

1 *Article - discoveries*

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3 *Seascape genomics reveals candidate*
4 *molecular targets of heat stress adaptation in*
5 *three coral species*

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27

28 **Abstract**

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30 Anomalous heat waves are causing a major decline of hard corals around the world and
31 threatening the persistence of coral reefs. There are, however, reefs that had been exposed
32 to recurrent thermal stress over the years and whose corals appeared tolerant against heat.
33 One of the mechanisms that could explain this phenomenon is local adaptation, but the
34 underlying molecular mechanisms are poorly known.

35 In this work, we applied a seascape genomics approach to study heat stress adaptation in
36 three coral species of New Caledonia (southwestern Pacific) and to uncover molecular actors
37 potentially involved. We used remote sensing data to characterize the environmental trends
38 across the reef system, and sampled corals living at the most contrasted sites. These samples
39 underwent next generation sequencing to reveal single-nucleotide-polymorphisms (SNPs) of
40 which frequencies associated with heat stress gradients. As these SNPs might underpin an
41 adaptive role, we characterized the functional roles of the genes located in their genomic
42 neighborhood.

43 In each of the studied species, we found heat stress associated SNPs notably located in
44 proximity of genes coding for well-established actors of the cellular responses against heat.
45 Among these, we can mention proteins involved in DNA damage-repair, protein folding,
46 oxidative stress homeostasis, inflammatory and apoptotic pathways. In some cases, the same
47 putative molecular targets of heat stress adaptation recurred among species.

48 Together, these results underscore the relevance and the power of the seascape genomics
49 approach for the discovery of adaptive traits that could allow corals to persist across wider
50 thermal ranges.

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53 **Introduction**

54

55 One of the most dramatic consequences of climate change is the worldwide decline of coral
56 reefs, which are the most biodiverse ecosystems in the marine environment (Hughes et al.
57 2017). Among the main drivers of this decline is coral bleaching, a stress response to
58 anomalous heat waves that eventually causes the death of hard corals (Bellwood et al. 2004;
59 Hughes et al. 2017). In the most severe episodes, coral bleaching provoked local living coral
60 cover loss of up to 50% (Hughes et al. 2017; Hughes et al. 2018), with climate change
61 projections expecting for bleaching conditions to be persistent worldwide by 2050 (Van
62 Hooidonk et al. 2013).

63 Despite these catastrophic perspectives, a glimpse of hope is brought by coral reefs that show
64 resistance after recurrent heat waves (Thompson and van Woesik 2009; Penin et al. 2013;
65 Krueger et al. 2017; Dance 2019; Hughes et al. 2019). One of the mechanisms that might
66 promote heat tolerance in corals is genetic adaptation (Sully et al. 2019). In recent years,
67 there has been a growing body of literature investigating how coral thermal adaptation might
68 alter the predictions of reef persistence, and how conservation policies could be modified
69 accordingly (Logan et al. 2014; Van Oppen et al. 2015; Matz et al. 2018).

70 Given the crucial role adaptation will play in long-term reef persistence, there is an urgent
71 need to characterize the adaptive potential of corals (Logan et al. 2014; Van Oppen et al.
72 2015). For instance, there are still open questions concerning the spatial and temporal scales
73 at which local adaptation operates (Matz et al. 2018; Roche et al. 2018). Changes in adaptive
74 potential have been observed along thermal gradients over hundreds of kilometres (e.g.
75 (Thomas et al. 2017), but also at reefs with distinct thermal variations located only a few
76 hundreds of meters apart (e.g. Bay and Palumbi 2014). Furthermore, different coral species
77 are reported to show differential vulnerability against thermal stress, leading to the question
78 of how different life-history traits (e.g. reproductive strategies) drive the pace of adaptation
79 (Loya et al. 2001; Darling et al. 2012; Hughes et al. 2018).

