

1 **Genomic-guided conservation actions to restore the most**
2 **endangered conifer in the Mediterranean Basin**

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12 **Abstract**

13 Species with extremely small population sizes are critically endangered due to reduced
14 genetic diversity, increased inbreeding, and the added threat of hybridization. Genomic tools
15 significantly advance conservation by revealing genetic insights into endangered species,
16 notably in monitoring frameworks. Sicilian fir is the most endangered conifer in Europe with
17 only 30 adult trees spread across an 84-hectare area. Using 20,824 SNPs from RAD-seq
18 employing the silver fir genome assembly and a custom 120 SNP-array, we evaluated genetic
19 diversity, mating patterns, and effective population size in adult trees, 118 natural seedlings,
20 and 2,064 nursery seedlings from past conservation actions. We assessed introgression from
21 neighboring non-native fir plantations and established an intra-population assisted gene flow
22 program selecting the most genetically dissimilar individuals and investigating the outcome
23 through simulations. Genomic analysis unveiled significant genetic diversity among adult
24 Sicilian firs, comparable to non-endangered Mediterranean firs with larger populations.
25 However, the genetic diversity of the forthcoming generation declined due to high self-
26 fertilization, leading to marked inbreeding ($F_{is} = 0.38$) and an alarmingly low effective
27 population size ($N_e = 6$). Nursery seedling monitoring revealed similar selfing rates but
28 significant introgression (~50%) from non-native firs. Although intra-population assisted gene
29 flow could help to mitigate genetic loss, it may not alleviate the species vulnerability to
30 imminent environmental challenges, perpetuating the risk of an extinction vortex. Hence,
31 investigating the impact of Sicilian fir population decline and selfing on inbreeding
32 depression, along with exploring the potential of hybrids for genetic load alleviation and
33 future adaptation, is crucial for effective conservation strategies. This study stands as a
34 compelling model for guiding conservation strategies in similarly imperiled species
35 characterized by extremely small populations.

36 **Introduction**

37 Over the past few centuries, many species of animals and plants have become extinct,
38 while many others teeter on the brink of extinction due to habitat loss, fragmentation, and the
39 reduction of population size to inviable levels (Palomares et al. 2012; Benazzo et al. 2017;
40 Humphreys et al. 2019). Species with extremely small populations present a pressing
41 challenge in contemporary biodiversity conservation, referred to as the 'small population
42 paradigm' (Norris 2004). These species face an increased risk of extinction due to a
43 confluence of factors, including reduced genetic diversity, increased inbreeding, vulnerability
44 to stochastic events, and susceptibility to environmental pressures (Frankham 2015; Ralls et
45 al. 2018). Hybridization presents an additional peril, endangering the species integrity and
46 exacerbating the risks for its already impoverished populations (Balao et al. 2015). To persist
47 at short-term without facing these detrimental effects, it is imperative to maintain an effective
48 population size above 100 individuals (i.e., the minimum viable population sensu Frankham
49 et al. 2014). Consequently, genetic management has become an essential tool to identify
50 strategies for preserving their genetic integrity, mitigating inbreeding depression, and
51 fostering the adaptive potential needed for their survival and long-term persistence in their
52 respective ecosystems (Hogg et al. 2022). Furthermore, genomic tools facilitate the
53 monitoring of current conservation status and the assessment of outcomes from conservation
54 actions (Flanagan et al. 2018; Humble et al. 2023).

55 Assisted gene flow (AGF) replacing natural gene flow by guided-outcrossing can thus
56 be a powerful conservation tool to preserve biodiversity, particularly for populations that are
57 experiencing genetic erosion (Pregler et al. 2023). Restoring gene flow into small, isolated
58 populations can increase genetic diversity and fitness (i.e., genetic rescue) reducing the risk of
59 inbreeding depression and ultimately population extinction and, at the same time, accelerate
60 the adaptive potential of species to environmental modifications such as climate change

61 (Aitken & Whitlock 2013; Castilla et al. 2019). Despite the benefits and risks of this practice
62 have been extensively discussed (Frankham 2016; Ralls et al. 2020; Grummer et al. 2022), the
63 practical use of the AGF in the tree conservation remains largely unexplored (Aitken et al.
64 2015; Browne et al. 2019).

65 Conifers, such as pines and spruces, have specific conservation needs due to their slow
66 growth, long reproductive cycles, low mutation rates, and specific breeding, pollination, and
67 dispersal systems, requiring a meticulous and sustained approach to population restoration
68 (Givnish 1980; Savolainen & Pyhäjärvi 2007; De La Torre et al. 2017). *Abies nebrodensis*
69 (Lojac.) Mattei (Sicilian fir) is the most threatened conifer in the Mediterranean Basin, and
70 likely one of the most endangered species in the world. It is categorized as “critically
71 endangered” by the IUCN Red List (Thomas 2017), listed in Appendix I of the Bern
72 Convention and considered a priority species in Annexes II and IV of the 92/43 EC Habitats
73 Directive (Code 9220). This species constitutes an example of a highly vulnerable species due
74 to its very low population size, with only 30-adult trees and 118 seedlings (over 5-yr old)
75 growing in a sole population over approximately 84 hectares within the integral reserve of
76 Parco delle Madonie in Sicily (Venturella et al. 1997). Previous studies suggested that
77 reproductive trees maintain high genetic diversity but typically exhibit reduced seed viability
78 (Vicario et al. 1995; Vendramin et al. 1996; Parducci et al. 2001; Scialabba 2019). This low
79 crop viability strongly indicates the existence of high rates of inbreeding. Moreover, fir
80 species show a pronounced tendency for hybridization (Klaehn & Winieski 1962; Balao et al.
81 2020). The presumed hybridization between the Sicilian fir and the non-native silver and
82 Greek firs (*A. alba* and *A. cephalonica*, respectively) from the surrounding areas of the natural
83 population may adversely impact seed viability and species integrity (Conte & Cristofolini
84 2003; Scialabba et al. 2005).

85 Many *in-situ* and *ex-situ* conservation efforts have been implemented over the last two
86 decades, encompassing breeding programs generating thousands of seedlings via natural
87 pollination and the removal of non-native fir species from the vicinity of the population
88 (Venturella et al. 1997; Frascella et al. 2022; Rogatis et al. 2022). However, there is scarce
89 genetic information available on both the natural occurring and the nursery-propagated
90 seedlings, hindering the assessment of a viable long-term conservation plan. To address this
91 limitation, we conducted a genome-wide analysis of Sicilian fir genetic variation and placed
92 our results in the context of genetic diversity observed in other Mediterranean fir species. We
93 then developed an SNP-array for genetic monitoring of the Sicilian firs. We evaluated the
94 genetic diversity and relationships among the 30 adult trees in the natural population. Through
95 simulations, we designed an AGF program that prioritizes crosses using individuals with
96 greater genetic dissimilarity, with the goal of ensuring the short-term survival. We determined
97 the parentage origins of 118 naturally regenerated seedlings, along with 2,064 seedlings
98 cultivated in the forest nursery 'Piano Noce,' aiming to ascertain outcrossing rates, and
99 occurrences of self-fertilization. Finally, we assessed the degree of introgression with
100 neighboring plantations of non-native firs.

