

# 1 Metagenomic analysis of ecological niche overlap and 2 community collapse in microbiome dynamics

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35

36 **Abstract**

37 Species utilizing the same resources often fail to coexist for extended periods of time. Such  
38 competitive exclusion mechanisms potentially underly microbiome dynamics, causing  
39 breakdowns of communities composed of species with similar genetic backgrounds of  
40 resource utilization. Although genes responsible for competitive exclusion among a small  
41 number of species have been investigated in pioneering studies, it remains a major challenge  
42 to integrate genomics and ecology for understanding stable coexistence in species-rich  
43 communities. Here, we show that community-scale analyses of functional gene redundancy  
44 can provide a useful platform for interpreting and predicting collapse of bacterial  
45 communities. Through 110-day time-series of experimental microbiome dynamics, we  
46 analyzed the metagenome-assembled genomes of co-occurring bacterial species. We then  
47 inferred ecological niche space based on the multivariate analysis of the genome  
48 compositions. The analysis allowed us to evaluate potential shifts in the level of niche overlap  
49 between species through time. We hypothesized that community-scale pressure of competitive  
50 exclusion could be evaluated by quantifying overlap of genetically determined resource-use  
51 profiles (metabolic pathway profiles) among coexisting species. We found that the degree of  
52 community compositional changes observed in the experimental microbiome was explained  
53 by the magnitude of gene-repertoire overlaps among bacterial species. The metagenome-  
54 based analysis of genetic potential for competitive exclusion will help us forecast major  
55 events in microbiome dynamics such as sudden community collapse (i.e., dysbiosis).

56

57 **INTRODUCTION**

58 Classic niche theory predicts that coexistence of species requires interspecific difference in  
59 resource use patterns [1–6]. Although some specific mechanisms can promote stable  
60 coexistence even with complete resource overlap (e.g., spatial structure of habitats and  
61 temporal variability in resource availability), similarity/dissimilarity in basic resource  
62 dependency among species is the basic factor determining the occurrence of competitive  
63 exclusion [7–9]. Therefore, evaluating the overlap of “fundamental niches”, which are defined  
64 by species’ fundamental resource requirements and resource-use capabilities [10, 11], is an  
65 essential step for understanding and predicting community-level dynamics.

66        Insights into fundamental niches are encrypted in species' genomes [12–14]: as species'  
67 traits are encoded in their DNA, genomic information provides the ultimate basis for  
68 evaluating target species' fundamental niches [15, 16]. Thus, potential strength of competitive  
69 interactions within ecological guilds or communities could be evaluated based on the  
70 distribution of species' gene repertoires within ecological niche space inferred with  
71 metagenomic data [12, 15, 16], also referred to as "metagenomic niche space". Although  
72 overlap of niches does not always cause competitive exclusion [7–9], higher levels of gene  
73 repertoire overlap within a community may impose greater impacts on population dynamics  
74 of constituent species.

75        In considering coexistence of microbial species, it is essential to examine whether such  
76 competition-driven population-level phenomena underly drastic ecological events observed at  
77 the community level. Microbial communities sometimes show sudden and substantial changes  
78 in species and/or taxonomic compositions [17–20]. Human gut microbiomes, for example,  
79 have been reported to show drastic shifts from species-rich states to "imbalanced" states with  
80 low  $\alpha$ -diversity and overrepresentation of pathogenic species [21–24] (e.g., *Clostridium*  
81 *difficile*). Elucidating the ecological mechanisms causing such drastic community-level events  
82 provide fundamental insights into microbiome dynamics [24–26]. In this respect, an important  
83 challenge is to test the hypothesis that high levels of gene-repertoire overlap are observable  
84 prior to drastic community compositional changes. However, this hypothesis, to our  
85 knowledge, has not yet been tested presumably due to the paucity of time-series observations  
86 of microbiomes with substantial compositional changes. Even if such microbiome time-series  
87 data are available, analyses of potential niche (gene repertoire) overlap require another line of  
88 information. Specifically, we need data of respective species' genomes at multiple time points.  
89 Therefore, developing research systems that can overcome these constraints will deepen our  
90 understanding of microbiome ecological processes.

91        In this study, we test the degree to which gene-repertoire overlap changes through  
92 dynamics of species-rich microbial communities. By targeting an experimental microbial  
93 system showing rapid and substantial changes in taxonomic compositions [19], we infer niche  
94 space depicting species' gene repertoires. Work in this system using a metabolic modeling  
95 analysis demonstrated that interactions between species were keys to understand the drastic  
96 microbiome dynamics [27]. Now, by compiling the shotgun metagenomic data collected at 13  
97 time points across the 110-day time-series of the experiment, we reveal temporal shifts in the  
98 magnitude of gene repertoire overlap among microbial species. We then examine whether a

99 high level of fundamental-niche overlap is observed prior to drastic changes in community  
100 structure. Overall, we explore how signs of drastic shifts in community structure are detected  
101 by inferring community-scale degree of fundamental niche overlap with the aid of genomic  
102 information.

103

104 **RESULTS**

105 **Functional dynamics of microbiomes**

106 We focused on the experimental microbiome showing drastic shifts in taxonomic  
107 compositions [19]. In a previous study [19], a 110-day monitoring of microbiomes was  
108 performed with six experimental settings. To set up experimental microbiomes with high  
109 diversity of bacterial species/taxa, we used natural microbial communities derived from soil  
110 or pond-water ecosystems as source inocula, rather than “synthetic” communities with pre-  
111 defined diversity. Specifically, microbiomes were set up with combinations of two source  
112 inoculum types (soil- or pond-water-derived inoculum microbiomes) and three medium types  
113 (oatmeal, oatmeal-peptone, or peptone broth medium) with eight replications ( $2 \times 3 \times 8 = 48$   
114 microbiomes; see Materials and Methods for details). From each of the 48 microbiomes, a  
115 fraction of each replicate community was sampled every 24 hours. The collected samples  
116 were subjected to the amplicon sequencing of the 16S rRNA region and the temporal changes  
117 in community compositions were monitored throughout the time-series [19]. By calculating  
118 the magnitude of time-series changes in community compositions [19] (Fig. 1A), we focused  
119 on a water-inoculum/oatmeal-medium replicate community showing the most abrupt (rapid  
120 and substantial) changes in community compositions among the 48 microbiomes examined as  
121 described in a study on metabolic interactions between species [27] (Fig. S1).

122 By targeting the replicate community mentioned above, we compiled shotgun  
123 metagenomic data at 13 time points across the time-series [27]. In total, 32 high-quality (>  
124 80 % completeness and < 5 % contamination) metagenome-assembled genomes (MAGs)  
125 belonging to 20 genera (16 families; 12 orders) were detected [27] (Figs. 1B-C and 2; Fig. S2;  
126 Table S1). As indicated in the amplicon sequencing analysis [19] (Fig. 1A), drastic shifts from  
127 taxon-rich community states to oligopolistic states was observed around Day 20 in the  
128 shotgun sequencing analysis (Fig. 1B).