80 There are also open questions concerning the molecular mechanisms that might be targeted
81 by heat stress adaptation in corals (van Oppen and Lough 2009; Mydlarz et al. 2010; Palumbi
82 et al. 2014). Some cellular responses to heat stress are now well characterized, such as DNA
83 repair mechanisms, the activation of the protein folding machinery in the endoplasmic
84 reticulum (ER) or the accumulation of reactive oxygen species (ROS, either endogenous or

85 produced by the symbiont) that progressively elicits inflammatory and apoptotic responses
86 (van Oppen and Lough 2009; Mydlarz et al. 2010; Maor-Landaw and Levy 2016; Oakley et al.
87 2017; Patel et al. 2018). However, little is known about which of the many molecular actors
88 participating to these cascades could be hijacked by evolutionary processes to increment
89 thermal tolerance.

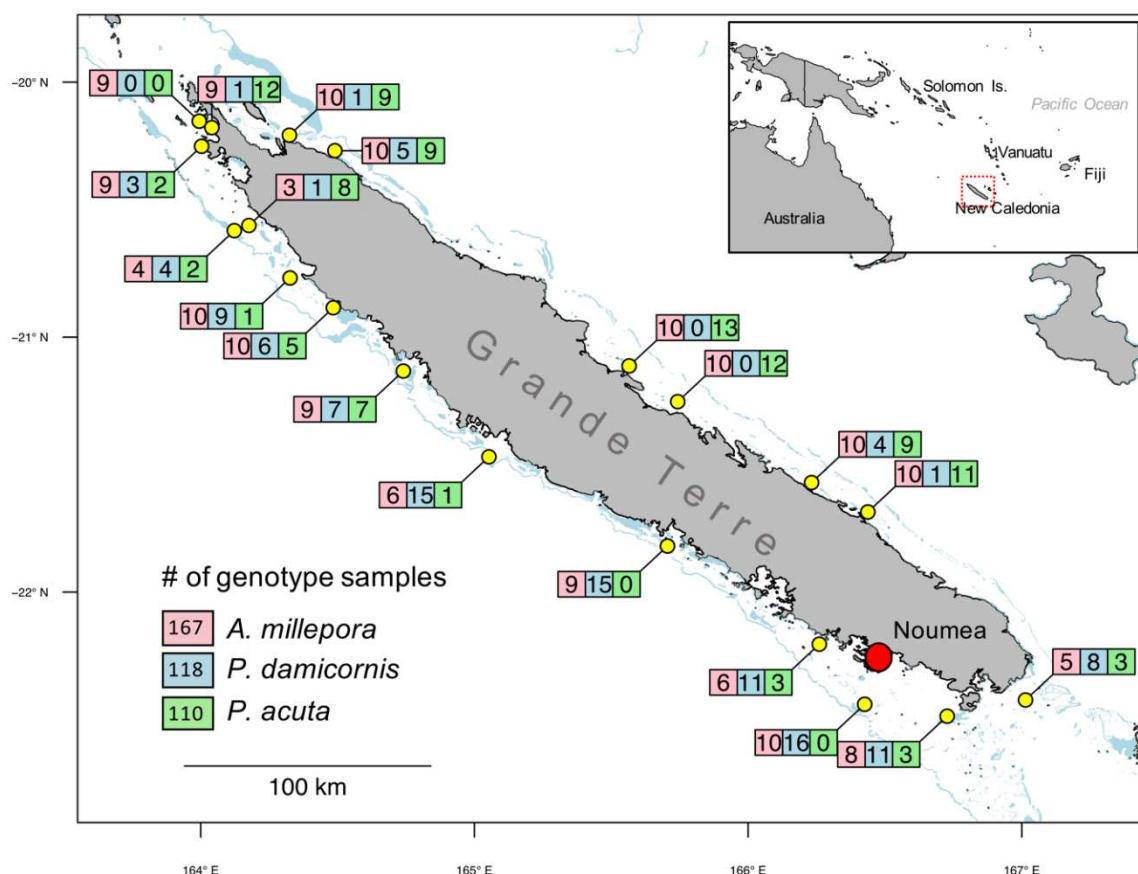
90 Seascapes genomics could contribute to filling these gaps. Seascapes genomics is a budding
91 field of population genomics that allows the study of local adaptation in wild populations
92 (Riginos et al. 2016). This method combines the environmental characterization of the
93 seascapes with a genomic analysis of its population (Rellstab et al. 2015). The goal is to identify
94 genetic variants that correlate with environmental gradients that might underpin an adaptive
95 role (Rellstab et al. 2015). Seascapes genomics could enhance the characterization of coral
96 adaptive potential because: (1) it requires an extensive sampling strategy that allows for
97 studying adaptation at different geographic scales, and against different types of
98 environmental constraints simultaneously (e.g. mean temperatures, standard deviations,
99 accumulated heat stress; Leempoel et al. 2017; Selmoni et al. 2020a); (2) its experimental
100 protocol is less laborious in comparison to traditional approaches used for studying coral
101 adaptation (e.g. aquarium experiments, transplantations), and therefore facilitates scaling-
102 up to a multiple species analysis; (3) it is based on genomic data and thus reports candidate
103 molecular targets of adaptation (Rellstab et al. 2015; Riginos et al. 2016). Moreover, recent
104 work described how the results of seascapes genomics studies on corals can be directly
105 transposed to a conservation perspective and support reef prioritization (Selmoni et al.
106 2020b).