101 **Materials and methods**

102 *Abies species sampling and RAD-seq libraries*

103 We used restriction site associated DNA sequencing (RAD-Seq) to identify high-
104 quality and information-rich SNPs for genotyping of Sicilian firs and detect hybridization
105 with alien silver and Greek firs. We sampled five individuals from the unique population of
106 Sicilian fir, 15 individuals from three Italian silver fir populations and 15 individuals from
107 three populations of Greek fir (Supplementary Material; Table S1). We extracted DNA and
108 prepared RAD libraries following Balao et al. (2020). Genomic DNA was digested by using

109 high-fidelity SbfI (New England Biolabs) and the resulting fragments were double barcoded.
110 The library was sequenced in a separate lane of an Illumina flowcell HiSeq 2500 at the VBCF
111 NGS Unit (www.vbcf.ac.at/ngs) as 100 bp single-end (SE) reads.

112 *SNP identification and population genomics*

113 Quality filtering and demultiplexing of the RAD-seq library were performed with
114 deML (Renaud et al. 2015) and STACKS ver. 2.53 (Catchen et al. 2011, 2013, Rochette et al.,
115 2019). Following this, we aligned each raw read fastq file to the silver fir genome (Mosca et
116 al 2019), using Bowtie2 (Langmead & Salzberg 201) with --sensitive and --no-unal settings.
117 Single-best alignments were then sorted and converted from SAM to BAM format using
118 Samtools v1.10 (Li et al., 2009). To assemble RAD loci, we used the ref_map.pl pipeline
119 implemented through Stacks v.2.53. Then, we utilized the populations script to export SNP
120 data into various formats for subsequent analyses. We investigated the nucleotide diversity
121 (π), heterozygosity (H_s) and inbreeding (F_{IS}) of the Sicilian fir and compared to silver and
122 Greek firs populations.

123 *OpenArray development for genotyping and hybrid detection*

124 We customized a panel of 120 SNPs by PCR-based OpenArray Technology (Thermo
125 Fisher Scientific, USA) for genotyping the Sicilian firsamples. Firstly, we selected 100 highly
126 informative SNPs to assess the genetic structure of the natural population as well as the origin
127 of the seedling from the forest nursery ‘Piano Noce’. These unlinked high-quality SNPs with
128 a maximum information for genotyping and paternity exclusion were selected using a
129 combination of Stacks, Plink2 v. 2.00a3.7 (Chang et al., 2015), PopLDdecay 3.40 (Zhang et
130 al., 2019) based on the following criteria: ‘Biallelic’, ‘No other SNPs within 4 kb’, ‘ $r^2 < 0.01$ ’,
131 ‘Minor allele frequency [MAF] > 0.05’, ‘Minimum 1 individual heterozygote’, ‘Minimum 1

132 individual homozygote for alternative alleles’, ‘100 bp flanking sequence available on both
133 sides of the variable position’. To minimize the probability of SNPs being affected by natural
134 selection, we kept the SNPs in intergenic regions (avoiding 1kb downstream and 2 kb
135 upstream regions of genes) using SnpEff and SnpSift v4.3.1t (Cingolani et al., 2012).
136 Secondly, we selected 20 high-quality SNPs (unlinked markers with high MAF) to detect
137 hybridization between the Sicilian fir and the other firs inhabiting the Madonie Natural Park.
138 These selected makers followed the same selection criteria previously described and showed
139 the highest species-specific diagnostic power following the highest loadings in the first two
140 axes of a Discriminant Analysis of Principal Components (DAPC) performed with the
141 *adegenet* package v. 2.1.10 (Jombart, 2008) in R v.4.0.3 (R Core Team 2020). A Principal
142 Component Analysis (PCA) using the R package *dartR* 2.7.2 (Mijangos et al., 2022) was used
143 to corroborate the discriminant power of the 20 selected SNPs. We customized a panel of 120
144 SNPs by PCR-based OpenArray Technology (Thermo Fisher Scientific, USA) for genotyping
145 Sicilian fir samples. Firstly, we selected 100 highly informative SNPs to assess the genetic
146 structure of the natural population as well as the origin of the seedling from the forest nursery
147 ‘Piano Noce’. These unlinked high-quality SNPs with a maximum information for genotyping
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151 ‘Minimum 1 individual heterozygote’, ‘Minimum 1 individual homozygote for alternative
152 alleles’, ‘100 bp flanking sequence available on both sides of the variable position’. To
153 decrease the probability of selecting SNPs under selection, we kept the SNPs in intergenic
154 regions (avoiding 1kb downstream and 2 kb upstream regions of genes) using SnpEff and
155 SnpSift v4.3.1t (Cingolani et al. 2012). Secondly, we selected 20 high-quality SNPs (unlinked

156 markers with high MAF) to detect hybridization between the Sicilian fir and the other firs
157 occurring at the Madonie Regional Natural Park. These selected makers followed the same
158 selection criteria previously described and showed the highest species-specific diagnostic
159 power following the highest loadings in the first two axes of a Discriminant Analysis of
160 Principal Components (DAPC) performed with the *adegenet* package v. 2.1.10 (Jombart &
161 Ahmed 2011) in R v.4.0.3 (R Core Team 2020). A Principal Component Analysis (PCA)
162 using the R package *dartR* 2.7.2 (Gruber et al. 2018) was used to corroborate the discriminant
163 power of the 20 selected SNPs.

164 *SNP-Array validation and Genotyping of plants from the natural population*

165 To investigate the genetic diversity and relatedness of Sicilian fir individuals, we first
166 collected leaf material for genetic analysis from the 30 adult trees and the 118 seedlings
167 inhabiting the population. DNA was isolated using the NucleoMag Plant kit (Macherey-
168 Nagel, Germany) according to the manufacturer's protocol. Before starting the genotyping of
169 samples, we initially validated the specially developed 120-SNPs arrays for this species. To
170 do so, we genotyped 24 samples (12 in duplicate) and we assessed replicability and genotype
171 accuracy, as well as we calculated the genotyping error rate and allele drop-out. Then, we
172 carried out a PCA to verify that replicates clustered together.

173 We used the 120 SNPs panel to genotype the individuals in the natural population. We
174 explored the population genetic structuring of the natural population using Structure v. 2.3.4
175 (Pritchard et al. 2000). We estimated the number of genetic clusters (K) by assigning
176 individuals in undefined mixture clusters under a Bayesian framework. We conducted 10
177 independent runs of 1,000,000 iterations each one, with a burn-in period of 100,000 for each
178 value of K from 1 to 5. Then, the best K that fit the data was calculated using the Δk method

179 (Evanno et al. 2005) in Structure Harvester (Earl & VonHoldt 2011). Finally, the effective
180 population size (N_e) and inbreeding level (F_{IS}) were estimated using the software Colony
181 v.2.0.6.6 (Jones & Wang 2010).

182 We inferred the pedigree relationships in the natural population of *A. nebrodensis*,
183 using Colony2 with the full-likelihood approach. Our analysis incorporated prior information
184 on parental (all reproductive adults) and maternal genotypes (the closest adult to the seedling).
185 We assumed monoecy, diploidy, both male and female polygamy, and potential inbreeding.
186 For the full likelihood calculation, we employed very high precision setting, very long run
187 length, sibship scaling, updated allele frequencies, and weak sibship priors. The genotyping
188 error rate (estimated from replicates) was set to 0.01 for all loci. We performed three runs
189 with different random number seeds to check the reliability of the results. Additionally, the
190 seedlings with a putative hybrid origin (i.e., with an uncertain genetic origin) were examined
191 with a PCA using the 20 SNPs designed for the discrimination of fir species.

