

1 **Metatranscriptomic analysis reveals synergistic activities of comammox and anammox**
2 **bacteria in full-scale attached growth nitrogen removal system.**

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12 **0- Abstract**

13 Leveraging comammox *Nitrospira* and anammox bacteria for shortcut nitrogen removal can
14 drastically lower the carbon footprint of wastewater treatment facilities by decreasing aeration
15 energy, carbon, alkalinity, and tank volume requirements while also potentially reducing nitrous
16 oxide emissions. However, their co-occurrence as dominant nitrifying bacteria is rarely reported
17 in full-scale wastewater treatment. As a result, there is poor understanding of how operational
18 parameters, in particular dissolved oxygen, impact their activity and synergistic behavior. Here,
19 we report the impact of dissolved oxygen concentration (DO = 2, 4, 6 mg/L) on the microbial
20 community's transcriptomic expression in a full-scale integrated fixed film activated sludge (IFAS)
21 municipal wastewater treatment facility predominantly performed by comammox *Nitrospira* and
22 anammox bacterial populations. 16S rRNA transcript compositions revealed anammox bacteria
23 and *Nitrospira* were significantly more active in IFAS biofilms compared to suspended sludge
24 biomass. In IFAS biofilms, anammox bacteria significantly increased *hzo* expression at lower
25 dissolved oxygen concentrations and this increase was highly correlated with the *amoA* expression
26 levels of comammox bacteria. Interestingly, the genes involved in nitrite oxidation by comammox
27 bacteria were significantly more upregulated relative to the genes involved in ammonia oxidation
28 with decreasing dissolved oxygen concentrations. Ultimately, our findings suggest that
29 comammox *Nitrospira* supply anammox bacteria with nitrite via ammonia oxidation and that this
30 synergistic behavior is dependent on dissolved oxygen concentrations.

31 **Keywords:** comammox-anammox cooperation, metatranscriptomics, IFAS, wastewater

32 **Synopsis** Comammox bacteria differentially regulate ammonia and nitrite oxidation in response
33 to dissolved oxygen concentration suggesting dissolved oxygen dependence of their synergistic
34 nitrogen removal with anammox bacteria in IFAS biofilms.

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38 **1- Introduction**

39 Recent studies have suggested that co-operation between comammox bacteria (CMX) and
40 anammox bacteria (AMX) can successfully result in nitrogen removal from mainstream
41 wastewater in both laboratory¹⁻³ and full-scale systems⁴. Such cooperation between CMX and
42 AMX could result in nitrogen removal with reduced oxygen and carbon demand, lower carbon,
43 alkalinity, and tank volume requirements, while also producing significantly reducing biotic
44 production of nitrous oxide (N₂O)^{5,6} which is a highly potent greenhouse gas (GHG). This contrasts
45 with conventional nitrogen removal systems, where estimates suggest 1-1.5% of incoming
46 nitrogen is lost as fugitive N₂O emissions⁷. Thus, the cooperation of CMX and AMX for
47 wastewater treatment has the potential to significantly decarbonize wastewater treatment.

48 While CMX have been found in soils⁸⁻¹⁰, drinking water systems¹¹⁻¹³, and traditional
49 activated sludge systems¹⁴⁻¹⁷, they are far more likely to be prefer biofilms in attached growth
50 systems and systems with long solids retention time (SRT)¹⁸ due to their slow growth rates^{19,20}.
51 Similarly, AMX may also prefer biofilms and granules due to their slower growth rates and because
52 such aggregate environments may offer protection from their sensitivity to nitrite inhibition in
53 nitrogen-rich wastewater sources^{21,22}, while also preventing exposure to dissolved oxygen^{23,24}. The
54 cohabitation of CMX and AMX in biofilms provides an ideal scenario to drive nitrogen loss via
55 their metabolic coupling. Specifically, CMX can drive down oxygen concentrations to ensure
56 anaerobic environments for AMX while providing them with nitrite. CMX could further oxidize
57 excessive nitrite to nitrate via nitrite oxidoreductase (*nxr*) activity, and potentially resupply nitrite
58 from nitrate in anaerobic conditions since activity of the nitrite oxidoreductase enzyme is
59 reversible. This sets up AMX to reduce ammonia and nitrite into dinitrogen gas establishing a
60 strong symbiotic relationship.

61 Our previous work identified CMX and AMX are dominant nitrifying bacteria in biofilms
62 in a full-scale integrated fixed film activated sludge (IFAS) process⁴; this was the first report to
63 observe their potential cooperation in a full-scale municipal wastewater treatment system. In this
64 study, we observed that anaerobic ammonia oxidation decreased, while aerobic ammonia oxidation
65 to nitrate increased with increasing dissolved oxygen concentration. In the IFAS biofilms, two
66 AMX bacterial populations associated with *Ca. Brocadia*, comprised $6.53 \pm 0.34\%$ of the
67 population while CMX bacteria closely associated with *Ca. Nitrospira nitrosa* comprised $0.37 \pm$
68 0.03% of the population. This system still had an abundance of strict ammonia oxidizing bacteria
69 (AOB) and nitrite oxidizing bacteria (NOB), particularly in the suspended phase making it unclear
70 the extent to which organisms were likely supplying nitrite to AMX¹⁸.

71 The current study follows-up our previous metagenomic and kinetic investigation by
72 systematically characterizing metatranscriptomic changes of the microbial communities in the
73 suspended phase and IFAS biofilms in response to changing DO concentrations. To do this, we
74 combine genome-centric metatranscriptomics as well as amplicon sequencing and quantitation of
75 major functional gene expressions (*amoA* for AOB, *amoA* for CMX, *nxrB* for *Nitrospira* bacteria,
76 and *hzo* for AMX) to determine (1) potential metabolic coupling between different nitrifying
77 populations within this mixed community, and (2) transcriptomic responses of key nitrifying
78 populations as a function of DO concentrations and growth mode (i.e., attached vs suspended

79 phase). By understanding the activity of all nitrifying organisms (AOB, CMX, NOB, AMX) in a
80 complex integrated fixed film activated sludge (IFAS) wastewater treatment facility, we can
81 identify DO concentrations that produce the highest activity and potential synergy between CMX
82 and AMX. Such insights have the potential to advance the development of mainstream partial-
83 nitritation-anammox nitrogen removal systems with lower energy requirements and GHG
84 emissions.

85 **2- Materials and Methods**

86 *2.1 Overview of Experimental Design*

87 A detailed overview of the wastewater treatment facility's process flow diagrams,
88 operational parameters, and conditions is reported in Vilardi et al 2023⁴. In brief, the secondary
89 treatment consists of five-zones in series including an anaerobic stage followed by an anoxic stage
90 receiving internal nitrate recycle divided into two cells, an aerobic stage with IFAS, and a small
91 deaeration zone. Samples were taken at the anaerobic stage (R1), end of the anoxic zone (R3), and
92 twice within the aerobic IFAS stage, at the beginning and the end (R4 and R5). Full-scale tank
93 dissolved oxygen concentrations in the aerobic zone were stabilized for six-hours prior to sampling
94 and experiments for multiple DO setpoints were run-on consecutive days within a week to
95 eliminate potential confounding effects from changes in microbial community composition.
96 Samples for DNA-based analyses were immediately stored in dry ice, while samples for RNA-
97 based analyses were mixed at a 1:10 ratio of LifeGuard Soil Preservation (Qiagen, Hilden
98 Germany) before being put on dry ice. At the end of each day's sampling, all samples were stored
99 at -80°C until extraction.