129 After the drastic community compositional change, the system reached a quasi-stable  
130 state represented by the dominance of a *Hydrotalea* (Chitinophagaceae) bacterium (Fig. 1B).

131 The MAG of the *Hydrotalea* was characterized by relatively low GC content (38 %) and  
132 relatively small genome size within the community (ca. 3.1 Mb; Fig. 2A). In contrast, the two  
133 bacterial MAGs consistently coexisted with the dominant *Hydrotalea* through the time-series  
134 (i.e., *Terracidiphilus* and *Mangrovibacter*) had larger genome size (4.2 and 5.4 Mb,  
135 respectively; Fig. 1C), characterized by various genes absent from the *Hydrotalea* genome  
136 (Fig. 2; Fig. S3). Specifically, the *Terracidiphilus* MAG showed metabolic  
137 pathways/processes for degrading plant-derived biopolymers (e.g., cellulose; Fig. 2),  
138 potentially surviving as a primary user of polymer compounds within the plant-derived  
139 (oatmeal) medium. Meanwhile, the *Mangrovibacter* MAG had pathways/processes related to  
140 starch degradation (e.g., amylase) and vitamin-B<sub>12</sub> transportation, which were absent from the  
141 genomes of *Hydrotalea*, *Terracidiphilus*, and the other MAG (*Rhizomicrombium*) detected on  
142 Day 40-60 (Fig. 2).

143

#### 144 **Multivariate analysis of gene repertoires**

145 Next, we used the shotgun metagenomic data to evaluate how the level of gene repertoire  
146 overlap among microbes shifted through time. We anticipated that microbial species with  
147 similar resource-use abilities or restrictions have similar genomic structure. Therefore, it is  
148 expected that species competing for the same resource tend to form clusters within the space  
149 defined based on the principal coordinate analysis (PCoA) of dissimilarity in gene repertoires.  
150 For each pair of the 32 MAGs, dissimilarity (Jaccard distance) of gene repertoires was  
151 calculated based on the matrix representing the presence/absence of the 6,999 genes annotated  
152 with the program Prokka [28]. A PCoA was then performed using the  $\beta$ -diversity information  
153 (Fig. 3A). At each of the 13 time points, detected MAGs were plotted on the PCoA space.  
154 Since we did not have *a priori* knowledge of specific metabolic pathways keys to the  
155 microbe-to-microbe competition within the experimental microbiome, all datasets were  
156 included in this multivariate analysis. Given general characteristics of multivariate analysis  
157 based on  $\beta$ -diversity metrics, the multivariate reconstruction of ecological niche space  
158 depends greatly on the genes whose presence/absence profiles vary among species, while  
159 housekeeping genes possessed by most species are expected to contribute little to the  
160 multivariate analysis.

161 We found that alphaproteobacterial and gammaproteobacterial MAGs respectively  
162 constituted some clusters within the niche space reconstructed based on the multivariate

163 analysis early in the microbiome dynamics (Day 1-20; Fig. 3B). This state with high niche  
164 overlap and potential within-guild competition for resources then collapsed into a simpler  
165 community state represented by *Hydrotalea*, *Mangrovibacter*, *Terracidiphilus*, and  
166 *Rhizomicrobium* as detailed above (Fig. 3B). The space once occupied by many  
167 alphaproteobacterial and gammaproteobacterial MAGs remained unoccupied or sparsely  
168 occupied after the community compositional collapse. Even when the number of MAGs  
169 detectable with our shotgun-metagenomic sequencing increased again late in the time-series,  
170 dense aggregations of microbes with similar genomic compositions remained unobserved  
171 (Fig. 3B).

172

### 173 **Metagenomic niche overlap**

174 We next quantitatively evaluated dynamics in the magnitude of community-scale niche  
175 overlap within the multivariate space (Fig. 3). We developed two types of simple indices for  
176 evaluating community-scale niche overlap. The one is defined as the overall mean of gene-  
177 repertoire similarity between pairs of MAGs within a community. For a time point, the niche  
178 overlap index is calculated as:

$$179 \text{niche overlap score (overall mean)} = 1 - \frac{\sum_{i \in T, j \in T, i \neq j} D_{ij}}{N_T(N_T - 1)},$$

180 where  $T$  is the set of MAGs detected on a focal day (relative abundance  $> 0.1\%$ ),  $D_{ij}$  is the  
181 Jaccard metric of dissimilarity [29] in gene compositions, and  $N_T$  is the number of MAGs  
182 detected on the day. By definition, this niche overlap value based on Jaccard dissimilarity  
183 varies from 0 (completely different repertoires of genes in all pairs of MAGs) and 1  
184 (completely identical gene repertoires in all pairs of MAGs), allowing us to evaluate niche  
185 overlap levels of target communities within the standardized ranges. The other index is  
186 defined as mean value of gene-repertoire similarity with nearest neighbors. The alternative is  
187 calculated as:

$$188 \text{niche overlap score (nearest mean)} = 1 - \frac{\sum_{i \in T, i \neq j} \min_{j \in T} (D_{ij})}{N_T}.$$

189 This index can be modified by incorporating the information of the relative abundance of  
190 MAGs ( $p_i$ ) as follows:

191           niche overlap score (weighted nearest mean) =  $1 - \sum_{i \in T, i \neq j} p_i \min_{j \in T} (D_{ij})$ .

192           The results indicated that the level of niche overlap was the highest on Day 1 or Day 10  
193           and that it decreased until Day 30 (Fig. 4A). Although the niche overlap score remained low  
194           between Day 40 and 60, it increased again late in the microbiome time-series (Fig. 4B). Note  
195           that  $\alpha$ -diversity of the community showed similar temporal shifts and it was significantly  
196           associated with the niche overlap indices (Fig. 4B-C). Through the time-series, the estimated  
197           niche overlap level was significantly associated with the magnitude of the observed  
198           community compositional changes (Fig. 5)

199

## 200 **DISCUSSION**

201           By developing simple metrics of among-species overlap of gene repertoires, we examined  
202           potential relationship between community-scale niche overlap and drastic changes in  
203           community structure. Early in the experimental microbiome dynamics, alphaproteobacterial  
204           and gammaproteobacterial species were present, resulting in relatively high niche-overlap  
205           scores at the community level (Figs. 3 and 4). The quasi-equilibrium state of microbial  
206           compositions then collapsed into another quasi-equilibrium represented by a small number of  
207           bacteria varying in genome size and metabolic capabilities. Throughout the time-series,  
208           higher niche overlap levels entailed greater changes in microbial community compositions  
209           (Fig. 5). These findings lead to the working hypothesis that collapse of microbiome structure  
210           is predicted by the level of potential niche overlap within multivariate metagenomic space. In  
211           light of the “limiting similarity” rule of ecological niches [30], microbial species that exceed a  
212           critical limit of genome compositional similarity are expected to compete for the same  
213           resources intensively, eventually driving competitive exclusion processes. Thus, as examined  
214           in this study, similarity/dissimilarity in genetically determined resource-use properties (i.e.,  
215           fundamental niches) sets baselines for consequences of interspecific interactions.