192 For the adult trees, we used the *adegenet* package to estimate the genetic diversity at
193 the individual level by calculating the following coefficients: *HsExp*, standardized
194 heterozygosity based on the mean expected heterozygosity; *HL*, homozygosity by loci; and
195 *INBR*, inbreeding coefficient. Additionally, we used Colony2 to estimate the effective
196 population size (N_e) and inbreeding level (F_{IS}). To assess whether genetic information could
197 be used in management strategies to mitigate the alleged low population diversity/inbreeding,
198 we studied the population genetic structure and relatedness of the 30-adult-trees to design a
199 genome-informed assisted gene flow (AGF) program. To do so, we used the *adegenet*
200 package to perform a DAPC using a *K*-mean clustering ($K = 1-30$, the total number of adult
201 trees). Additionally, we used the *related* package (Pew et al. 2015) to calculate the Ritland
202 estimator (RIT) of pairwise co-ancestry (Ritland 1996) to infer the genetic relatedness of all

203 adult trees. Finally, considering the absence of other Sicilian fir populations for outcrossing,
204 we selected the 12 pairs of adults with the lowest RIT values (i.e., those individuals
205 genetically more dissimilar) to simulate an artificial population in the AGF conservation
206 program. We used the function *hybridize* from the *adegenet* package to generate 10 progenies
207 from each of the 12 proposed crosses. Then, we compared the genetic diversity, inbreeding
208 and effective size of adults, natural seedlings, the simulated AGF population (n = 120) and the
209 AGF next generation population, which is composed by the 120 seedlings from the AGF
210 population, the 118 natural seedlings and 19 seedlings from the nursery originate by
211 outcrossing (see Results section).

212 *Pedigree analysis of seedlings from the forest nursery*

213 We investigated the genetic diversity and pedigree from 2,064 seedlings at the forest
214 nursery from eight maternal batches from the nursery. To assess a feasible parentage with any
215 of the 30-adult trees and detect putative hybrids with other alien firs, we genotyped the
216 samples with the 120 SNP panel and then we conducted paternity tests and determined the
217 rate of outcrossing, inbreeding and self-fertilization using Colony2, as described above.

218 **Results**

219 *Genomic diversity of the Sicilian fir and relatives*

220 The RAD-seq of the 35 individuals from populations of Sicilian, silver and Greek firs
221 produced a final STACKS catalogue a total of 365,582 RAD loci (composed of 34,325,726
222 genomic sites) with an average coverage per sample (\pm s.d.) of 51.4 ± 38.5 reads per locus.
223 After filtering for polymorphic RAD loci with $MAF > 0.05$, we obtained a final data set with
224 20,824 SNPs. The raw data was deposited in the NCBI Short Reads Archive (BioProject ID
225 PRJNA563575).

226 Estimates of genetic diversity in the populations were notably consistent across the
227 three fir species (Figure 1). Although the Sicilian fir exhibited the lowest values for π and H_s ,
228 these differences were not statistically significant when compared to the populations of the
229 other two firs (Kruskal-Wallis tests with $p > 0.05$). The Sicilian fir population showed similar
230 F_{IS} to those of the silver (0.07 ± 0.014) and Greek (0.06 ± 0.007) firs. Furthermore, it is
231 noteworthy that the Sicilian fir displayed the highest number of private alleles, with 1029,
232 surpassing the populations of the other species which ranged from 483 to 728 (Figure 1). The
233 pairwise F'_{ST} values varied between 0.07 (silver fir pop1- silver fir pop2) and 0.26 (Sicilian fir
234 – Greek fir pop 2). Finally, the Sicilian fir showed the lowest F'_{ST} pairwise with the silver fir
235 pop1 (0.14; Table S2).

236 *Development of SNP-Array for conservation of Sicilian fir*

237 We selected 120 high-quality SNPs from the 3,536 SNPs called with $MAF > 0.05$
238 present in the five individuals of the Sicilian fir, remaining only 459 SNPs after the quality
239 filtering. Functional annotations showed that 351 SNPs were in intergenic regions, and 268
240 SNPs remained after excluding promoters genes regions (46 downstream and 59 upstream).
241 The 20 SNPs selected to discriminate putative hybrids maximized the separation of the three
242 fir species (the first two components of the PCA accounted for 42.5% and 20.8% of the total
243 variation, respectively; Figure S1). The validation of the designed 120 SNP-array revealed a
244 replicability higher than 99% and the PCA successfully separated samples and clustered
245 duplicates together. Call rates (i.e., proportion of samples that were assigned a genotype call
246 compared to the total number of samples) of the analyzed samples ranged from 64.6% to
247 77.1%. High genotyping failures ($> 50\%$) were detected at 15 loci and were discarded for
248 downstream analyses.

249 *Gene diversity and population structure in the natural population*

250 The STRUCTURE analysis showed that the most likely number of genetic groupings
251 was $K = 2$ (Figure 2) with no clear clustering pattern. The first cluster comprised three adult
252 trees and 41 out of the 42 seedlings related to one of these trees. The second cluster included
253 the remaining adult trees and seedlings from the population. The effective population size was
254 very low ($N_e = 6$; 3–21; lower and upper confidence interval at 95%) and the level of
255 inbreeding was moderate ($F_{IS} = 0.373$).

256 The results of the paternity assignments disclosed the potential hybrid origin in nine
257 out of 118 seedlings from the natural population. The genetic origin of three of them could
258 not be assigned to any of the 30-adult trees occurring in the population, and for the remaining
259 six seedlings only one parental was assigned. 109 seedlings (92.4%) were identified as
260 purebred Sicilian firs. Of those, 103 seedlings originated by self-fertilization and only six of
261 them resulted from outcrossing (only five out of the 30 adult-trees were involved in
262 outcrossing). We analyzed in the PCA the nine seedlings with a putative hybrid origin,
263 finding that only one of them was genetically closer to the cloud point of the silver fir and,
264 therefore, it was confirmed as a hybrid (Figure 3).

265 For adult trees, the estimations of parental similarity showed mild genetic diversity at
266 the individual level. The mean value of the standardized observed heterozygosity ($HsExp$)
267 was 0.74 ± 0.19 , varying from 0.32 to 1.13. The individual homozygosity by loci (HL) ranged
268 from 0.50 to 0.86, with an average value of 0.67 ± 0.09 . Finally, the inbreeding coefficient
269 ($INBR$) showed values that ranged from -0.20 to 0.79, with an average value of 0.21 ± 0.23 .
270 DAPC analysis conclusively found three genetic clusters with almost lacking genetic
271 admixture that were comprised of 12, 11 and seven individuals, respectively (Figure S2A).
272 Co-ancestry analysis using *RIT* pairwise estimations revealed overt differences of the genetic

273 relatedness of the 30-adult trees, ranging from -0.3837 to 1.0587 (a matrix of pairwise co-
274 ancestry among all adult trees is depicted in Figure S2B). Finally, the Mantel test showed
275 absence of any spatial correlation between geographical distance and genetic diversity from
276 adult trees ($r = -0.06$; $p > 0.05$).

277 *Gene diversity and population structure in seedlings from the forest nursery*

278 We were able to genetically characterize 1,776 of the 2,064 selected seedlings from
279 the nursery. Paternity tests confidently assigned 1,525 seedlings (85.8%) to one or more
280 parents of the Sicilian fir. We identified 897 purebred seedlings from this species with a very
281 high autogamy rate (97.9%), only 19 of them being the result of outcrossing. Additionally, we
282 identified 879 seedlings as putative hybrids. Finally, the effective population size of the
283 seedlings in the nursery was very low ($N_e = 12$; 6–26; lower and upper confidence interval at
284 95%) and its level of inbreeding was moderately high ($F_{IS} = 0.354$).