100 *2.2 Nucleic acid Extraction*

101 DNA was extracted using protocols from the DNeasy PowerSoil Pro Kit with some
102 modifications (Qiagen, Hilden, Germany) as described previously⁴. Extracted DNA was quantified
103 using a Qubit 4 Fluorometer and standard protocols for dsDNA with High Sensitivity
104 (ThermoFisher Scientific, Waltham, Massachusetts, USA). RNA was extracted using protocols
105 from the Quick-RNA Fecal/Soil Microbe Microprep Kit with some modifications (Zymo, Irvine,
106 California, USA). RNA samples were initially preserved with 1 mL of activated sludge (i.e.,
107 suspended phase) in 9 mL of LifeGuard Soil Preservation (Qiagen, Hilden Germany). This solution
108 was thawed and centrifuged (16,000 g for 1-minute) to pellet the activated sludge. The 1 mL of
109 activated sludge was then transferred into a 2 mL microcentrifuge tube and 1 mL of ultrapure
110 DNAse/RNase free water was added. To remove residual RNA stabilization solutions, three rinses
111 were performed where the samples were vortex for 30 seconds, centrifuged for 1-minute at 16,000
112 g, and 1.5 mL of supernatant was exchanged.

113 To remove the biomass from the IFAS plastic, IFAS and LifeGuard Soil Preservation were
114 transferred into a 5 mL bead beating tube from DNeasy PowerWater kits for 5-minutes of vortexing
115 (Qiagen, Hilden, Germany). Afterwards, 500 µL of biomass and LifeGuard Soil Preservation were
116 transferred into a clean 2 mL microcentrifuge tube for rinsing to remove the RNA stabilization
117 solution as described above.

118 Modifications made during the Quick-RNA Fecal/Soil Microbe Microprep Kit include
119 using a FastPrep-24 homogenizer for 40 seconds at 6 m/s during bead beating. An additional
120 DNase_1 treatment was performed to degrade residual DNA (Zymo, Irvine, California, USA)
121 followed by final elution of RNA in 50 μ L of buffer. The eluted RNA and residual DNA were
122 quantified via Qubit Fluorometer (ThermoFisher Scientific, Waltham, Massachusetts, USA). After
123 quantification several aliquots were made where some eluted RNA was immediately frozen at -
124 80°C for metatranscriptomic sequencing, while another aliquot was processed for cDNA synthesis
125 for RT-qPCR and amplicon sequencing before being frozen at -20°C.

126 *2.4 Quantification of Transcripts and Genes*

127 All samples were quantified for expression which included duplicate extractions from two
128 technical replicates from all three dissolved oxygen settings, at all reactor locations (R1, R3, R4,
129 and R5 for both suspended sludge biomass and IFAS). For DNA-based gene quantification, only
130 technical duplicates from R1, R5 for suspended sludge biomass and IFAS were used as there was
131 no significant change in community structure over the short experimental period (i.e., less than
132 one week). The extracted RNA was synthesized into cDNA using standard protocols for
133 SuperScript IV First-Strand Synthesis System for RT-PCR (Invitrogen, Waltham, Massachusetts,
134 USA). The cDNA was quantified via Qubit Fluorometer (ThermoFisher Scientific, Waltham,
135 Massachusetts, USA). RT-qPCR and qPCR were then performed on a QuantStudio 7 Flex
136 (ThermoFisher Scientific, Waltham, Massachusetts, USA). The reaction mastermix contained 0.5
137 μ L of both forward and reverse primers at 10 μ M, 10 μ L of Luna Universal qPCR Master Mix
138 (New England Biolabs, Ipswich, Massachusetts, USA), 0.5 μ L of template DNA/cDNA, and 8.5
139 μ L of nuclease-free water. All primers were purchased by Integrated DNA Technologies (IDT,
140 Coralville, Iowa, USA).

141 Primers and their thermocycling conditions are shown in **Supplemental Table 1** using
142 standard protocols for 16S rRNA²⁵, ammonia monooxygenase subunit A (*amoA*) for strict AOB²⁶,
143 *amoA* for *Ca. Nitrospira nitrosa*²⁷, nitrite oxidoreductase subunit B (*nxrB*) for *Nitrospira* bacteria²⁸,
144 and hydrazine oxidoreductase (*hzo*) found in AMX²⁹. Quantification data of transcripts is shown
145 in **Supplemental Figure 1** with a comparison between reporting RT-qPCR results as
146 log(copies/mL) and log(copies/g of Total Solids) in **Supplemental Figure 2**.

147 *2.5 Amplicon Sequencing*

148 Extracted DNA and cDNA were submitted to Georgia Institute of Technology's Molecular
149 Evolution Core where for sequencing using the Illumina MiSeq v3 kit PE300 using Illumina
150 Nextera Adapter Sequences (Illumina Inc, San Diego, California, USA). The samples were
151 demultiplexed and barcodes removed by the sequencing core. The raw sequences were then
152 imported into RStudio v2023.06.2³⁰ and processed using DADA2 v1.26³¹. In DADA2, 16S rRNA
153 amplicons used standard protocols with a lower, default truncQ = 2. Other amplicons did not use
154 any truncation, and during merging maxMismatch was increased to 5 to allow longer amplicons
155 to have substantial overlap. Taxonomic assignments referenced the SILVA SSU 138.1 database³²
156 and additionally checked in the with the MiDAS 5.0 database³³. Merged reads outside \pm 2 of the
157 expected amplicon length were discarded. ASV's were then consolidated into OTUs with a 97%

158 identity cutoff using BioStrings v2.66.0³⁴, dplyr v1.1.2³⁵, tibble v3.2.1³⁶, and DECIPHER
159 v2.26.0³⁷. Histograms of postprocessed amplicon reads are available in the supplemental materials
160 (**Supplemental Figure 3**). Amplicon compositions for *amoA* for AOB, *amoA* for CMX, *nxrB* for
161 *Nitrospira* sp., and *hzo* for AMX can be found in **Supplemental Figure 4** with top BLAST matches
162 based on e-value with the reported sequence percent ID in similarity in **Supplemental Table 2**.

163 Absolute quantification of OTUs was estimated by multiplying the OTU relative
164 abundance with RT-qPCR copies for each gene. The same primers were used for RT-qPCR
165 amplification and amplicon sequencing to ensure parity with previous similar studies³⁸⁻⁴².

166 2.6 Metatranscriptomic sequencing and data processing

167 A limited subset of samples were submitted for metatranscriptomic sequencing in
168 duplicate, limited to only DO concentrations of 2 mg/L and 6 mg/L across reactor locations R1,
169 R3, and R5 for both suspended sludge biomass and IFAS. Metagenomic analysis of this wastewater
170 treatment plant is previously described⁴.