216           The results also indicated that niche overlap level does not necessarily show monotonic  
217           decrease through microbial community processes. Although gene-repertoire overlap level and  
218           detectable species richness sharply declined early in the microbiome dynamics, both variables  
219           gradually increased again around Day 80 (Figs. 1A and 4B). In the resurgence process,  
220           however, the dense clusters of alphaproteobacterial or gammaproteobacterial species detected  
221           until Day 20 did not appear again within the niche space (Fig. 3B). These observations

222 suggest that once collapsed, microbial communities may not return to previous states with  
223 highest levels of niche overlap, but refilling of poorly-used niches can occur under the  
224 constraint of limiting similarity within niche space. Although these insights are useful, our  
225 present analysis is based only on 13 time points of a microbiome experiment. Due to the  
226 limitation, it remained difficult to separate effects of  $\alpha$ -diversity from those of gene-repertoire  
227 overlap (Fig. 4B-C). Thus, the statistical analysis proposed in this study need to be expanded  
228 by reducing the cost of metagenomic sequencing as well as by developing more efficient  
229 pipelines for the computationally intensive analyses of metagenomic datasets.

230 The approach of systematically evaluating potential overlap of ecological niches have  
231 been previously explored in “community phylogenetics”, in which phylogenetic  
232 overdispersion/clustering is evaluated based on null model analysis of random assembly from  
233 species pools [5, 31, 32]. In those studies based on phylogenetic analyses, similarity of niches  
234 has been inferred based on the assumption that phylogenetically similar species have similar  
235 ecological properties (e.g., resource requirements). Nonetheless, given that convergent  
236 evolution of ecologically important traits is ubiquitous in the history of life [33–35], the  
237 assumption of phylogenetic niche conservatism is not always met [36]. Therefore, because  
238 gene repertoires are more direct proxies of species traits than phylogeny, metagenome-based  
239 analyses will deepen our understanding of community processes driven by competitive  
240 exclusion. Meanwhile, in the present analyses of gene repertoire overlap, we included whole  
241 metagenomic datasets of the examined microbes due to the lack of *a priori* insights into the  
242 metabolic pathways/processes playing essential roles in interspecific competition for  
243 resources. In this respect, our analysis is a preliminary conceptual step for evaluating potential  
244 overlap of fundamental niches at the community level. In future studies, analyses excluding  
245 housekeeping genes [37, 38] or those focusing on specific functional groups of genes (e.g.,  
246 carbohydrate degrading genes [39]) may provide more reliable inference of niche overlap.  
247 Because such selection of genes can critically influence threshold niche-overlap values for  
248 anticipating abrupt community compositional changes, setting a commonly applicable  
249 criterion of choosing target gene sets will help us perform comparative analyses across a wide  
250 range of microbial communities.

251 The simple framework for evaluating overlap of fundamental niches is applicable to  
252 diverse types of microbiomes. Given that our Jaccard-dissimilarity-based indices are  
253 standardized within the range from 0 to 1, the next crucial step is to examine how threshold  
254 niche overlap values for anticipating microbial community collapse vary among different

255 types of ecosystems. Such threshold values can vary among ecosystems depending on their  
256 basic levels of sustainable functional redundancy. In our laboratory microbiome, for example,  
257 the lack of spatial structure (e.g., refuges for inferior species) and environmental fluctuations  
258 (e.g., temperature fluctuations) might have severely limited coexistence of functionally  
259 similar species (species with similar metabolic capabilities). In contrast, in human gut  
260 microbiomes, spatial complexity [40, 41] and temporally fluctuating environmental  
261 conditions [22] may reduce the risk of competitive exclusion, allowing higher levels of niche  
262 overlap within communities. Thus, extension of time-series metagenomic analyses to diverse  
263 types of ecosystems [42–45] will enhance our knowledge of relationship among ecosystem  
264 properties, functional redundancy, and community stability.

265 While genomic information provides an ultimate platform for inferring fundamental  
266 niches [12–14], overlap of gene repertoires may not always result in competitive exclusion of  
267 species within communities. Even in a pair of species with similar gene repertoires,  
268 differentiation in gene expression patterns may occur to avoid overlap of resource-use  
269 patterns between species, allowing coexistence of the two species in an environment. Such  
270 differentiation of “realized niches [10]“ through phenotypic plasticity is potentially evaluated  
271 by transcriptomic or metabolomic analyses [46, 47]. Consequently, integration of  
272 (meta)transcriptome and (meta)metabolome analyses [48–50] with metagenome-based  
273 analyses will reorganize our understanding of deterministic processes in microbiome  
274 dynamics.

275  
276

277 **MATERIALS AND METHODS**

278 **Time-series data of experimental microbiomes**

279 We used the experimental system of the microbiome time-series monitoring described in a  
280 previous study [19]. In the experiment, microbiomes differing in the magnitude of community  
281 compositional shifts were constructed across the six treatments defined by the combinations  
282 of two inoculum sources and three types of media. One of the source microbiomes derived  
283 from the soil collected from the A layer (0-10 cm in depth) in the research forest of Center for  
284 Ecological Research, Kyoto University, Otsu, Japan (34.972 °N; 135.958 °E). The other  
285 source inoculum was prepared by collecting water from a pond (“Shoubuike”) near Center for  
286 Ecological Research (34.974 °N, 135.966 °E). Each of the source inocula was introduced into  
287 oatmeal (Medium-A), oatmeal-peptone (Medium-B), or peptone (Medium-C) broth media  
288 with eight replicates. Thus, in total, 48 experimental microcosms (two source microbiomes ×  
289 three media × eight replicates) were constructed in a deep-well plate (1000-µL-scale culture in  
290 each well). The plate was kept shaken at 1,000 rpm at 23 °C. After five-day pre-incubation,  
291 200 µL out of the 1,000-µL culture medium was sampled from each well every 24 hours for  
292 110 days. In each sampling event, 200 µL of fresh medium was added to each well so that the  
293 total culture volume was kept constant. In total, 5,280 samples (48 communities/day × 110  
294 days) were collected through the time-series experiment. After DNA extraction, the samples  
295 were subjected to the amplicon sequencing analysis of the 16S rRNA region [19]. To quantify  
296 the speed and magnitude of community shifts through time, the “abruptness” index was  
297 calculated through the time-series of each replicate microcosm in each experimental treatment  
298 [19]. Specifically, an estimate of the abruptness index for time point  $t$  was obtained as the  
299 Bray-Curtis  $\beta$ -diversity between average community compositions from time points  $t - 4$  to  $t$   
300 and those from  $t + 1$  to  $t + 5$  (i.e., dissimilarity between 5-day time-windows). The Bray-

301 Curtis  $\beta$ -diversity [51] was calculated as  $\frac{\sum_{i=1}^n |X_{ij} - X_{ik}|}{\sum_{i=1}^n (X_{ij} + X_{ik})}$ , where  $X_{ij}$  and  $X_{ik}$  denoted relative  
302 abundance of microbial amplicon sequence variant (ASV)  $i$  in the compared time windows ( $j$ ,  
303 from  $t - 4$  to  $t$ ;  $k$ , from  $t + 1$  to  $t + 5$ ). An abruptness score larger than 0.5 indicates that  
304 turnover of more than 50 % of community compositions occurred between the time-windows  
305 [19].