107 Here we applied the seascapes genomics approach to study the adaptive potential against heat
108 stress in three bleaching-prone coral species of New Caledonia, in the southwestern Pacific
109 (Fig. 1). We first used publicly available satellite data to characterize the seascapes conditions
110 for over 1,000 km of the reef system. A sampling campaign was then organized to collect
111 colonies at the 20 sites exposed to the most contrasted environmental conditions. The
112 collected samples underwent a genotype-by-sequencing (DART-seq) genomic
113 characterization, followed by a seascapes genomics analysis accounting for the confounding
114 role of demographic structure. This allowed us to uncover single nucleotide polymorphisms
115 (SNPs) associated with heat stress. We then performed the functional annotations of genes
116 surrounding these SNPs and found molecular targets that notably recurred among species

117 and that referred to well established heat stress responses in coral cells. Our study lays the
118 foundations for the discovery of adaptive traits that could allow corals to persist across wider
119 thermal ranges.

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Figure 1. Study area and sampling sites. The 20 sampling sites around Grande Terre, the main island of New Caledonia (South Western Pacific), are shown in yellow. For every sampling site, the number of genotyped individuals per species (*Acropora millepora*: red, *Pocillopora damicornis*: blue, *Pocillopora acuta*: green) are given in the corresponding boxes.



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123 Results

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125 Three coral species were sampled at 20 sites across the reef system of New Caledonia:
126 *Acropora millepora* (Ehrenberg, 1834; n=360), *Pocillopora damicornis* (Linnaeus, 1758;
127 (n=128) and *Pocillopora acuta* (Lamarck, 1816 ; n=150; Tab. S2). The DArT-seq analytical

128 pipeline resulted in the genotyping of 188 samples by 57,374 bi-allelic single nucleotide
129 polymorphisms (SNPs) for *A. millepora*, and 128 and 150 samples by 70,640 SNPs for
130 *P. damicornis* and *P. acuta*, respectively (Tab. 1). After filtering for rare variants, missing
131 values and clonality, we obtained a final genotype matrix of 167 individuals by 11,935 SNPs
132 for *A. millepora*, of 118 individuals by 7,895 SNPs for *P. damicornis* and of 110 individuals by
133 8,343 SNPs for *P. acuta* (Tab. 1). The *A. millepora* genotyped samples distributed across all
134 the 20 sampling sites (18 of which counted five samples or more), while genotyped samples
135 of *P. damicornis* and *P. acuta* were distributed across 17 sites each (both with 10 sites
136 counting five samples or more), with 15 sites where both species were found in sympatry
137 (Fig. 1; Tab. S1).

Table 1. Workflow of the analysis. For each of the species of interest (*Acropora millepora*, *Pocillopora damicornis*, *Pocillopora acuta*), we report the number of individuals (ind.) and single nucleotide polymorphism (SNPs) obtained or retained after each the various step of the workflow.