171 Extracted RNA samples were submitted to the Georgia Institute of Technology's Molecular
172 Evolution Core for rRNA depletion, and RNA sequencing. RNA integrity, read length, and
173 quantification were performed on a 2100 Bioanalyzer with C0.1.069 firmware (Agilent, Santa
174 Clara, California, USA). Excess rRNA was depleted using QiaSeq FastSelect 5S/16S/23S
175 depletion kit as well as the FastSelect HMR and Plant rRNA probes (Qiagen, Hilden, Germany).
176 RNA sequencing was performed on an Illumina NovaSeq S1 PE100bp to obtain 1.6 billion reads
177 (Illumina Inc, San Diego, California, USA). Samples were demultiplex and barcodes removed by
178 the core facilities prior to releasing sequences.

179 Raw reads were preprocessed using fastp v0.23.4⁴³ before residual rRNA reads were
180 separated from mRNA reads using SortMeRNA v2.1⁴⁴. The remaining mRNA reads were then
181 competitively mapped to previously assembled metagenome-assembled genomes (MAGs) from
182 AOB, CMX, NOB, and AMX^{4,14,18}. The MAGs were annotated using bakta v1.8.2⁴⁵. The mapped
183 mRNA reads were converted into .bam files, sorted, and indexed using samtools v1.18⁴⁶. Finally,
184 the mapped mRNA reads were quantified for read counts on the annotated MAGs using dirseq
185 v0.43⁴⁷.

186 2.7 Statistical Methods

187 Statistical analysis was performed in RStudio³⁰ with the vegan package⁴⁸. ANOVA was
188 used to determine the impact of DO and reactor zonation across all conditions, while Tukey Honest
189 Significant Differences⁴⁹ post-hoc test was used to compare means between two subgroups. For
190 correlative regression analysis, ranked-based linear regressions were performed in R to determine
191 significant trends across RT-qPCR samples. Shannon Diversity Index⁵⁰ was additionally used to
192 compare the alpha diversity across all amplicon sequencing samples. Differential gene expression
193 analysis was performed with Deseq2⁵¹. For CMX, the minimum read counts per gene were set to
194 five due to lower transcriptomic reads, whereas the default of 10 read counts was used for AMX.

195 2.8 Data Availability

196 All code for amplicon sequencing, metatranscriptomic analysis, statistical analysis and
197 plots are available at github.com/queermsfrizzle/CMX_AMX_transcriptomics. Raw fastq files for
198 metatranscriptomic sequencing and amplicon sequencing are available via NCBI bioproject
199 submission number PRJNA1050761.

200 **3- Results and Discussion**

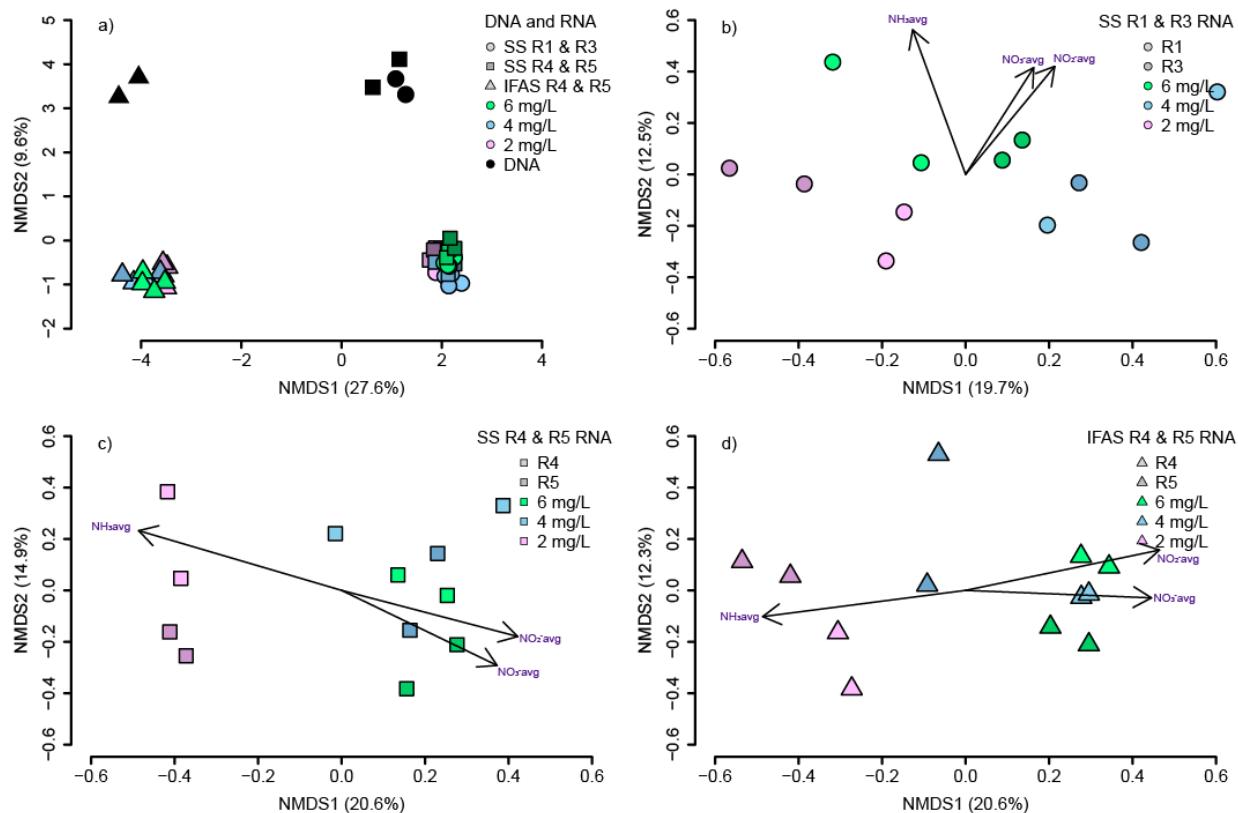
201 *3.1 Microbial community composition and activity was associated with growth phase and DO*
202 *concentrations.*

203 The total and active microbial community composition was distinct from each other
204 between the two phases (i.e., suspended sludge biomass vs attached IFAS phase) based on 16S
205 rRNA gene and transcript composition (**Figure 1**). The largest difference in community
206 composition was between IFAS and suspended sludge biomass which (**Figure 1a**) and explained
207 27.6% of the total variance between samples (PERMANOVA $p << 0.05$). There was also a
208 significant difference between each community's 16S rRNA transcript composition and 16S rRNA
209 gene composition which explained 9.6% of the total variance (**Figure 1a**) (PERMANOVA $p << 0.05$).
210 This is consistent with previous reports which showed the most significant changes to
211 community composition based on reactor zonation followed by the separation of DNA and RNA,
212 and further by the experimental variable of seasonality³⁸. When analyzing the Shannon Diversity
213 Index of 16S rRNA transcripts and genes, there was a significantly higher amount of diversity in
214 16S rRNA genes, compared to transcripts (ANOVA $p << 0.05$) (**Supplemental Figure 4**). This
215 suggests that a large proportion of the microbial community is likely less active.⁵²

