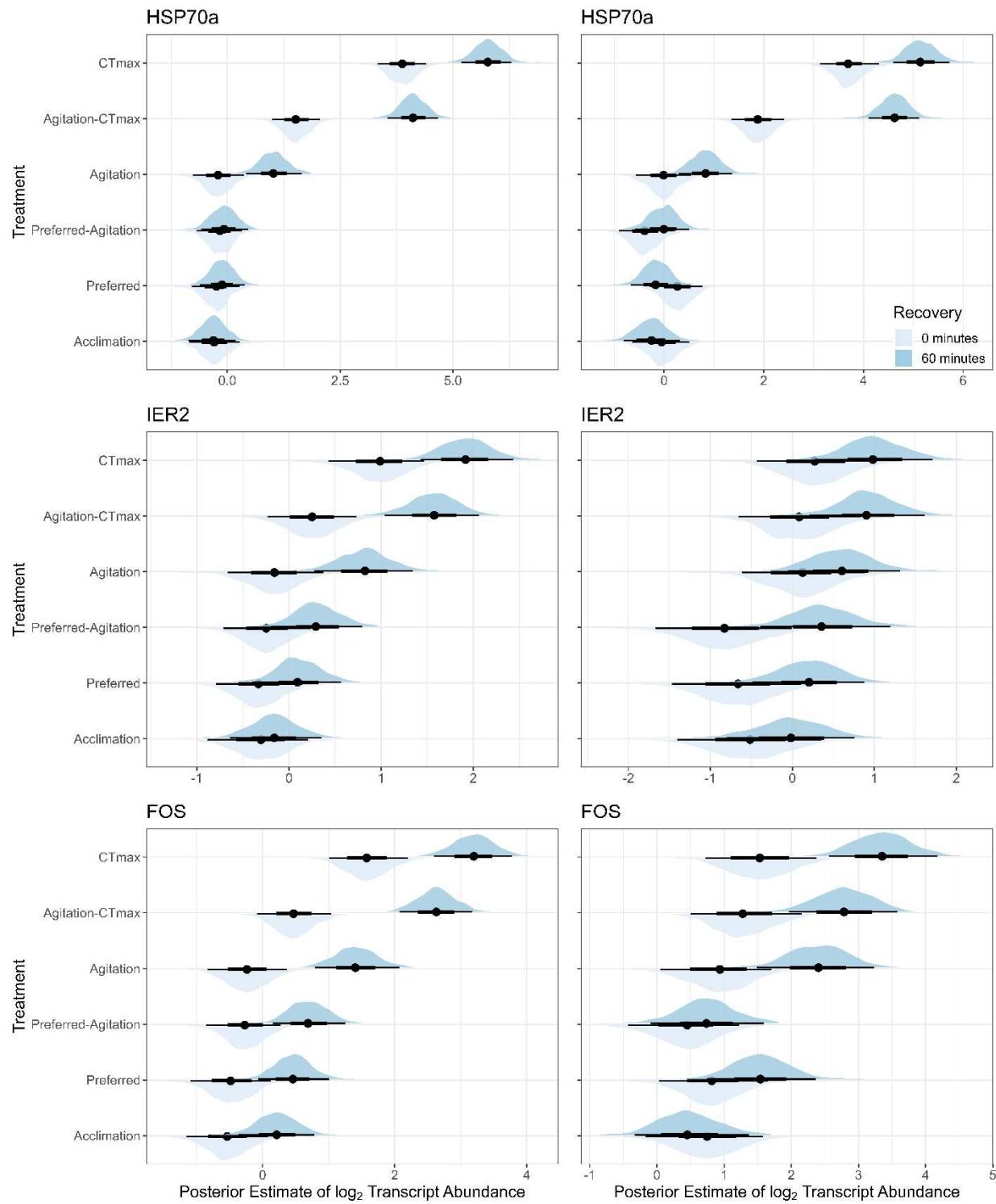
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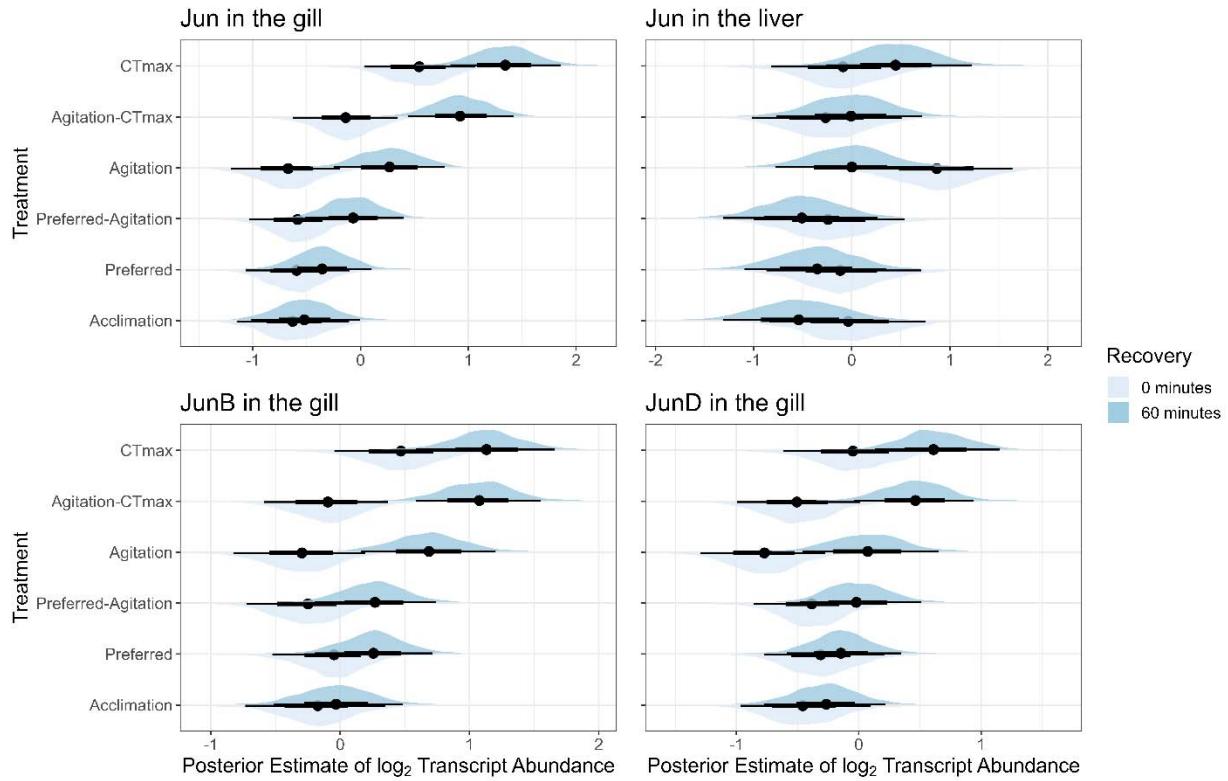
843 **Figure 4** – Plasma Lactate levels (nmol·μL⁻¹) increased with treatment temperature and were generally higher in the 60-min
844 recovery fish (T^1) than the fish sampled immediately at treatment temperature (T^0), except at the agitation temperature. Blue and
845 red lines are LOESS smoothed regression lines for T^0 and T^1 groups, respectively, to emphasize the general pattern.



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Figure 5 – Posterior estimates of abundance of *HSP70a*, *IER2* and *FOS* in the gill (left) and liver (right). Transcript abundance of these genes was observed to be different at agitation temperature and higher between the 0-min (T^0) sampling and 60-min (T^1) sampling periods.



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851 **Figure 6** – *Jun* transcript abundance in the gill tissue of juvenile Brook Trout. Lower ellipses represent tissues sampled at the 0-
852 min (T^0) sampling point and upper ellipses represent the 60-min (T^1) sampling point.

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