306

307 **Shotgun metagenomics**

308 Focusing on a replicate microcosm in which the most rapid and substantial turnover of  
309 community compositions was observed (replicate no. 5 of Water/Medium-A treatment; Fig.  
310 S1), shotgun metagenomic sequencing was conducted by targeting 13 samples (Day 1, 10, 20,  
311 24, 30, 40, 50, 60, 70, 80, 90, 100, and 110) as described elsewhere [27]. Specifically, each  
312 DNA sample was processed with Nextera XT DNA Library Preparation Kit (Illumina) and  
313 sequenced with the DNBSEQ-G400 (BGI; 200-bp paired-end sequencing). From the output  
314 data, sequencing adaptors were removed using Cutadapt [52] 2.5 and quality filtering was  
315 performed with Fastp [53] 0.21.0: ca. 10 Gb/sample was subjected to the analysis [in total,  
316 159.96 Gb (1000.301 M reads)]. The sequences of each sample were assembled with  
317 metaSPAdes [54] 3.15.2. Binning was then performed with MetaWRAP [55] 1.3.2, followed  
318 by quality assessing with CheckM [56] 1.1.3. The identity between MAGs were calculated  
319 using FastANI [57] 1.33 and MAGs with > 99 % identity were dereplicated through the time-  
320 series (Table S1). In the dereplication, the MAGs with the highest completeness and N50  
321 statistics were selected as representative MAGs. Read-coverage was then calculated with  
322 CoverM [58] 0.6.0, followed by taxonomic annotation was performed using GTDB-Tk [59,  
323 60] 1.6. Only the MAGs with > 80 % completeness and < 5 % contamination were used in the  
324 downstream analyses. Gene annotation was performed with Prokka [28] 1.14.6, yielding  
325 6,999 annotated genes (Data S1). To conduct additional functional annotation of genes, the  
326 orthology numbers of Kyoto Encyclopedia of Genomes (KEGG) were retrieved using  
327 GhostKOALA [61] 2.2. For respective microbial MAGs (bins), completeness of metabolic  
328 pathways was estimated with KEGG decoder [62] 1.3. Based on the matrix representing  
329 KEGG metabolic pathway/process profiles of respective MAGs (Data S2), a heatmap  
330 showing pathway/process completeness was drawn (Fig. S3).

331

### 332 **Background environmental conditions**

333 For the 13 samples subjected to the shotgun metagenomic analysis, concentrations of  
334 ammonium ( $\text{NH}_4^+$ ) and nitrate ( $\text{NO}_3^-$ ) were measured to obtain supplementary information of  
335 background environmental conditions. Colorimetric methods with a modified indophenol  
336 reaction [63, 64] and the VCl3/Griess assay were applied for the measurements of  $\text{NH}_4^+$  and  
337  $\text{NO}_3^-$ , respectively. Samples were run in triplicates via a standard addition method to account  
338 for individual matrix effects [65].

339

340 **Multivariate analysis of the metagenomic space**

341 Based on the whole matrix representing the profiles of the 6,999 genes (Data S1), the Jaccard  
342 metric of distance was calculated for each pair of the 32 microbial MAGs ( $D_{ij}$ , where  $i$  and  $j$   
343 represent MAGs). The Jaccard-distance estimates were then used to perform a principal  
344 coordinate analysis (PCoA). Using the obtained principal coordinate scores, all the microbial  
345 MAGs detected through the time-series were plotted on a multivariate space consisting of the  
346 first three PCoA axes (PCoA 1, PCoA 2, and PCoA 3). At each time point, the MAGs detected  
347 with the shotgun metagenomic sequencing (defined as the MAGs whose relative abundance is  
348 greater than 0.1 %) was plotted on the three-dimensional space defined with the PCoA axes.

349

350 **Evaluation of niche overlap level**

351 The community-scale magnitude of potential niche overlap among species was evaluated  
352 based on the shogun metagenomic sequencing dataset. We developed two types of simple  
353 indices for evaluating community-scale niche overlap as detailed in the Results section. To  
354 test whether a high level of fundamental-niche overlap is observed prior to drastic changes in  
355 microbial community structure, we examined relationship between the above niche overlap  
356 index and time-series shifts in community structure (Bray-Curtis  $\beta$ -diversity between present  
357 and next time points through the time-series of the shotgun metagenomic data).

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359 **REFERENCES**

- 360 1. Gause GF. The Struggle for Coexistence. 1934. Williams & Wilkins, Baltimore.
- 361 2. Hardin G. The competitive exclusion principle. *Science* (1979) 1960; **131**: 1292–1297.
- 362 3. Volterra V. Variations and fluctuations of the number of individuals in animal species  
363 living together. *ICES Journal of Marine Science* 1928; **3**: 3–51.
- 364 4. Zaret TM, Rand AS. Competition in tropical stream fishes: support for the competitive  
365 exclusion principle. *Ecology* 1971; **52**: 336–342.
- 366 5. Mayfield MM, Levine JM. Opposing effects of competitive exclusion on the  
367 phylogenetic structure of communities. *Ecol Lett* 2010; **13**: 1085–1093.

368 6. Grime JP. Competitive exclusion in herbaceous vegetation. *Nature* 1973; **242**: 344–  
369 347.

370 7. Letten AD, Ke PJ, Fukami T. Linking modern coexistence theory and contemporary  
371 niche theory. *Ecol Monogr* 2017; **87**: 161–177.

372 8. Chesson P. Updates on mechanisms of maintenance of species diversity. *Journal of*  
373 *Ecology* 2018; **106**: 1773–1794.

374 9. Chesson P. Mechanisms of maintenance of species diversity. *Annu Rev Ecol Syst* 2000;  
375 **31**: 343–366.

376 10. Chase JM, Leibold MA. Ecological niches: linking classical and contemporary  
377 approaches. *Biodiversity and Conservation* . 2004. University of Chicago Press,  
378 Chicago.

379 11. Hutchinson GE. Concluding Remarks. *Cold Spring Harb Symp Quant Biol* 1957; **22**:  
380 415–427.

381 12. Régimbeau A, Budinich M, Larhlimi A, Pierella Karlusich JJ, Aumont O, Memery L, et  
382 al. Contribution of genome-scale metabolic modelling to niche theory. *Ecol Lett* 2022;  
383 **25**: 1352–1364.

384 13. Smith SR, Dupont CL, McCarthy JK, Broddrick JT, Oborník M, Horák A, et al.  
385 Evolution and regulation of nitrogen flux through compartmentalized metabolic  
386 networks in a marine diatom. *Nat Commun* 2019; **10**: 4552.