216 For both IFAS biofilms and suspended sludge biomass communities, the DO concentration
217 had a significant impact on the active community composition (PERMANOVA $p_{IFAS} = 0.0015$, $p_{ss} = 0.0007$) (**Figure 1b, c, & d**). For the IFAS biofilm community, there was a statistically significant
218 difference between DO = 2 mg/L with DO = 4 mg/L as well as DO = 6 mg/L (ANOSIM $p_{IFAS}^{D2D4} = 0.027$ and $p_{IFAS}^{D2D6} = 0.003$), but not between DO = 4 mg/L and DO = 6 mg/L (ANOSIM $p_{IFAS}^{D4D6} = 0.5687$). In suspended sludge biomass, all dissolved oxygen concentrations showed statistically
219 significant separation (ANOSIM $p_{ss}^{D2D4} << 0.05$, $p_{ss}^{D2D6} << 0.05$, and $p_{ss}^{D4D6} = 0.0127$). While
220 the DO concentrations did result in significant changes, the measured concentrations of DO,
221 ammonia, nitrate, nitrite, alkalinity, and COD all showed statistically significant effects (all
222 Mantels test $p << 0.05$) on differences between the active community composition between
223 samples from the IFAS media or suspended sludge biomass. This indicates that the DO setpoint
224 has cascading effects on microbial community activity and redox conditions due to residual
225 oxygen, nitrite, and nitrate. Similar impacts of dissolved oxygen impacting the microbial
226 community composition has been previously shown.⁵³⁻⁵⁵

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233 **Figure 1:** NMDS plots constructed using Bray-Curtis dissimilarity estimates for a) all 16S rRNA gene and transcript OTUs,
234 b) SS for R1 and R3 16S rRNA transcript OTUs, c) SS for R4 and R5 16S rRNA transcript OTUs, and d) IFAS R4 and R5
235 16S rRNA transcript OTUs. Black data points represent genes while colored data points represent transcripts. Green,
236 blue, and pink shades represent the dissolved oxygen concentration at 6 mg/L, 4 mg/L and 2 mg/L, respectively. Shades
237 of each color represent different specific reactor locations while circles, squares and triangles represent the reactor
238 zones.

239
240 3.2 High diversity of nitrite oxidizing bacteria compared to strict AOB and AMX.

241 The diversity of nitrifying bacteria was analyzed using a combination of 16S rRNA
242 transcript sequencing and quantification and amplicon sequencing of major nitrogen cycling
243 functional genes (Figure 2). 16S rRNA transcript analysis averaged $63,049 \pm 11,365$ postprocessed
244 reads per sample (Supplemental Figure 3) and revealed one dominant AMX (Figure 2a), two
245 dominant AOB (Figure 2c) and three *Nitrospira*-like OTUs (Figure 2b).

246 Anammox activity was dominated by OTU48, *Candidatus Brocadia*, based on 16S rRNA
247 transcript quantification. OTU48 was particularly active in IFAS and comprised $16.5\% \pm 4.2\%$ of
248 the 16S rRNA transcript composition at low dissolved oxygen and declined down to $9.5\% \pm 2.2\%$
249 at higher dissolved oxygen concentrations. Similarly, there was a single OTU, *hzo* 7, which
250 dominated anammox activity based on *hzo* transcript absolute abundance (Figure 2d) and
251 consistently comprised $93.43\% \pm 2.85\%$ of the *hzo* transcript composition. IFAS had significantly
252 higher *hzo* expression at low dissolved oxygen (ANOVA $p < 0.05$) at 8.64 ± 0.13 log(copies/mL)

253 and decreased with higher dissolved oxygen to 7.92 ± 0.13 log(copies/mL). AMX's *hzo* expression
254 was stable in the suspended sludge biomass at 7.69 ± 0.24 log(copies/mL) and not impacted by
255 dissolved oxygen concentration (ANOVA $p = 0.337$). While AMX preferring lower dissolved
256 oxygen concentrations has been previously reported^{23,56}, what is interesting when comparing our
257 system is the lack of diversity in AMX, whereas typically multiple species of abundant AMX⁵⁷ are
258 observed to co-occur from wastewater bioreactors seeing *Ca. Brocadia* and *Ca. Kuenenia*
259 together⁵⁶, to rice-wheat paddy fields observing *Ca. Brocadia* and *Ca. Jettenia*⁵⁸. While there are
260 low abundance species observed (**Supplemental Figure 4**), these species are orders of magnitude
261 lower in abundance and also lower levels of activity.

262 Three dominant *Nitrospira* OTUs were observed based on 16S rRNA transcript analysis:
263 OTU544 closely associated with *Nitrospira defluvii*, OTU163 was associated with *Nitrospira*
264 *moscoviensis*, and OTU417 was unknown *Nitrospira* (**Figure 2b**). For redundancy, 16S rRNA
265 sequences were compared across SILVA SSU 138.1 and MIDAS 5.0 databases to assess whether
266 CMX associated 16S rRNA sequences were distinguishable from NOB, but this was not feasible
267 (data not shown). All 16S rRNA transcripts for *Nitrospira* were significantly more active in IFAS
268 biofilm than in suspended sludge biomass (ANOVA $p << 0.05$).

269 In contrast to the 16S rRNA transcript data which showed remarkable differences in
270 *Nitrospira* activity in IFAS biofilm as compared to suspended sludge biomass, RT-qPCR data
271 indicates relative stable transcriptional activity of the *nxrB* gene in the suspended sludge biomass
272 (7.99 ± 0.23 log(copies/mL)) which was comparable to that of the IFAS biofilm based *nxrB* activity
273 (**Figure 2e**). Interestingly, *nxrB* transcript abundance dropped significantly with increasing
274 dissolved oxygen (ANOVA $p << 0.05$) from 7.89 ± 0.08 log(copies/mL) to 7.31 ± 0.12
275 log(copies/mL) from DO = 2 mg/L up to DO = 6 mg/L in the IFAS biofilm community. If *nxrB*
276 activity was only associated with nitrite oxidation, this drop would be perplexing since ammonia
277 oxidation is higher at high dissolved oxygen concentrations. However, this drop in *nxrB* transcripts
278 could be indicative of lower nitrite concentrations found in IFAS biofilm since the available nitrite
279 is likely consumed by AMX, or alternatively in anoxic or low DO conditions *nxrB* could reversibly
280 reduce nitrate back to nitrite. This would explain the higher *nxrB* transcripts in IFAS biofilm at
281 lower dissolved oxygen concentrations.