285 *Simulating the effects of the Assisted Gene Flow*

286 RIT values of 12 pairs of adult trees selected to simulate an artificial AGF population
287 ranged from -0.3837 to -0.2843. The simulated AGF population from the selected crosses
288 showed higher H_o (Kruskal-Wallis's test, $\chi^2 = 91.882$, $df = 3$, $p < 0.001$) and lower
289 inbreeding coefficient ($\chi^2 = 179.82$, $df = 3$, $p < 0.001$) than the natural population (i.e.,
290 adults and seedlings) and the next generation population (including natural seedlings,
291 seedlings from the simulated AGF event, and the 19 outbreed seedlings from the nursery).
292 This latter population maintained similar values than current adults in the population whereas
293 natural seedlings showed significantly the lowest H_o and the highest F_{IS} (Figure 4). The
294 nucleotide diversity was similar for all the groupings ($\chi^2 = 1.984$, $df = 3$, $p > 0.05$). Natural

295 seedlings showed the lowest N_e (95% IC 2.6-2.9) and the adults the highest one (18-25.2).
296 AGP seedlings and the Next-Gen population displayed intermediate N_e values (Figure 4).

297 **Discussion**

298 *Differential genetic diversity across age cohorts of the Sicilian fir*

299 The genome-wide levels of genetic diversity among adult individuals in the Sicilian fir
300 population were comparable to those of the silver and Greek firs populations along their
301 native ranges. The significant genetic variation observed in adult Sicilian fir samples,
302 highlighted in prior studies (Vicario et al. 1995; Ducci et al. 1999; Conte et al. 2004),
303 correlated with the recent demographic dynamics of the Sicilian fir, as evidenced by the
304 increased private alleles, slower heterozygosity response, and stable inbreeding coefficients
305 (Llorens et al. 2018). Palaeobotanical and humanistic studies revealed the massive presence of
306 firs along Sicily from Paleogene to Holocene (Tinner et al. 2016), along with a significant
307 demographic decline since the Middle Ages (Pasta et al. 2020). Anthropogenic pressures
308 nearly decimated the Sicilian fir forests, driving the species to the verge of extinction, with
309 only a few surviving individuals comprising the current population (Venturella et al. 1997).
310 The extended lifespan of firs (> 150 years) and adaptive selection against homozygous
311 individuals could counteract the effects of demographic decline on genetic diversity (Yao et
312 al. 2007; Fraser et al. 2014; Su et al. 2018).

313 Despite the high genetic diversity observed in adult trees, genetic monitoring of
314 seedlings in the natural population using a developed SNP-array revealed a concerning
315 scenario for species conservation. The seedlings exhibited a substantial decrease in
316 heterozygosity and an increased inbreeding due to exceptionally high selfing rates (> 94%).
317 Only six outcrossed seedlings were found, which were traced back to just five parental
318 genotypes. This contrasts sharply with the typically high levels of outcrossing observed in

319 Mediterranean firs and other gymnosperms (Restoux et al. 2008), and it is likely influenced
320 by several factors such as the sparsely tree dispersion, with an average distance of 621.7
321 meters between them, and a limited pollen dispersal as occurs in other endangered
322 Mediterranean firs (Arista & Talavera 1994; Sánchez-Robles et al. 2014). The high selfing
323 rate and subsequent decline in genetic diversity across young cohorts of the Sicilian fir likely
324 account for the previously reported reduced seed and seedling viability (Conte et al. 2004;
325 Scialabba 2019; Mirabile et al. 2023). This reduced genetic diversity may also increase the
326 species' vulnerability to emerging environmental challenges, particularly those associated to
327 climate change (Frankham 2015; Ralls et al. 2020).

328 *Remnants of recent hybridization in Sicilian fir population and implications in the*
329 *conservation*

330 It is remarkable that, despite a high level of selfing, at least one seedling in the Sicilian
331 fir population likely derived from hybridization with an alien firs species. Previous
332 conservation efforts identified more than 5,000 such firs in reforested areas located within a
333 few hundred meters of the Sicilian fir population (Ducci 2014). While many of these non-
334 native firs were removed, some still persist, presenting a real risk of genetic exchange,
335 particularly the silver fir, the ancestral species that gave rise to the Sicilian fir (Balao et al.
336 2020). Hybridization is a widespread phenomenon present in evolutionary history of
337 Mediterranean firs (Krajmerová et al. 2016; Balao et al. 2020). While hybridization can
338 potentially harm endangered species (Balao et al. 2015), understanding the consequences of
339 hybridization requires detailed study (Arnold 2015; Gompert & Buerkle 2016). Notably, two
340 adult Sicilian firs exhibited indications of coancestry with silver firs in the PCA, suggesting
341 historical introgression of anthropical origin due to plantations from at least one century as
342 occurred in other tree species (Vanden Broeck et al. 2005; Meirmans et al. 2014; Scotti-

343 Saintagne et al. 2023). Considering that such introgressed trees could potentially enhance the
344 adaptability towards prevailing ecological stress (Kormutak et al. 2013; Stejskal et al. 2016),
345 it is cautious to prioritize their conservation as part of the Sicilian fir recovery program
346 (Jackiw et al. 2015; VonHoldt et al. 2022; Brauer et al. 2023).

347 *Monitoring ex-situ regenerated seedlings with SNP-array confirms the prevalence of*
348 *hybridization and selfing*

349 Similarly, SNP-array analyses of nursery-grown seedlings revealed a 50% prevalence
350 of hybridization, contrasting starkly with the minimal 1-in-118 ratio in natural seedlings. This
351 raises questions about the viability of hybrids in the wild (Campbell & Waser 2001). Clearly,
352 this prevalence of hybrids in the nursery may compromise the implementation of reforestation
353 actions. However, it seems worthwhile to delve deeper into investigating the fitness and
354 adaptability of these hybrids in their natural habitat, considering various environmental
355 factors.

356 Additionally, approximately 98% of the purebred Sicilian fir seedlings in the forest
357 nursery originated from self-pollination, mirroring the selfing rates observed in the wild.
358 While the effective population size (N_e) in the nursery population, established from cones
359 collected from adult trees, slightly surpassed that of the wild population, this discrepancy
360 likely stemmed from the inclusion of hybrid seedlings. However, the strikingly low ratio of
361 effective population size to census size ($N_e/N=0.0058$) highlights an alarming genetic
362 bottleneck. This compromises the genetic robustness of the forest restoration stock, akin to
363 endangered conifers such as the Boise araucaria or the Yellow Box (Kettle et al. 2008;
364 Broadhurst 2013).

365 *Implementing within-population AGF as the last chance to avoid extinction*

366 In a similar vein, the Sicilian fir population exhibited a notably low effective
367 population size ($N_e=3-21$), falling significantly below the recommended threshold (> 100),
368 which poses considerable challenges for its short-term conservation efforts (Frankham et al.
369 2014). The high selfing rate, coupled with the low mutation rate in firs (De La Torre et al.
370 2017), indicates that relying solely on spontaneous mutations will not sufficiently enhance the
371 population genetic diversity in the short- or medium- term, emphasizing the urgent need for
372 proactive genetic management strategies (Ralls et al. 2018). Consequently, genetic rescue,
373 through upcoming outcrossing events with genetically diverse individuals, stands as the final
374 opportunity to avert extinction by mitigating inbreeding and reducing levels of homozygosity
375 (Ingvarsson 2001; Fady et al. 2020). Given the single remaining wild Sicilian fir population,
376 we propose a within-population AGF approach, selecting crosses among trees with more
377 distant co-ancestry in the population. This method appears to be effective in enhancing
378 genetic diversity and N_e , reducing inbreeding, and simultaneously mitigating potential
379 negative effects, such as outbreeding and the introduction of maladaptive alleles associated
380 with AGF (Aitken & Whitlock 2013; Grummer et al. 2022). However, the gain in genetic
381 diversity through AGF restoration appears insufficient to recover past diversity, and the long-
382 term effectiveness of this conservation action to avoid the extinction vortex remains
383 uncertain. Hence, there is a necessity to investigate the impact of Sicilian fir population
384 decline and selfing on inbreeding depression (Pečnerová et al. 2023), along with exploring the
385 potential of hybrids to alleviate genetic load and facilitate future adaptation (vonHoldt et al.
386 2018).