387 14. Palomo A, Pedersen AG, Fowler SJ, Dechesne A, Sicheritz-Pontén T, Smets BF.  
388 Comparative genomics sheds light on niche differentiation and the evolutionary history  
389 of comammox *Nitrospira*. *ISME Journal* 2018; **12**: 1779–1793.

390 15. Alneberg J, Bennke C, Beier S, Bunse C, Quince C, Ininbergs K, et al. Ecosystem-wide  
391 metagenomic binning enables prediction of ecological niches from genomes. *Commun  
392 Biol* 2020; **3**: 1–10.

393 16. Fahimipour AK, Gross T. Mapping the bacterial metabolic niche space. *Nat Commun*  
394 2020; **11**: 1–8.

395 17. Carding S, Verbeke K, Vipond DT, Corfe BM, Owen LJ. Dysbiosis of the gut  
396 microbiota in disease. *Microb Ecol Health Dis* 2015; **26**: 26191.

397 18. Ravel J, Brotman RM, Gajer P, Ma B, Nandy M, Fadrosh DW, et al. Daily temporal  
398 dynamics of vaginal microbiota before, during and after episodes of bacterial  
399 vaginosis. *Microbiome* 2013; **1**: 29.

400 19. Fujita H, Ushio M, Suzuki K, Abe M, Yamamichi M, Iwayama K, et al. Alternative  
401 stable states, nonlinear behavior, and predictability of microbiome dynamics.  
402 *Microbiome* 2023; **11**: 63.

403 20. Yajima D, Fujita H, Hayashi I, Shima G, Suzuki K, Toju H. Core species and  
404 interactions prominent in fish-associated microbiome dynamics. *Microbiome* 2023; **11**:  
405 53.

406 21. Lahti L, Salojärvi J, Salonen A, Scheffer M, de Vos WM. Tipping elements in the  
407 human intestinal ecosystem. *Nat Commun* 2014; **5**: 4344.

408 22. David LA, Maurice CF, Carmody RN, Gootenberg DB, Button JE, Wolfe BE, et al.  
409 Diet rapidly and reproducibly alters the human gut microbiome. *Nature* 2014; **505**:  
410 559–563.

411 23. Kho ZY, Lal SK. The human gut microbiome - A potential controller of wellness and  
412 disease. *Front Microbiol* 2018; **9**: 1835.

413 24. Kriss M, Hazleton KZ, Nusbacher NM, Martin CG, Lozupone CA. Low diversity gut  
414 microbiota dysbiosis: drivers, functional implications and recovery. *Curr Opin  
415 Microbiol* 2018; **44**: 34–40.

416 25. Costello EK, Stagaman K, Dethlefsen L, Bohannan BJM, Relman DA. The application  
417 of ecological theory toward an understanding of the human microbiome. *Science  
418 (1979)* 2012; **336**: 1255–1262.

419 26. Huttenhower C, Gevers D, Knight R, Abubucker S, Badger JH, Chinwalla AT, et al.  
420 Structure, function and diversity of the healthy human microbiome. *Nature* 2012; **486**:  
421 207–214.

422 27. Fujita H, Ushio M, Suzuki K, Abe MS, Yamamichi M, Okazaki Y, et al. Facilitative  
423 interaction networks in experimental microbial community dynamics. *Front  
424 Microbiol* . 2023. , **14**: 1153952

425 28. Seemann T. Prokka: Rapid prokaryotic genome annotation. *Bioinformatics* 2014; **30**:

426 2068–2069.

427 29. Anderson MJ, Crist TO, Chase JM, Vellend M, Inouye BD, Freestone AL, et al.  
428 Navigating the multiple meanings of  $\beta$  diversity: A roadmap for the practicing  
429 ecologist. *Ecol Lett* 2011; **14**: 19–28.

430 30. MacArthur R, Levins R. The limiting similarity, convergence, and divergence of  
431 coexisting species. *American Naturalist* 1967; **101**: 377–385.

432 31. Webb CO, Ackerly DD, McPeek MA, Donoghue MJ. Phylogenies and community  
433 ecology. *Annu Rev Ecol Syst* 2002; **33**: 475–505.

434 32. Cavender-Bares J, Kozak KH, Fine PVA, Kembel SW. The merging of community  
435 ecology and phylogenetic biology. *Ecol Lett* 2009; **12**: 693–715.

436 33. Marvig RL, Sommer LM, Molin S, Johansen HK. Convergent evolution and adaptation  
437 of *Pseudomonas aeruginosa* within patients with cystic fibrosis. *Nat Genet* 2015; **47**:  
438 57–64.

439 34. McCutcheon JP, McDonald BR, Moran NA. Convergent evolution of metabolic roles  
440 in bacterial co-symbionts of insects. *Proc Natl Acad Sci U S A* 2009; **106**: 15394–  
441 15399.

442 35. Merhej V, Royer-Carenzi M, Pontarotti P, Raoult D. Massive comparative genomic  
443 analysis reveals convergent evolution of specialized bacteria. *Biol Direct* 2009; **4**: 13.

444 36. Losos JB. Phylogenetic niche conservatism, phylogenetic signal and the relationship  
445 between phylogenetic relatedness and ecological similarity among species. *Ecol Lett* .  
446 2008. , **11**: 995–1007

447 37. Gibson DG, Glass JI, Lartigue C, Noskov VN, Chuang R-Y, Algire M a, et al. Creation  
448 of a bacterial cell controlled by a chemically synthesized genome. *Science* 2010; **329**:  
449 52–6.

450 38. Maiden MCJ, van Rensburg MJJ, Bray JE, Earle SG, Ford SA, Jolley KA, et al. MLST  
451 revisited: The gene-by-gene approach to bacterial genomics. *Nat Rev Microbiol* 2013;  
452 **11**: 728–736.

453 39. Flint HJ, Scott KP, Duncan SH, Louis P, Forano E. Microbial degradation of complex  
454 carbohydrates in the gut. *Gut Microbes* 2012; **3**: 289–306.

455 40. Tropini C, Earle KA, Huang KC, Sonnenburg JL. The Gut Microbiome: Connecting  
456 Spatial Organization to Function. *Cell Host Microbe* 2017; **21**: 433–442.

457 41. Earle KA, Billings G, Sigal M, Lichtman JS, Hansson GC, Elias JE, et al. Quantitative  
458 Imaging of Gut Microbiota Spatial Organization. *Cell Host Microbe* 2015; **18**: 478–  
459 488.

460 42. Jansson JK, Hofmockel KS. Soil microbiomes and climate change. *Nat Rev Microbiol* .  
461 2020. , **18**: 35–46

462 43. Fierer N. Embracing the unknown: Disentangling the complexities of the soil  
463 microbiome. *Nat Rev Microbiol* 2017; **15**: 579–590.