282 To determine if any *nxrB* transcripts were from CMX, the *nxrB* transcripts were mapped to
283 all *Nitrospira* MAGS. Several low abundance *nxrB* OTUs mapped to CMX accounting for 1.02%
284 $\pm 0.84\%$ of the *nxrB* transcript composition without significant changes due to dissolved oxygen
285 concentration (ANOVA $p = 0.83$). Interestingly, the *amoA* CMX activity was significantly higher
286 in the IFAS biofilm compared to the suspended phase and demonstrated sensitivity to DO
287 concentration (ANOVA $p = 0.033$) increasing from 6.43 ± 0.32 log(copies/mL) up to 7.33 ± 0.58
288 log(copies/mL) with drop in DO from 6 to 2 mg/L (**Figure 2f**). Overall, in suspended sludge
289 biomass CMX appear to have a stable transcriptional activity of *nxrB* and a low activity of *amoA*
290 whereas in IFAS biofilm a significant increase in *amoA* activity is observed. Interestingly, CMX
291 *amoA* activity did significantly increase at reduced DO, even in suspended sludge biomass. This
292 is uncommon since *amoA* activity is regulated by ammonia concentration, not oxygen
293 concentration. This change in expression in the suspended phase is likely due to residual *amoA*

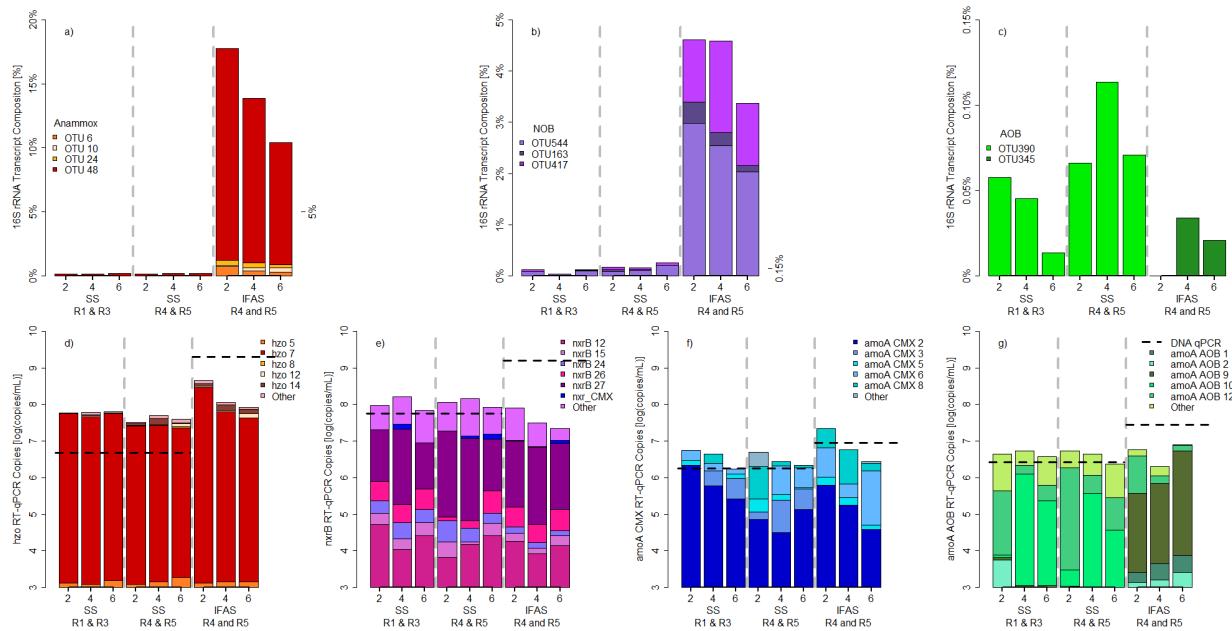
294 transcripts of biomass being sloughed off from IFAS biofilm into the suspended phase which
295 would experience fluctuations in ammonia concentration, therefore impacting *amoA* transcript
296 regulation based on DO concentration.

297 This aligns with current studies showing that CMX strongly prefer biofilms⁵⁹ and our
298 previous work using intrinsic assays and differential inhibition to suggest higher ammonia
299 oxidation from CMX in IFAS biofilm⁴. Other studies using differential inhibitors with CMX
300 enrichment cultures from rotating biological contactors have shown similar changes in CMX
301 activity between biofilm and suspended sludge biomass cultures⁶⁰. While our study suggests a
302 greater in nitrite oxidation activity of CMX in suspended sludge biomass, future studies are
303 required to investigate the mechanisms underpinning CMX differential expression in suspended
304 and attached phases.

305 Two dominant strict AOB (OTU390, OTU345) within the family *Nitrosomonadaceae* were
306 detected in the 16S rRNA transcript amplicon sequencing library. While both OTUs showed very
307 low levels of activity based on their relative abundance in the 16S rRNA transcript libraries,
308 OTU390 was only active in the suspended phase, while only OTU345 demonstrated low levels of
309 activity in the IFAS biofilms (**Figure 2c**). While the absolute abundance of AOB was significantly
310 higher in IFAS biofilm samples (7.45 ± 0.9 log(copies/mL)) compared to suspended sludge
311 biomass samples (6.42 ± 0.26 log(copies/mL)), their transcriptional activity based on RT-qPCR
312 assay targeting the *amoA* gene indicated no significant change with either reactor zonation or DO
313 concentration (ANOVA_{DO}=0.426, ANOVA_{zone}= 0.839) averaging at 6.63 ± 0.31 log(copies/mL)
314 (**Figure 2g**). This indicates a significant downregulation of *amoA* transcriptional activity (on a per
315 cell basis) in the IFAS biofilm community. Interestingly, dissolved oxygen or reactor zonation did
316 not impact the transcriptional activity of *amoA* for AOB. This stable activity of *amoA* by AOB is
317 consistent with past activated sludge systems⁶¹⁻⁶³ but overall lower than typically reported between
318 7-8 log(copies/mL). This is consistent with our previous findings that CMX and AMX significantly
319 contribute to nitrogen removal, especially in IFAS biofilm phase⁴.

320 Multiple *amoA* OTUs were identified in both the suspended sludge biomass and IFAS
321 biofilm phase which is in contrast with the single 16S rRNA gene OTU in each phase. This could
322 likely emerge due to both the presence of multiple *amoA* gene copies within genomes relative to a
323 single 16S rRNA gene per AOB genome⁶⁴, but also could emerge from the differences in the
324 sensitivities of two assays. Specifically, *amoA* gene amplicon sequencing was performed after
325 selectively amplifying this gene it from cDNA, and as a result is likely to detect sequences from
326 more rare populations as compared to when the entire microbial community is profiled using 16S
327 rRNA gene assay. Nonetheless, in alignment with the 16S rRNA transcript data, a single OTU,
328 *amoA* 9, was primarily active in the IFAS biofilm community. This *amoA* 9 transcript exhibited
329 sensitivity to DO concentration increasing from 57.2% at 2 mg/L to 74.4% of all AOB *amoA*
330 transcripts at 6 mg/L. Furthermore, the *amoA* 9 OTU sequence exhibited 99.6% sequence
331 similarity with an uncultured *Nitrosomonas* clone A12⁶⁵. This *Nitrosomonas* clone A12 was
332 previously found in a biofilm working alongside CMX and raises the possibility that specific
333 *Nitrosomonas* species are more inclined to co-exist with CMX or potentially both prefer biofilms
334 due to their growth kinetics.