387 **Conclusions**

388 Genomic data revealed that adult Sicilian firs harbor high genetic diversity despite
389 their small population size, comparable even to that observed in other Mediterranean firs
390 spread across large populations. Despite this high genetic richness found in adult trees, the
391 genetic variability of the population will dramatically decrease in the upcoming generations,
392 even if human-mediated actions are taken to increase, or at least maintain, the genetic
393 diversity of the population. Intra-population assisted gene flow can contribute to preventing
394 the loss of genetic variation in the population; however, this does not seem to be sufficient to
395 break free from the extinction spiral because of the predictable species' vulnerability in
396 upcoming generations to emerging environmental challenges, such as the climate change.
397 Interspecific crosses with other intersectional fir species, namely the silver fir, could represent
398 a potential alternative strategy for the future conservation of the Sicilian fir that merits
399 exploration. Our results emphasize the relevance of genomic-guided conservation actions,
400 which can assist in identifying suitable individuals for reforestation and provide a solid
401 foundation for conservation management. We expect that this approach could serve as a
402 model to be replicated in the conservation strategies for other endangered firs, or even for
403 other conifers, around the Mediterranean region.

404 **Credit authorship contribution statement**

405 **JC del Valle:** Methodology, Investigation, Writing – original draft. **M Arista:**
406 Conceptualization, Methodology, Investigation, Funding acquisition; **C Benítez-Benítez:**
407 Methodology, Investigation; **PL Ortiz:** Conceptualization, Methodology, Investigation; **FJ**
408 **Jiménez-López:** Methodology, Investigation; **A Terrab:** Conceptualization, Methodology,

409 Investigation; **Francisco Balao**: Conceptualization, Methodology, Investigation, Writing –
410 original draft. All authors have read and agreed to the published version of the manuscript.

411 **Declaration of competing interest**

412 The authors declare that they have no known competing financial interests or personal
413 relationships that could have appeared to influence the work reported in this paper.

414 **References**

- 415 Aitken SN, Bemmels JB, Sally Aitken CN. 2015. Time to get moving: assisted gene flow of
416 forest trees. *Evolutionary Applications* **9**:271-290
- 417 Aitken SN, Whitlock MC. 2013. Assisted gene flow to facilitate local adaptation to climate
418 change. *Annual Review of Ecology, Evolution, and Systematics* **44**:367–388.
- 419 Arista M, Talavera S. 1994. Phenology and anatomy of the reproductive phase of *Abies*
420 *pinsapo* Boiss (Pinaceae). *Botanical Journal of the Linnean Society* **116**:223–234.
- 421 Arnold ML. 2015. Divergence with Genetic Exchange. Page Divergence with Genetic
422 Exchange. Oxford University Press, Oxford, UK.
- 423 Balao F, Casimiro-Soriguer R, García-Castaño JL, Terrab A, Talavera S. 2015. Big thistle
424 eats the little thistle: Does unidirectional introgressive hybridization endanger the
425 conservation of *Onopordum hinojense*? *New Phytologist* **206**:448–458. Blackwell
426 Publishing Ltd.
- 427 Balao F, Lorenzo MT, Sánchez-Robles JM, Paun O, García-Castaño JL, Terrab A. 2020.
428 Early diversification and permeable species boundaries in the Mediterranean firs. *Annals*
429 *of Botany* **125**:495–507.
- 430 Benazzo A et al. 2017. Survival and divergence in a small group: The extraordinary genomic
431 history of the endangered Apennine brown bear stragglers. *Proceedings of the National*

- 432 Academy of Sciences of the United States of America **114**:E9589–E9597.
- 433 Brauer CJ, Sandoval-Castillo J, Gates K, Hammer MP, Unmack PJ, Bernatchez L,
434 Beheregaray LB. 2023. Natural hybridization reduces vulnerability to climate change.
435 Nature Climate Change **13**:282–289
- 436 Broadhurst LM. 2013. A genetic analysis of scattered Yellow Box trees (*Eucalyptus*
437 *melliodora* A.Cunn. ex Schauer, Myrtaceae) and their restored cohorts. Biological
438 Conservation **161**:48–57.
- 439 Browne L, Wright JW, Fitz-Gibbon S, Gugger PF, Sork VL. 2019. Adaptational lag to
440 temperature in valley oak (*Quercus lobata*) can be mitigated by genome-informed
441 assisted gene flow. Proceedings of the National Academy of Sciences of the United
442 States of America **116**:25179–25185.
- 443 Campbell DR, Waser NM. 2001. Genotype-by-environment interaction and the fitness of
444 plant hybrids in the wild. Evolution **55**:669–676.
- 445 Castilla AR, Garrote PJ, Żywiec M, Calvo G, Suárez-Esteban A, Delibes M, Godoy JA, Picó |
446 F Xavier, Fedriani JM. 2019. Genetic rescue by distant trees mitigates qualitative pollen
447 limitation imposed by fine-scale spatial genetic structure. Molecular Ecology **28**:4363.
- 448 Chang CC, Chow CC, Tellier LCAM, Vattikuti S, Purcell SM, Lee JJ. 2015. Second-
449 generation PLINK: Rising to the challenge of larger and richer datasets. GigaScience
450 **4**:1–16.
- 451 Cingolani P, Platts A, Wang LL, Coon M, Nguyen T, Wang L, Land SJ, Lu X, Ruden DM.
452 2012. A program for annotating and predicting the effects of single nucleotide
453 polymorphisms, SnpEff: SNPs in the genome of *Drosophila melanogaster* strain w1118;
454 iso-2; iso-3. Fly **6**:80–92.
- 455 Conte L, Cotti C, Schicchi R, Raimondo PM, Cristofolini G. 2004. Detection of ephemeral

- 456 genetic sub-structure in the narrow endemic *Abies nebrodensis* (Lojac.) Mattei
457 (Pinaceae) using RAPD markers. *Plant Biosystems* **138**:279–289.
- 458 Conte L, Cristofolini G. 2003. Assessment of RAPD variation in *Abies nebrodensis* (Lojac.)
459 Mattei (Pinaceae) using haploid tissue analysis. *Israel Journal of Plant Sciences* **51**:199–
460 206.
- 461 De La Torre AR, Li Z, Van de Peer Y, Ingvarsson PK. 2017. Contrasting Rates of Molecular
462 Evolution and Patterns of Selection among Gymnosperms and Flowering Plants.
463 *Molecular Biology and Evolution* **24**:1586–1591.
- 464 Ducci F. 2014. Species restoration through dynamic ex situ conservation: *Abies nebrodensis*
465 as a model in Genetic considerations in ecosystem restoration using native tree species.
466 *State of the World’s Forest Genetic Resources – Thematic Study*. Bozzano M., Jalonen
467 R., Thomas E., Boshier D., Gallo L., Cavers S., Bordács S., Smith P., Loo J. eds., p. 225-
468 233.
- 469 Ducci F, Proietti R, Favre JM. 1999. Allozyme assessment of genetic diversity within the relic
470 Sicilian fir *Abies nebrodensis* (Lojac.) Mattei. *Annals of Forest Science* **56**:345–355.
- 471 Earl DA, VonHoldt BM. 2011. STRUCTURE HARVESTER: a website and program for
472 visualizing STRUCTURE output and implementing the Evanno method. *Conservation*
473 *Genetics Resources* **4**:359–361.
- 474 Evanno G, Regnaut S, Goudet J. 2005. Detecting the number of clusters of individuals using
475 the software structure: a simulation study. *Molecular Ecology* **14**:2611–2620.
- 476 Fady B, Aravanopoulos F, Benavides R, González-Martínez S, Grivet D, Lascoux M, Lindner
477 M, Rellstab C, Valladares F, Vinceti B. 2020. Genetics to the rescue: managing forests
478 sustainably in a changing world. *Tree Genetics and Genomes* **16**.
- 479 Flanagan SP, Forester BR, Latch EK, Aitken SN, Hoban S. 2018. Guidelines for planning