464 44. Trivedi P, Leach JE, Tringe SG, Sa T, Singh BK. Plant–microbiome interactions: from  
465 community assembly to plant health. *Nat Rev Microbiol* 2020; **18**: 607–621.

466 45. Venter JC, Remington K, Heidelberg JF, Halpern AL, Rusch D, Eisen JA, et al.  
467 Environmental Genome Shotgun Sequencing of the Sargasso Sea. *Science* (1979)  
468 2004; **304**: 66–74.

469 46. Nowinski B, Moran MA. Niche dimensions of a marine bacterium are identified using  
470 invasion studies in coastal seawater. *Nat Microbiol* 2021; **6**: 524–532.

471 47. Pereira FC, Berry D. Microbial nutrient niches in the gut. *Environ Microbiol* 2017; **19**:  
472 1366–1378.

473 48. Heintz-Buschart A, Wilmes P. Human gut microbiome: function matters. *Trends  
474 Microbiol* 2018; **26**: 563–574.

475 49. Schirmer M, Franzosa EA, Lloyd-Price J, McIver LJ, Schwager R, Poon TW, et al.  
476 Dynamics of metatranscription in the inflammatory bowel disease gut microbiome. *Nat  
477 Microbiol* 2018; **3**: 337–346.

478 50. Turner TR, Ramakrishnan K, Walshaw J, Heavens D, Alston M, Swarbreck D, et al.  
479 Comparative metatranscriptomics reveals kingdom level changes in the rhizosphere  
480 microbiome of plants. *ISME Journal* 2013; **7**: 2248–2258.

481 51. Legendre P, de Cáceres M. Beta diversity as the variance of community data:  
482 Dissimilarity coefficients and partitioning. *Ecol Lett* 2013; **16**: 951–963.

483 52. Martin M. Cutadapt removes adapter sequences from high-throughput sequencing  
484 reads. *EMBnet J* 2011; **17**: <https://doi.org/10.14806/ej.17.1.200>.

485 53. Chen S, Zhou Y, Chen Y, Gu J. Fastp: An ultra-fast all-in-one FASTQ preprocessor.  
486 *Bioinformatics* 2018; **34**: i884–i890.

487 54. Bankevich A, Nurk S, Antipov D, Gurevich AA, Dvorkin M, Kulikov AS, et al.  
488 SPAdes: A new genome assembly algorithm and its applications to single-cell  
489 sequencing. *Journal of Computational Biology* 2012; **19**: 455–477.

490 55. Uritskiy G v., DiRuggiero J, Taylor J. MetaWRAP—a flexible pipeline for genome-  
491 resolved metagenomic data analysis. *Microbiome* 2018; **6**: 158.

492 56. Parks DH, Imelfort M, Skennerton CT, Hugenholtz P, Tyson GW. CheckM: Assessing  
493 the quality of microbial genomes recovered from isolates, single cells, and  
494 metagenomes. *Genome Res* 2015; **25**: 1043–1055.

495 57. Jain C, Rodriguez-R LM, Phillippe AM, Konstantinidis KT, Aluru S. High throughput  
496 ANI analysis of 90K prokaryotic genomes reveals clear species boundaries. *Nat  
497 Commun* 2018; **9**: 5114.

498 58. Woodcroft B. CoverM: program available at <https://github.com/wwood/CoverM>. 2021.

499 59. Chaumeil PA, Mussig AJ, Hugenholtz P, Parks DH. GTDB-Tk: A toolkit to classify  
500 genomes with the genome taxonomy database. *Bioinformatics* 2020; **36**: 1925–1927.

501 60. Parks DH, Chuvochina M, Rinke C, Mussig AJ, Chaumeil P-A, Hugenholtz P. GTDB:  
502 an ongoing census of bacterial and archaeal diversity through a phylogenetically  
503 consistent, rank normalized and complete genome-based taxonomy. *Nucleic Acids Res*  
504 2022; **50**: D785–D794.

505 61. Kanehisa M, Sato Y, Morishima K. BlastKOALA and GhostKOALA: KEGG Tools for  
506 Functional Characterization of Genome and Metagenome Sequences. *J Mol Biol* 2016;  
507 **428**: 726–731.

508 62. Graham ED, Heidelberg JF, Tully BJ. Potential for primary productivity in a globally-  
509 distributed bacterial phototroph. *ISME Journal* 2018; **12**: 1861–1866.

510 63. Kandeler E, Gerber H. Short-term assay of soil urease activity using colorimetric  
511 determination of ammonium. *Biol Fertil Soils* 1988; **6**: 68–72.

512 64. Hood-Nowotny R, Umana NH-N, Inselbacher E, Oswald- Lachouani P, Wanek W.  
513 Alternative methods for measuring inorganic, organic, and total dissolved nitrogen in  
514 soil. *Soil Science Society of America Journal* 2010; **74**: 1018–1027.

515 65. Taylor BW, Keep CF, Hall RO, Koch BJ, Tronstad LM, Flecker AS, et al. Improving  
516 the fluorometric ammonium method: Matrix effects, background fluorescence, and  
517 standard additions. *J North Am Benthol Soc* 2007; **26**: 167–177.

518

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526

527 **AUTHOR CONTRIBUTIONS**

528 H.T. designed the work with H.F. H.F. and A.C. performed experiments. H.F. analyzed the  
529 data with Y.O., and H.T. H.F. and H.T. wrote the paper with all the authors.

530

531 **COMPETING INTERESTS**

532 The authors declare no competing interests.

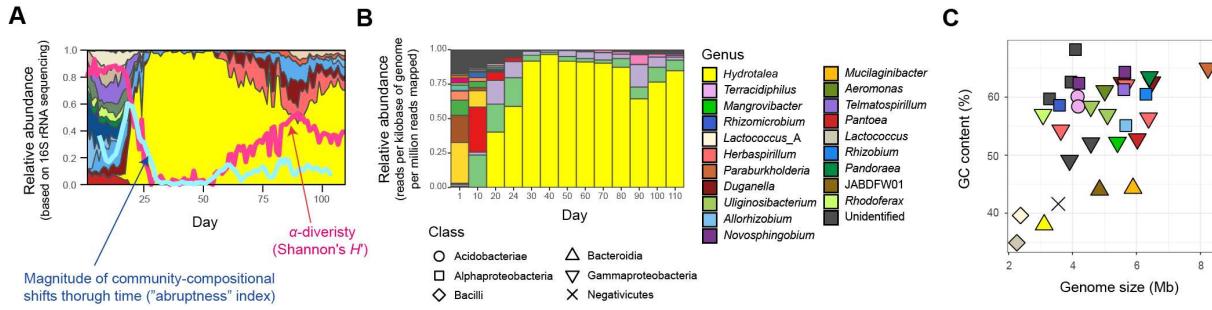
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534 **DATA AVAILABILITY STATEMENT**

535 The 16S rRNA sequencing data reported in a previous study [19] are available from the DNA  
536 Data Bank of Japan (DDBJ) with the accession number DRA013352, DRA013353,  
537 DRA013354, DRA013355, DRA013356, DRA013368 and DRA013379. The shotgun  
538 metagenomic data reported previous [27] are available with the DDBJ accession number  
539 DRA013382. The microbial community data are deposited at our GitHub repository  
540 ([https://github.com/hiroakif93/MTS\\_nicheSpace](https://github.com/hiroakif93/MTS_nicheSpace)) [to be publicly available after acceptance of  
541 the paper]. The matrices of the shotgun metagenomic data are available as Data S1 and 2. All  
542 the scripts used to analyze the data are available at the GitHub repository  
543 ([https://github.com/hiroakif93/MTS\\_nicheSpace](https://github.com/hiroakif93/MTS_nicheSpace)) [to be publicly available after acceptance of  
544 the paper].