335



336

337 **Figure 2; Summary of dominant nitrifying bacteria's transcriptional activity based on 16S rRNA transcript sequencing**
338 **(a,b,c) and functional gene expression (d,e,f,g).** 16S rRNA transcript activity with a) anammox bacteria b) all nitrite
339 oxidizing bacteria (including comammox which cannot be differentiated using OTUs of 16S rRNA V4 amplicons), and c)
340 strict ammonia oxidizing bacteria. The expression of functional genes are quantified via RT-qPCR and then multiplied
341 with OTU relative abundance in amplicon sequencing data (y-axis) for d) hydrazine oxidoreductase found in AMX e)
342 nitrite oxidoreducatase of all nitrite oxidizing bacteria (including comammox *Nitrospira*), f) ammonia monooxygenase
343 found in comammox *Nitrospira* and g) ammonia monooxygenase found in strict ammonia oxidizing bacteria. The dashed
344 line represents the average abundance of each functional gene via qPCR to determine transcripts-to-gene ratios.

345

346 *3.3 Comammox Nitrospira and anammox bacteria have strong synergistic relationship in the IFAS*
347 *biofilm phase*

348 CMX and AMX show a strong symbiotic relationship in IFAS biofilm as indicated by
349 significant correlative relationships (**Figure 3a**) and transcripts-per-million gene expression
350 analysis of major nitrogen cycling pathways (**Figure 3b & c**). Using RT-qPCR transcript
351 abundances, *hzo* transcripts were regressed against *amoA* for CMX transcripts in IFAS biofilm
352 (**Figure 3a**) and a significant ranked linear correlation was observed (linear regression $R^2 = 56.1\%$
353 & $p = 0.003$) suggesting that AMX *hzo* and CMX *amoA* activities may be potentially co-regulated
354 by changes in DO concentrations. While similar regressions can be performed using *nxrB*, CMX
355 *amoA* and AMX *hzo* are the only genes that exhibited significantly higher transcriptional activity
356 in the IFAS biofilm relative to the suspended phase. This contrasts with expression levels of AOBs
357 *amoA* and total *nxrB* transcripts which were either stable or declined in the IFAS biofilm,
358 respectively. This suggests that aerobic ammonia oxidation by CMX was strongly associated with
359 anaerobic ammonia oxidation by AMX, suggesting that CMX may play a role in supplying nitrite
360 to AMX. This has been suggested since the discovery of CMX coexisted with an AMX

361 population^{66,67} but this study is the first to demonstrate this symbiotic relationship leveraging
362 transcriptomic expression.

363 This relationship was further investigated by examining differential expression of key
364 nitrogen cycling genes in MAGS from CMX (**Figure 3b**) and AMX (**Figure 3c**) by mapping
365 metatranscriptomic reads to them. Similar analysis for AOB and NOB can be found in the
366 supplemental materials (**Supplemental Figure 6**). Nitrogen cycling genes from CMX were most
367 impacted by changes in DO with significant increases in *nxrA* across every reactor setting (Tukey
368 all p << 0.05) at lower DO concentrations. In suspended sludge biomass, nitrogen cycling
369 pathways accounted for 7.75% ± 1.62% of the CMX transcripts at low dissolved oxygen
370 concentrations and only 4.31% ± 0.85% at high DO concentrations. This was further highlighted
371 when comparing activity in IFAS biofilm with 18.53% ± 2.89% and 7.18% ± 1.40% at DO = 2
372 mg/L and DO = 6 mg/L of all transcripts respectively mapped to nitrogen cycling pathways. The
373 largest increase in activity was associated with *nxrA* in IFAS biofilm which increased 9.1-fold at
374 low DO concentrations. This was surprising since *nxrC* expression did not significantly change
375 between DO setting and was often undetected but this may be a result of a fragmented MAG
376 resulting in low mapped transcripts since other studies have shown consistent transcriptional
377 activity across all *nxr* subunits⁶⁸⁻⁷⁰. Further work needs to investigate biofilm depth-dependent
378 activities and oxygen gradients to understand the how CMX regulates expression of nitrogen
379 cycling genes across a redox gradient.⁷¹

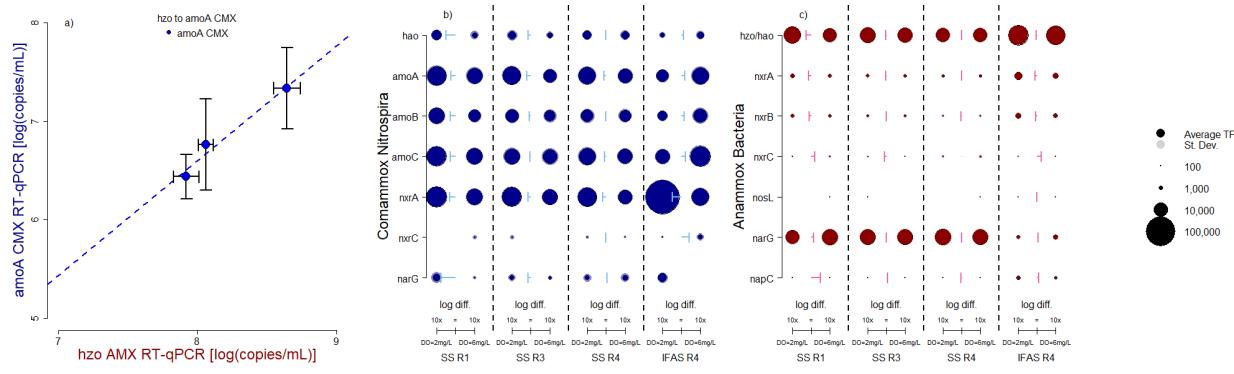
380 AMX maintained mostly consistent expression of nitrogen cycling genes in suspended
381 sludge biomass comprising 2.52% ± 0.19% of the total mapped AMX transcripts (**Figure 3c**). In
382 comparison to IFAS biofilm, *narG* (nitrate reductase), was significantly more active in suspended
383 sludge biomass (Tukey p << 0.05) as compared to the IFAS biofilm. This suggests that in
384 suspended sludge biomass AMX may need to supply their own nitrite from residual nitrate due to
385 competition for nitrite with NOB. The downregulation of *narG* in IFAS biofilm could indicate
386 sufficient supply of nitrite, likely from CMX. Additionally, in IFAS biofilm, most pathways
387 experienced a significant increase in transcripts-per-million at low DO concentrations. The
388 expression of genes associated with *hao/hzo* pathway increased 1.2x, *nxrA* increased 1.68x, and
389 *nxrB* increased 1.36x. These increases results in AMX's nitrogen cycling transcripts in IFAS
390 biofilm increasing from 2.65% ± 0.23% at DO = 6 mg/L up to 3.23% ± 0.08% at DO = 2 mg/L
391 (Tukey p << 0.05) with overall activity in IFAS biofilm being significantly higher than in
392 suspended sludge biomass.