- 480 genomic assessment and monitoring of locally adaptive variation to inform species
481 conservation. *Evolutionary Applications* **11**:1035–1052.
- 482 Frankham R. 2015. Genetic rescue of small inbred populations: meta-analysis reveals large
483 and consistent benefits of gene flow. *Molecular Ecology* **24**:2610–2618.
- 484 Frankham R. 2016. Genetic rescue benefits persist to at least the F3 generation, based on a
485 meta-analysis. *Biological Conservation* **195**:33–36.
- 486 Frankham R, Bradshaw CJA, Brook BW. 2014. Genetics in conservation management:
487 Revised recommendations for the 50/500 rules, Red List criteria and population viability
488 analyses. *Biological Conservation* **170**:56–63.
- 489 Frascella A et al. 2022. Innovative In Situ and Ex Situ Conservation Strategies of the
490 Madonie Fir *Abies nebrodensis*. *Sustainability* **14**:12643.
- 491 Fraser DJ, Debes P V, Bernatchez L, Hutchings JA. 2014. Population size, habitat
492 fragmentation, and the nature of adaptive variation in a stream fish. *Proceedings of the*
493 *Royal Society B: Biological Sciences* **281**.
- 494 Givnish T. 1980. Ecological constraints on the evolution of breeding systems in seed plants:
495 dioecy and dispersal in gymnosperms. *Evolution* **34**:959–972.
- 496 Gompert Z, Buerkle CA. 2016. What, if anything, are hybrids: enduring truths and challenges
497 associated with population structure and gene flow. *Evolutionary Applications* **9**:909–
498 923.
- 499 Gruber B, Unmack PJ, Berry OF, Georges A. 2018. dartr: An r package to facilitate analysis
500 of SNP data generated from reduced representation genome sequencing. *Molecular*
501 *Ecology Resources* **18**:691–699.
- 502 Grummer JA, Booker TR, Matthey-Doret R, Nietlisbach P, Thomaz AT, Whitlock MC. 2022.
503 The immediate costs and long-term benefits of assisted gene flow in large populations.

- 504 Conservation Biology **36**:e13911.
- 505 Hogg CJ, Ottewell K, Latch P, Rossetto M, Biggs J, Gilbert A, Richmond S, Belov K. 2022.
- 506 Threatened Species Initiative: Empowering conservation action using genomic
- 507 resources. Proceedings of the National Academy of Sciences of the United States of
- 508 America **119**.
- 509 Humble E et al. 2023. Conservation management strategy impacts inbreeding and mutation
- 510 load in scimitar-horned oryx. Proceedings of the National Academy of Sciences of the
- 511 United States of America **120**.
- 512 Humphreys AM, Govaerts R, Ficinski SZ, Nic Lughadha E, Vorontsova MS. 2019. Global
- 513 dataset shows geography and life form predict modern plant extinction and rediscovery.
- 514 Nature Ecology and Evolution **3**:1043–1047.
- 515 Ingvarsson PK. 2001. Restoration of genetic variation lost - The genetic rescue hypothesis.
- 516 Trends in Ecology and Evolution **16**:62–63.
- 517 Jackiw RN, Mandil G, Hager HA. 2015. A framework to guide the conservation of species
- 518 hybrids based on ethical and ecological considerations. Conservation Biology **29**:1040–
- 519 1051.
- 520 Jombart T, Ahmed I. 2011. adegenet 1.3-1: new tools for the analysis of genome-wide SNP
- 521 data. Bioinformatics **27**:3070–1.
- 522 Jones OR, Wang J. 2010. COLONY: A program for parentage and sibship inference from
- 523 multilocus genotype data. Molecular Ecology Resources **10**:551–555.
- 524 Kettle CJ, Ennos RA, Jaffré T, Gardner M, Hollingsworth PM. 2008. Cryptic genetic
- 525 bottlenecks during restoration of an endangered tropical conifer. Biological Conservation
- 526 **141**:1953–1961.
- 527 Klaehn FU, Winieski JA. 1962. Interspecific hybridization in the genus *Abies*.

- 528 Kormutak A et al. 2013. Artificial hybridization of some *Abies* species. *Plant Systematics and*
529 *Evolution* **299**:1175–1184.
- 530 Krajmerová D, Paule L, Zhelev P, Voleková M, Evtimov I, Gagov V, Gömöry D. 2016.
531 Natural hybridization in eastern-Mediterranean firs: The case of *Abies borisii-regis*.
532 *Plant Biosystems* **150**:1189–1199.
- 533 Llorens TM, Ayre DJ, Whelan RJ. 2018. Anthropogenic fragmentation may not alter pre-
534 existing patterns of genetic diversity and differentiation in perennial shrubs. *Molecular*
535 *Ecology* **27**:1541–1555.
- 536 Meirmans PG, Gros-Louis MC, Lamothe M, Perron M, Bousquet J, Isabel N. 2014. Rates of
537 spontaneous hybridization and hybrid recruitment in co-existing exotic and native mature
538 larch populations. *Tree Genetics and Genomes* **10**:965–975.
- 539 Mirabile G, Cirlincione F, Venturella G, Torta L. 2023. Seed vitality and fungal
540 contamination in *Abies nebrodensis*. *Plant Biosystems* **157**:112–118.
- 541 Norris K. 2004. Managing threatened species: the ecological toolbox, evolutionary theory and
542 declining-population paradigm. *Journal of Applied Ecology* **41**:413–426.
- 543 Palomares F, Godoy JA, López-Bao JV, Rodríguez A, Roques S, Casas-Marce M, Revilla E,
544 Delibes M. 2012. Possible Extinction Vortex for a Population of Iberian Lynx on the
545 Verge of Extirpation. *Conservation Biology* **26**:689–697.
- 546 Parducci L, Szmidt AE, Madaghiele A, Anzidei M, Vendramin GG. 2001. Genetic variation
547 at chloroplast microsatellites (cpSSRs) in *Abies nebrodensis* (Lojac.) Mattei and three
548 neighboring *Abies* species. *Theoretical and Applied Genetics* **102**:733–740.
- 549 Pasta S, Sala G, La Mantia T, Bondi C, Tinner W. 2020. The past distribution of *Abies*
550 *nebrodensis* (Lojac.) Mattei: results of a multidisciplinary study. *Vegetation History and*
551 *Archaeobotany* **29**:357–371.