545

546



547 **Fig. 1 Community and ecosystem dynamics. A** Time-series data of community structure.  
548 For the replicate microcosm that showed the most abrupt community compositional changes  
549 through the 110-day microbiome experiment [19] (Fig. S1), family-level taxonomic  
550 compositions inferred with 16S rRNA sequencing are shown. The blue line represents the  
551 speed and magnitude of community compositional changes around each time point  
552 (“abruptness” index [19]; see Materials and Methods). The red line indicates  $\alpha$ -diversity  
553 (Shannon’s  $H'$ ) of microbial ASVs [19]. Note that a value larger than 0.5 represents turnover  
554 of more than 50 % of microbial ASV compositions. See Fig. S1 for color profiles of bacterial  
555 families. Reproduced from the data of a previous study [19]. **B** Taxonomic compositions  
556 inferred with shotgun metagenomic sequencing. At each of the 13 time points through the  
557 time-series of the target microcosm, the relative abundance of each MAG was estimated  
558 based on the normalized read coverage value (reads per kilobase of genome per million reads  
559 mapped). **C** Genome size and GC nucleotide content of the MAGs detected in the target  
560 microcosm. See panel **A** for colors and symbols.

561



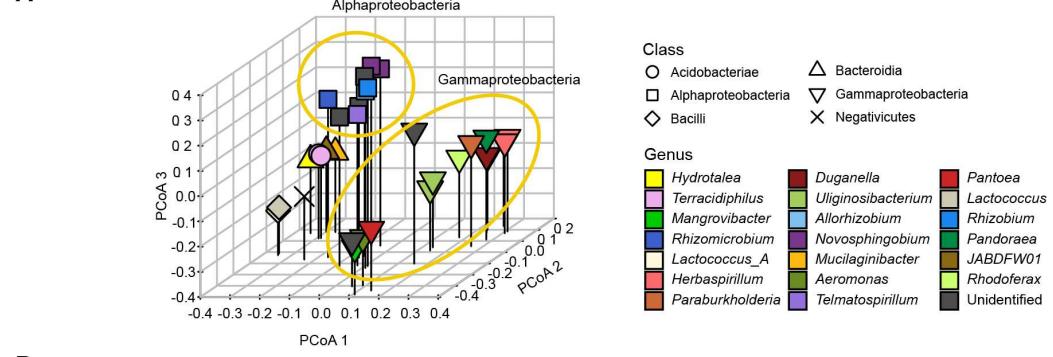
562

563 **Fig. 2 Metabolic pathway/process profiles of the MAGs.** KEGG metabolic  
 564 pathways/profiles of the reconstructed bacterial genomes (MAGs) are shown. The detection  
 565 (relative abundance > 0.1 %) of each microbial MAG on each day within the shotgun  
 566 metagenomic data is indicated in the panel below. Only the microbial MAGs with > 80 %  
 567 completeness and < 5 % contamination were included (Table S1). The five MAGs that co-  
 568 occurred from Day 40 to 60 and metabolic pathways/processes mentioned in the main text are  
 569 highlighted. Only the metabolic pathways/processes with highly heterogeneous patterns  
 570 across microbial MAGs are shown. See Fig. S3 for detailed profiles of the metabolic

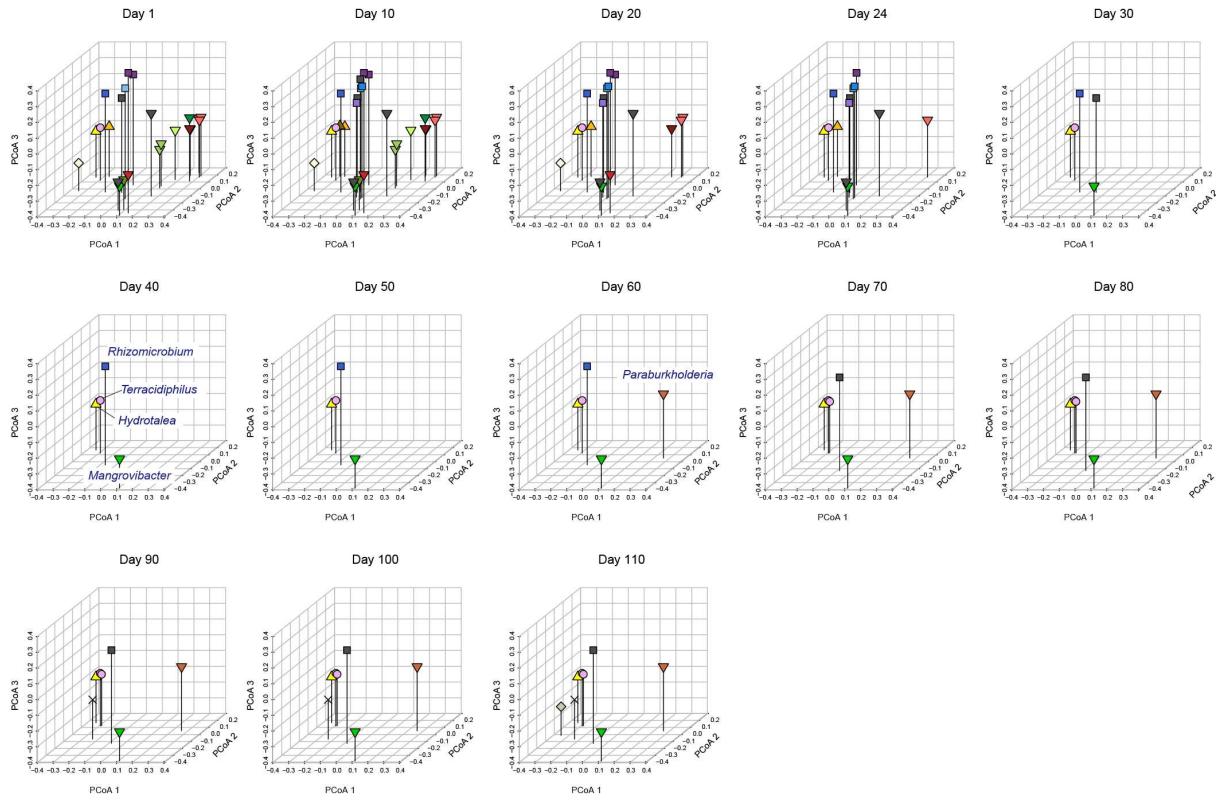
571 pathways/processes.