393

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Figure 3: Anammox bacteria and comammox Nitrospira have a strong linear relationship of their activity in IFAS biofilm based on a) RT-qPCR transcript quantification of hydrazine oxidoreductase and ammonia monooxygenase from comammox Nitrospira. This symbiotic relationship was further explored using transcripts-per-million activity profiles of major nitrogen cycling pathways mapped to MAGS constructed of b) comammox Nitrospira and c) anammox bacteria. The size of each point relates to the magnitude of transcripts-per-million, with an outline for standard deviation. The arrows show the magnitude difference between each data point DO = 2 mg/L and DO = 6 mg/L.

403

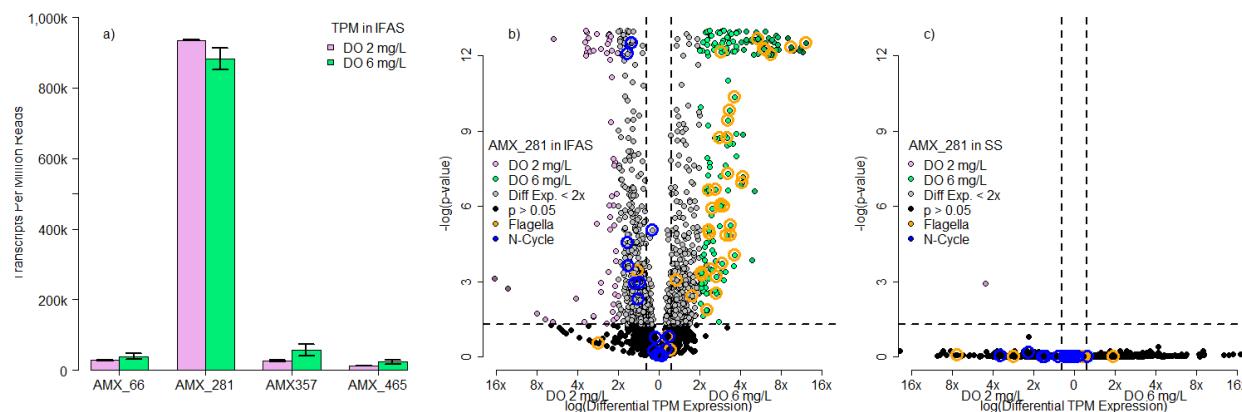
404 3.4 Differential Gene Expression Analysis Anammox Increases Flagella Activity in IFAS biofilm 405 phase under high DO conditions

406 AMX_281 was the most actively transcribed MAG in IFAS biofilm with $93\% \pm 0.2\%$ of
407 the transcripts mapping at DO = 2 mg/L, and $88.3\% \pm 3.1\%$ at DO = 6 mg/L (Figure 4a). This is
408 consistent with dominance of 16S rRNA transcript OTU 48 whose transcripts completely mapped
409 to the MAG, and the dominance of *hzo* transcript OTU 7, which did not map to the MAG due to
410 an incomplete genome.

411 AMX_281 was further analyzed in IFAS biofilm (Figure 4b) and in suspended sludge
412 biomass (Figure 4c) with genes considered significantly differentially transcribed with a Wald test
413 $p < 0.05$ and greater than 2x transcripts-per-million between two conditions being compared (i.e.,
414 2 vs 6 mg/L of DO or biofilm vs suspended phase). Most genes associated with nitrogen cycling
415 functions in AMX_281 showed insignificant changes in relative expression abundance, similar to
416 all AMX in Figure 3c. If nitrogen cycling genes are up-regulated as shown with *hzo* quantification
417 via RT-qPCR in Figure 2d, they increase proportionally to maintain similar relative expression
418 abundance via transcripts-per-million and differential gene expression (Figure 3c and Figure 4b).
419 At lower DO concentrations, most up-regulated gene expression was from hypothetical proteins
420 and was not further investigated.

421 At high dissolved oxygen concentrations in IFAS biofilm, there was a significant increase
422 in most flagella-related transcriptional activity, which was not similarly observed in suspended
423 sludge biomass. This was observed for genes across multiple flagellar related proteins from FlgN,
424 FliS, Flagellar hook-associated protein 2, and FlaG, all of which increased more than five-fold at
425 high DO concentrations. Yan et al. 2020²³ suggests this maybe be oxygen related-stress response
426 and trigger biofilm formation. While reactive oxygen species could impact AMX in IFAS biofilm,
427 this stress response is not consistent with AMX in suspended sludge biomass (Figure 4c) which

428 did not show significant changes in flagella-related activities and accounted for an overall lower
 429 abundance of transcripts-per-million (IFAS_{DO6} = 1.78%, SS_{DO6} = 0.08%). We hypothesize that
 430 anammox bacteria in IFAS biofilms may increase mobility due to chemotaxis and cell growth
 431 within the competitive environment of a biofilm. Research into anammox located in deep sea
 432 anoxic nitrate-ammonium transition zones observed expression of full gene sets for flagella
 433 mobility despite limited energy for mobility, suggesting flagellar activity is linked to in situ
 434 growth⁷². Similarly, other *Planctomycetes* bacteria have shown increased flagella mobility in
 435 highly rich media⁷³ suggesting AMX may utilize chemotaxis within high nutrient concentrations
 436 as opposed to mobility to avoid oxygen stress⁷⁴.



437
 438 **Figure 4: Anammox MAG, AMX_281, was highly expressive in IFAS biofilm compared to other MAGS (Figure 4a), and**
 439 **further analyzed using Deseq2 for differential gene expression analysis in b) IFAS biofilm and c) suspended sludge**
 440 **biomass. Points in pink were statistically significantly more active and expressive at DO = 2 mg/L while points in green**
 441 **were statistically significantly more active and expressive at DO = 6 mg/L. Blue circles represent shifts in nitrate**
 442 **transporting, while the orange circles represent transcripts related to flagella activity. In instances where the -log(p-value)**
 443 **exceeded 12, they are plotted as 12.5 with a jitter for spacing.**

444
 445 *3.5 Comammox Nitrospira alters ammonia and nitrite metabolic preference based on redox*
 446 *conditions.*