- 552 Pew J, Muir PH, Wang J, Frasier TR. 2015. related: An R package for analysing pairwise
553 relatedness from codominant molecular markers. *Molecular Ecology Resources* **15**:557–
554 561.
- 555 Pregler KC, Obedzinski M, Gilbert-Horvath EA, White B, Carlson SM, Garza JC. 2023.
556 Assisted gene flow from outcrossing shows the potential for genetic rescue in an
557 endangered salmon population. *Conservation Letters* **16**.
- 558 Pritchard JK, Stephens M, Donnelly P. 2000. Inference of population structure using
559 multilocus genotype data. *Genetics* **155**:945–959.
- 560 Ralls K, Ballou JD, Dudash MR, Eldridge MDB, Fenster CB, Lacy RC, Sunnucks P,
561 Frankham R. 2018. Call for a Paradigm Shift in the Genetic Management of Fragmented
562 Populations. *Conservation Letters* **11**.
- 563 Ralls K, Sunnucks P, Lacy RC, Frankham R. 2020. Genetic rescue: A critique of the evidence
564 supports maximizing genetic diversity rather than minimizing the introduction of
565 putatively harmful genetic variation. *Biological Conservation* **251**:108784.
- 566 Restoux G, Silva DE, Sagnard F, Torre F, Klein E, Fady B. 2008. Life at the margin: The
567 mating system of Mediterranean conifers. *Web Ecology* **8**:94–102.
- 568 Ritland K. 1996. Estimators for pairwise relatedness and individual inbreeding coefficients.
569 *Genetical Research* **67**:175–185.
- 570 Rogatis A De, Ducci F, Guerri S, De Rogatis A, Ducci F, Guerri S, Teani A, Proietti R. 2022.
571 Genotyping ex situ trees of *Abies nebrodensis* translocated from the original Sicilian
572 population to enrich the gene pool. *Journal of Forestry Research* **34**:1095–1106.
- 573 Sánchez-Robles JMJM, García-Castaño JLJL, Balao F, Terrab A, Navarro-Sampedro L,
574 Tremetsberger K, Talavera S. 2014. Effects of tree architecture on pollen dispersal and
575 mating patterns in *Abies pinsapo* Boiss. (Pinaceae). *Molecular Ecology* **23**:6165–6178.

- 576 Savolainen O, Pyhäjärvi T. 2007. Genomic diversity in forest trees. *Current Opinion in Plant*
577 *Biology* **10**:162–167.
- 578 Scialabba A. 2019. Seed germination in *Abies nebrodensis* (Pinaceae). *Flora Mediterranea*
579 **29**:272–276.
- 580 Scialabba A, Schicchi R, Cordi R, others. 2005. Monitoraggio e analisi del polline. Pages 94–
581 98 *Rendiconto sul progetto LIFE Natura" Conservazione in situ ed ex situ di Abies*
582 *nebrodensis* (Lojac.) Mattei. ITA.
- 583 Scotti-Saintagne C, de Sousa Rodrigues A, Roig A, Fady B. 2023. A comprehensive strategy
584 for the conservation of forest tree genetic diversity: an example with the protected *Pinus*
585 *nigra* subsp. *salzmannii* (Dunal) Franco in France. *Conservation Genetics*.
586 <https://doi.org/10.1007/s10592-023-01581-8>
- 587 Stejskal J, Horák J, Typta J. 2016. Effect of hybridization in the firs: artificial hybridization
588 may lead to higher survival rate. *European Journal of Forest Research* **135**:1097–1105.
- 589 Su J, Yan Y, Song J, Li J, Mao J, Wang N, Wang W, Du FK. 2018. Recent fragmentation
590 may not alter genetic patterns in endangered long-lived species: Evidence from *Taxus*
591 *cuspidata*. *Frontiers in Plant Science* **9**:1571.
- 592 Thomas P. 2017. *Abies nebrodensis*. *The IUCN Red List of Threatened Species:2012–2017*.
- 593 Tinner W et al. 2016. Holocene vegetation and fire history of the mountains of Northern
594 Sicily (Italy). *Vegetation History and Archaeobotany* **25**:499–519.
- 595 Vanden Broeck A, Villar M, Van Bockstaele E, Van Slycken J. 2005. Natural hybridization
596 between cultivated poplars and their wild relatives: Evidence and consequences for
597 native poplar populations. *Annals of Forest Science* **62**:601–613.
- 598 Vendramin GG, Michelozzi M, Lelli L, Tognetti R. 1996. Genetic variation in *Abies*
599 *nebrodensis*: a case study for a highly endangered species. *Forest Genetics* **2**:171–175.

- 600 Venturella G, Mazzola P, Raimondo F. 1997. Strategies for the conservation and restoration
601 of the relict population of *Abies nebrodensis* (Lojac.) Mattei. *Bocconea* **7**:417–425.
- 602 Vicario F, Vendramin GG, Rossi P, Liò P, Giannini R. 1995. Allozyme, chloroplast DNA and
603 RAPD markers for determining genetic relationships between *Abies alba* and the relict
604 population of *Abies nebrodensis*. *Theoretical and Applied Genetics* **90**:1012–1018.
- 605 vonHoldt BM, Brzeski KE, Wilcove DS, Rutledge LY. 2018. Redefining the role of
606 admixture and genomics in species conservation. *Conservation Letters* **11**.
- 607 VonHoldt BM, Hinton JW, Shutt AC, Murphy SM, Karlin ML, Adams JR, Waits LP, Brzeski
608 KE. 2022. Reviving ghost alleles: Genetically admixed coyotes along the American Gulf
609 Coast are critical for saving the endangered red wolf. *Science Advances* **8**:eabn7731.
- 610 Yao X, Ye Q, Kang M, Huang H. 2007. Microsatellite analysis reveals interpopulation
611 differentiation and gene flow in the endangered tree *Changiostyrax dolichocarpa*
612 (Styracaceae) with fragmented distribution in central China. *New Phytologist* **176**:472–
613 480.
- 614 Zhang C, Dong SS, Xu JY, He WM, Yang TL. 2019. PopLDdecay: A fast and effective tool
615 for linkage disequilibrium decay analysis based on variant call format files.
616 *Bioinformatics* **35**:1786–1788.
- 617

618 **Figure Legends**

619 **Figure 1.** Violin plots display genetic diversity estimators based in 20,824 SNPs for silver
620 (orange), Greek (light yellow), and Sicilian (green) firs. Metrics include nucleotide diversity
621 (π), heterozygosity (H_s), private alleles, and inbreeding (F_{IS}).

622 **Figure 2.** Individual ancestry at $K= 2$, the most supported numbers of clusters in Structure
623 analysis, for the 30 adult trees and the 118 seedlings in the Sicilian fir population. Adults are
624 sorted by proximity and arrows show seedlings origin from mother plants.

625 **Figure 3.** PCA plot displaying nine seedlings of uncertain origin from the natural Sicilian fir
626 population and other alien firs species. Colored dots represent Sicilian (green), silver (light
627 yellow), and Greek (orange) firs, and seedlings with ambiguous origins (red). Darker colors
628 highlight two adult trees and one seedling showing introgression.

629 **Figure 4.** Genetic population parameters based in 120 SNP-array for the adult Sicilian firs
630 (red) and seedlings (orange) from the natural population, the simulated AGF population
631 (green), and the AGF-next-generation population (blue). Violin plots display the distribution
632 of homozygosity (H_o), inbreeding (F_{IS}), nucleotide diversity (π), and effective size population
633 (N_e).

634 **Table 1.** Results of the genetic diversity estimations calculated for the 30 adult trees of the Sicilian fir.
635 *HsExp* (standardized heterozygosity based on the mean expected heterozygosity), *HL* (homozygosity
636 by loci), and *INBR* (inbreeding coefficient).