572

A

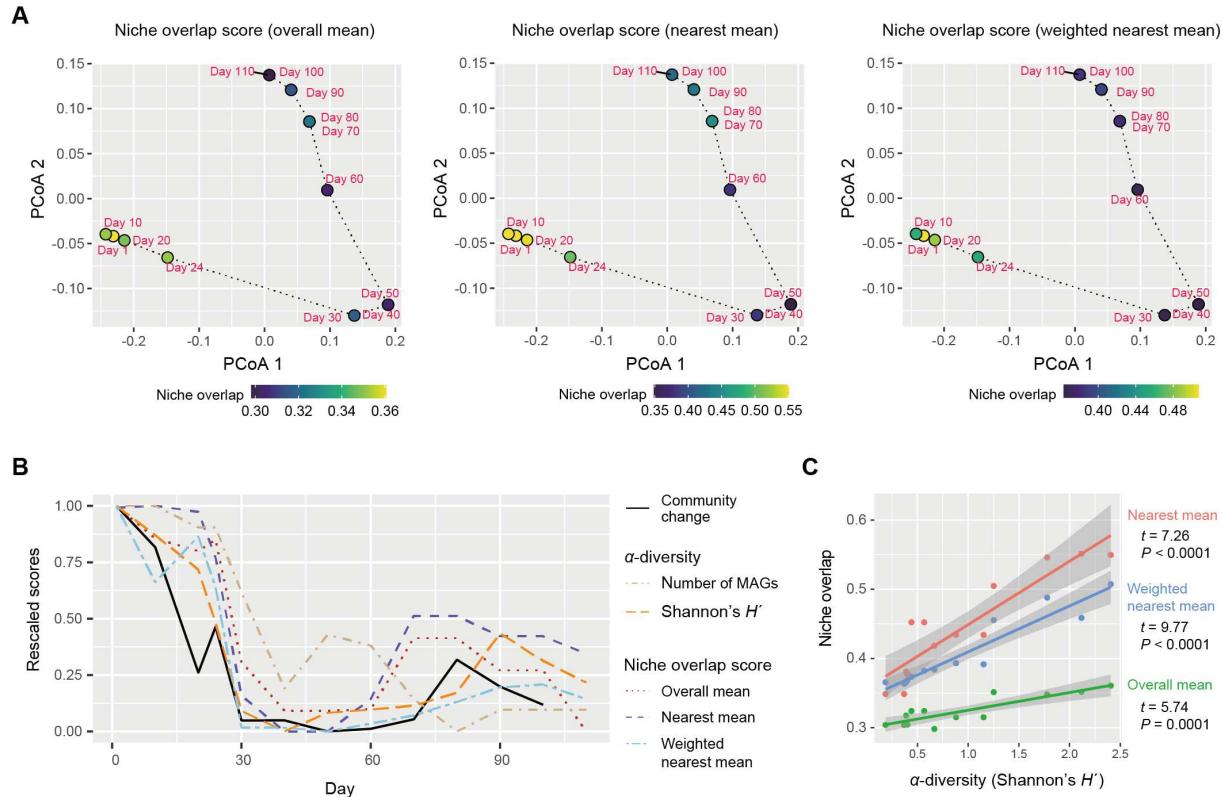


B



573

574 **Fig. 3 Metagenomic niche space. A** Distributions of MAGs within metagenomic niche  
 575 space. Based on dissimilarity in gene repertoires, microbial MAGs that appeared in the time-  
 576 series of the target microcosm were plotted on the three-dimensional space defined by the  
 577 principal coordinate analysis (PCoA) of 6,999 genes. **B** Changes in the distributions of  
 578 microbial MAGs within niche space. At each time point, detected MAGs (relative abundance  
 579 > 0.1 %) were plotted on the space defined in the multivariate analysis in the in the panel A.  
 580



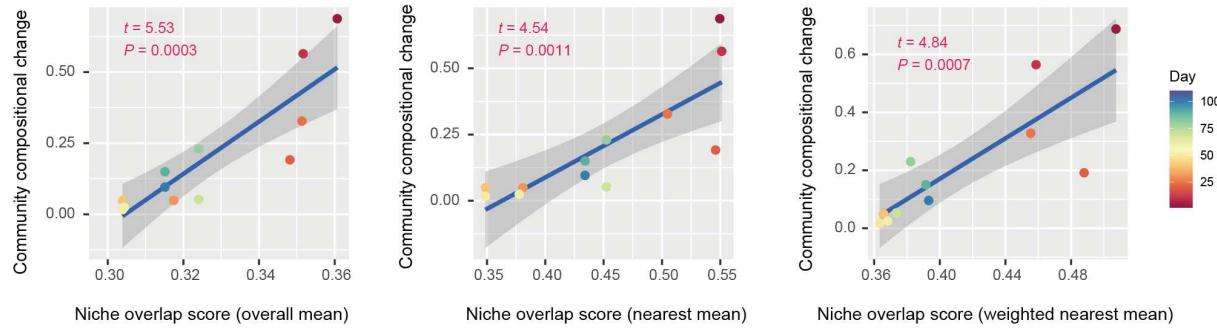
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582

583 **Fig. 4 Dynamics of niche-overlap level.** **A** Community-level profiles of metabolic  
 584 pathways/processes and niche overlap index. The niche overlap indices were defined based on  
 585 the Jaccard similarity/dissimilarity of gene compositions between pairs of the microbial  
 586 MAGs detected at a target time point. Three types of niche overlap indices are shown on a  
 587 PCoA surface representing community-level compositions of genes. On the PCoA surface,  
 588 time points are distributed based on the sum of the gene repertoires of the detected MAGs. **B**  
 589 Dynamics of niche-overlap levels. Niche overlap scores are shown across the time-series. The  
 590 magnitude of community compositional changes (Bray-Curtis  $\beta$ -diversity between present  
 591 and next time points through the time-series of the shotgun metagenomic data) and  $\alpha$ -  
 592 diversity indices of the communities are shown as well. **C** Relationship between  $\alpha$ -diversity  
 593 and niche overlap scores. The lines represent linear regressions (with 95 % confidence  
 594 intervals).

595

596



597

Niche overlap score (overall mean)

Niche overlap score (nearest mean)

Niche overlap score (weighted nearest mean)

598 **Fig. 5 Niche overlap level and community compositional shifts.** The magnitude of  
599 community compositional changes observed in the microbiome was regressed on each niche  
600 overlap index obtained based on the shotgun metagenomic analysis. Niche overlap index at  
601 each time point and time-series shifts in community structure (Bray-Curtis  $\beta$ -diversity  
602 between present and next time points through the time-series of the shotgun metagenomic  
603 data) are shown along horizontal and vertical axes, respectively. The regression lines are  
604 shown with 95 % confidence intervals.