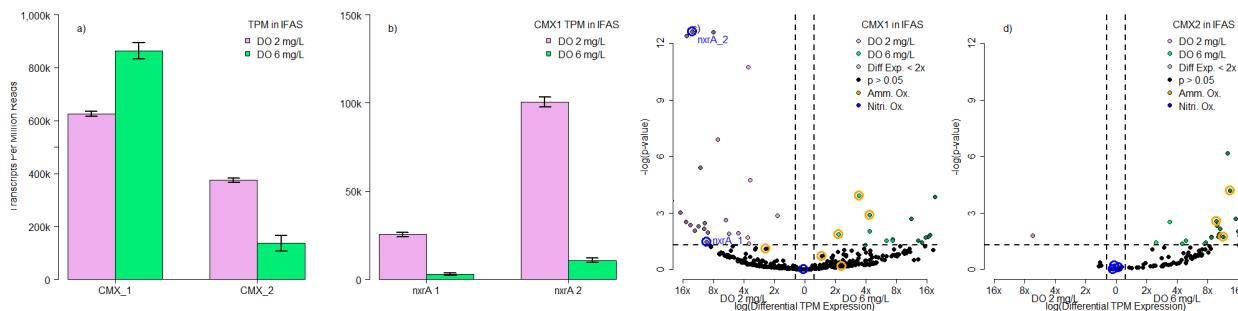
447 Two CMX MAGS revealed a significant shift in metabolic preferences based on dissolved
 448 oxygen concentration (**Figure 5**). CMX_1 was overall more active than CMX_2 comprising
 449 $62.5\% \pm 0.9\%$ of transcripts-per-million at DO = 2 mg/L and $86.4\% \pm 3.0\%$ at DO = 6 mg/L
 450 whereas CMX_2 comprised $37.5\% \pm 0.9\%$ at DO = 2 mg/L and $13.6\% \pm 3\%$ at DO = 6 mg/L
 451 respectively (**Figure 5a**). Further, genes associated with ammonia oxidation were significantly
 452 upregulated at 6 mg/L DO, while genes associated with nitrite oxidation were significantly
 453 upregulated at 2 mg/L DO.

454 Both MAGS, CMX_1 in **Figure 5c** and CMX_2 in **Figure 5d**, showed significant increases
 455 in transcription of genes involved in nitrite oxidation pathways at lower DO concentrations in IFAS
 456 biofilm, further reinforcing *nxrB* increases shown in **Figure 3b**. Interestingly, while this analysis
 457 aligns with *nxrB* RT-qPCR results (**Figure 2e**), the relative downregulation of *amoA* in CMX gene
 458 expression conflicts with *amoA* transcript quantification in CMX via RT-qPCR (**Figure 2f**) which

459 showed an increase in expression. This is likely a result of transcripts-per-million being a relative
460 abundance metric whereas RT-qPCR is quantitative. At low DO levels, CMX increases expression
461 of genes involved with both ammonia oxidation and nitrite oxidation, but those associated with
462 nitrite oxidation have a much larger increase resulting in the expression of genes associated with
463 ammonia oxidation appearing to decrease using relative abundance metrics.

464 CMX_1 had sufficient mapped transcripts to compare differential expressions between the
465 two copies of *nxrA* within the genome. The second copy of *nxrA* had significantly more activity
466 than the first copy of *nxrA* (Tukey p << 0.05). Furthermore, the ratio of expression between the
467 second copy and the first copy was constant at 4.14x copies and no significant differences in the
468 ratio based on dissolved oxygen concentration (Tukey p = 0.739). This suggests that there are
469 transcription regulations for these two *nxr* pathways. Different regulation for the paralogs has been
470 previously observed in cultures⁷⁵, but not previously reported in a full-scale wastewater treatment
471 system.

472 Interestingly, the *nxr* pathway is also a reversible pathway, which can either oxidize nitrite
473 or reduce nitrate in anaerobic conditions⁷⁶. This dynamic further reinforces the synergistic
474 relationship shown in **Figure 3** where CMX are likely supplying AMX nitrite in IFAS biofilm as
475 previously reported by Vilardi et al⁴. In aerobic conditions CMX could perform ammonia oxidation
476 and supply nitrite to AMX. In anaerobic conditions, the *nxr* pathway could be reversed for nitrate
477 reduction to resupply nitrite to AMX. This scenario could explain why CMX significantly
478 increases ammonia oxidation and greatly decrease *nxr* expression in IFAS biofilm, whereas strict
479 NOB do not significantly increase *nxrA* expression in IFAS biofilm phase (**Supplemental Figure**
480 **6c**). Furthermore, there is an overall decrease in *nxrB* expression in IFAS biofilm compared to
481 suspended sludge biomass (**Figure 2e**) since CMX dominate the limited *Nitrospira* community in
482 IFAS biofilm and strict NOB dominate the suspended sludge biomass. While others have reported
483 differential expression of *amo* and *nxr* in CMX^{68,69}, this would be the first report suggesting DO
484 role and redox zonation promoting differential expression.



485
486 **Figure 5: Comammox Nitrospira had two dominant MAGs in IFAS biofilm (Figure 5a)** which were further analyzed using
487 volcano plots for differential gene expression analysis in IFAS biofilm. The more abundant MAG, CMX_1, saw significant
488 changes in the expression between the two copies of *nxrA* within the genome (Figure 5b). Each MAG was analyzed with
489 Deseq2 for differential gene expression analysis in IFAS biofilm with c) CMX_1 and d) CMX_2. Points in pink were
490 statistically significantly more active and expressive at DO = 2 mg/L while points in green were statistically significantly
491 more active and expressive at DO = 6 mg/L. Blue circles represent shifts in nitrite oxidation activity, while the orange circles
492 represent shifts in ammonia oxidation activity. In instances where the -log(p-value) exceeded 12, they are plotted as 12.5
493 with a jitter for spacing.

494

495 *3.6 Implications for development of a CMX-AMX system for mainstream nitrogen removal and*
496 *future research.*

497 The cooperation of CMX and AMX in IFAS biofilms at lower DO concentrations has
498 significant beneficial implications for wastewater treatment facilities. Such coupling effectively
499 doubles as both cost saving due to lower aeration energy consumption⁷⁷, and decarbonization by
500 potentially minimizing the release of nitrous oxide during nitrification⁵. Further, integrating an
501 IFAS biofilm component within conventional nitrification-denitrification process does not require
502 significant infrastructure changes and can be accomplished by integrating carrier media within
503 existing reactors. The metabolic flexibility of CMX serves three critical roles benefiting AMX by
504 1) aerobically converting ammonia to nitrite for AMX's anaerobic ammonia oxidation 2)
505 potentially anaerobically converting nitrate back to nitrite for further AMX anaerobic ammonia
506 oxidation and 3) utilizing oxygen via aerobic ammonia oxidation and nitrite oxidation to create
507 anaerobic zones for AMX. While this study demonstrates that nitrifying activities of CMX and
508 their co-operation with AMX are sensitive to DO concentrations, it also opens up further research
509 questions and needs for process optimization. What is the optimal DO concentration to maximize
510 anaerobic ammonia removal considering the more pronounced upregulation of genes associated
511 with nitrite oxidation as compared to ammonia oxidation for CMX bacteria at lower DO
512 concentrations? Specifically, do CMX bacteria start competing with AMX for nitrite below a
513 certain DO setpoint despite substantially lower nitrite affinities⁴. Further, the conditions critical for
514 establishment and maintenance of AMX and CMX co-operation in a full-scale mainstream system
515 are as yet unknown since our work was conducted on already established nitrifying community.

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528 **Data Availability**

529 All code for amplicon sequencing, metatranscriptomic analysis, statistical analysis and plots are
530 available at github.com/queermsfrizzle/CMX_AMX_transcriptomics. Raw fastq files for
531 metatranscriptomic sequencing and amplicon sequencing are available via NCBI bioproject
532 submission number PRJNA1050761.

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