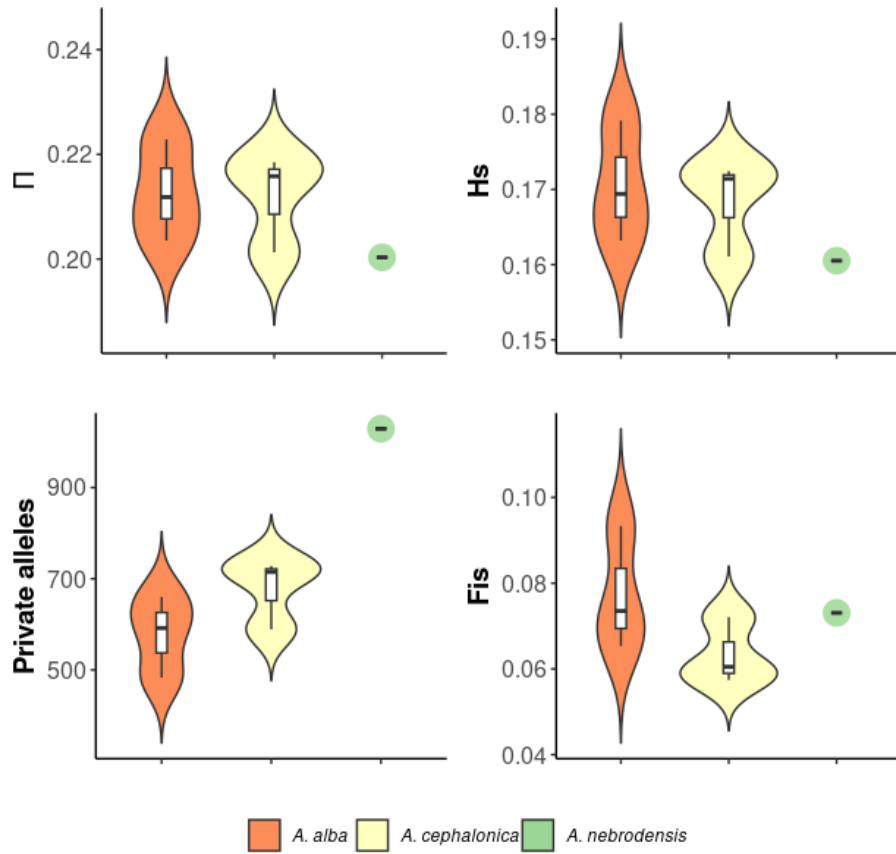
637

Adult tree	<i>HsExp</i>	<i>HL</i>	<i>INBR</i>
01M	0.60	0.74	0.25
02M	0.76	0.65	0.15
04M	0.72	0.67	0.25
06M	0.69	0.68	0.20
07M	0.69	0.69	0.20
08M	0.79	0.64	0.57
09M	0.76	0.66	0.12
10M	0.90	0.61	0.17
11M	0.83	0.61	0.11
12M	0.93	0.57	-0.01
13M	1.13	0.50	-0.20
14M	0.93	0.59	0.02
15M	0.74	0.67	0.08
16M	0.37	0.83	0.40
17M	0.83	0.63	0.10
18M	0.90	0.61	0.12
19M	0.42	0.83	0.79
20M	0.81	0.63	0.06
21M	0.58	0.75	0.55
22M	0.97	0.57	-0.02
23M	0.79	0.63	0.07
24M	0.74	0.66	0.13
25M	0.74	0.66	0.18
26M	0.93	0.58	0.04
27M	0.60	0.73	0.31
28M	0.65	0.71	0.17
29M	0.97	0.57	-0.01
30M	0.49	0.78	0.61
31M	0.32	0.86	0.65
32M	0.58	0.74	0.36

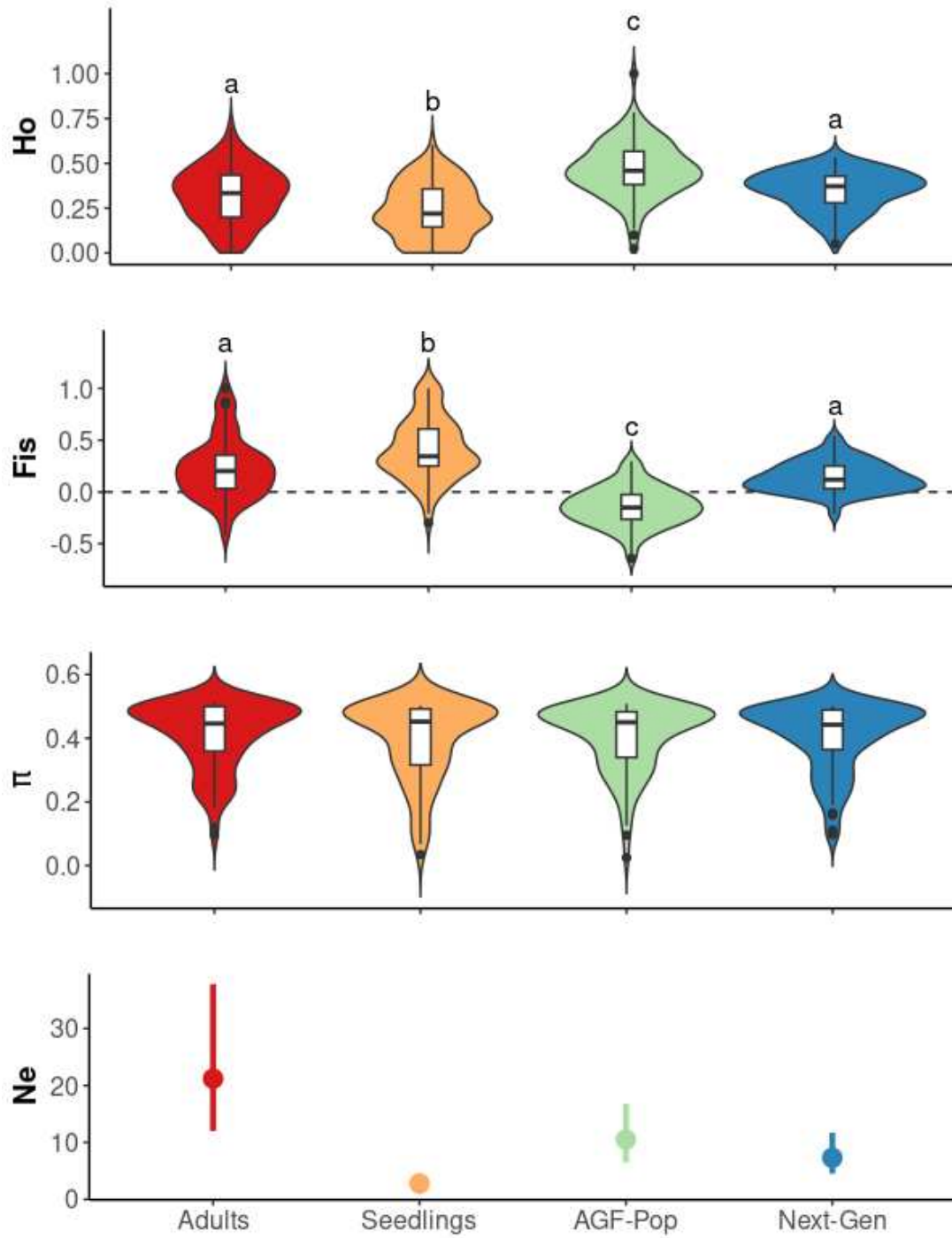
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639 **Figure 1**
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641



646 **Figure 4**



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