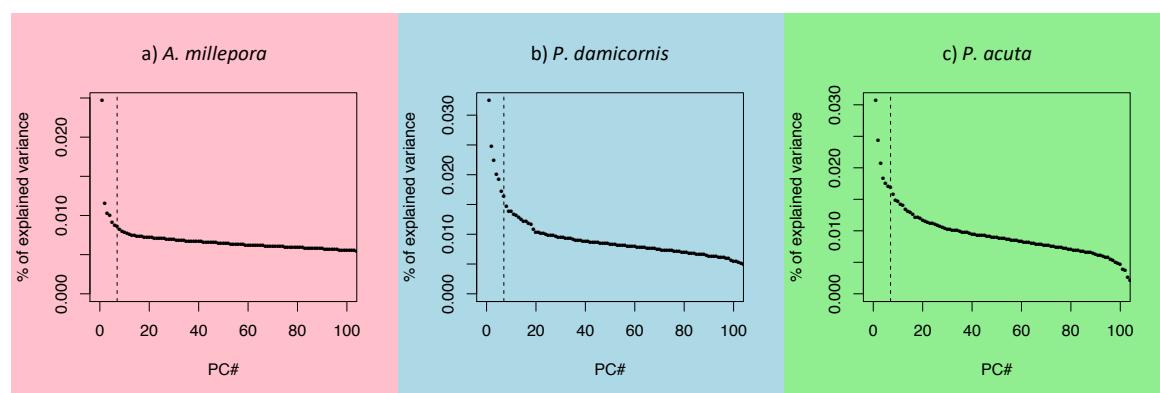
	<i>A. millepora</i>	<i>Pocillopora</i>	
		<i>P. damicornis</i>	<i>P. acuta</i>
Sampling	370 ind.	360 ind.	
Microsatellite	-	128 ind.	150 ind.
DArT-seq	188 ind. x 57,374 SNPs	128 ind. x 70,640 SNPs	150 ind. x 70,640 SNPs
BLAST against reference	188 ind. x 47,529 SNPs	127 ind. x 48,049 SNPs	145 ind. x 48,049 SNPs
Filtering (Missing values, MAF, MGF, LD, clonality)	167 ind. x 11,935 SNPs	118 ind. x 7,895 SNPs	110 ind. x 8,343 SNPs

138
139 *Neutral genetic structure*
140
141 We ran a principal component analysis (PCA) of the genotype matrix of each species to
142 anticipate possible confounding role of neutral genomic variation on the adaptation study
143 (Fig. 2, Fig. S1-S2). The Tracy-Widom test ($P < 0.05$) revealed that the number of PCs underlying
144 a non-random structure were seven for all the species, accounting for 8% of the total variance
145 in *A. millepora*, 15% in both *P. damicornis* and *P. acuta* (Fig. 2). In *P. damicornis*, the spatial
146 distribution of PC1 values appeared to be spatially structured following a north-south

147 separation along the west coast of Grande Terre (Fig. S1B). In *P. acuta*, colonies in the north-
148 west of Grande Terre displayed lower PC1 values, compared to those on the eastern coast
149 (Fig. S1C). In *A. millepora*, no clear geographical patterns emerged as individuals with
150 different values on PC1 were often located on the same reef. Finally, we analysed the
151 presence of genomic widows clustering SNPs with high PC1-loadings, expected to be frequent
152 in genetically isolated groups (cryptic species, hybrids). These genomics windows were rare,
153 as we observed no more than three per species (Fig. S2).

154

Figure 2. Principal component analysis (PCA) of the genotype matrices for three species studied, *Acropora millepora*, *Pocillopora damicornis* and *Pocillopora acuta*. The three graphs display the percentage of variance explained by the first 100 principal components (PC) of the genotype matrix for the three studied species. The vertical dotted lines represent the number of PCs deemed as underlying a non-random structure by the Tracy-Widom test ($P < 0.05$).



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156 *Local adaptation*

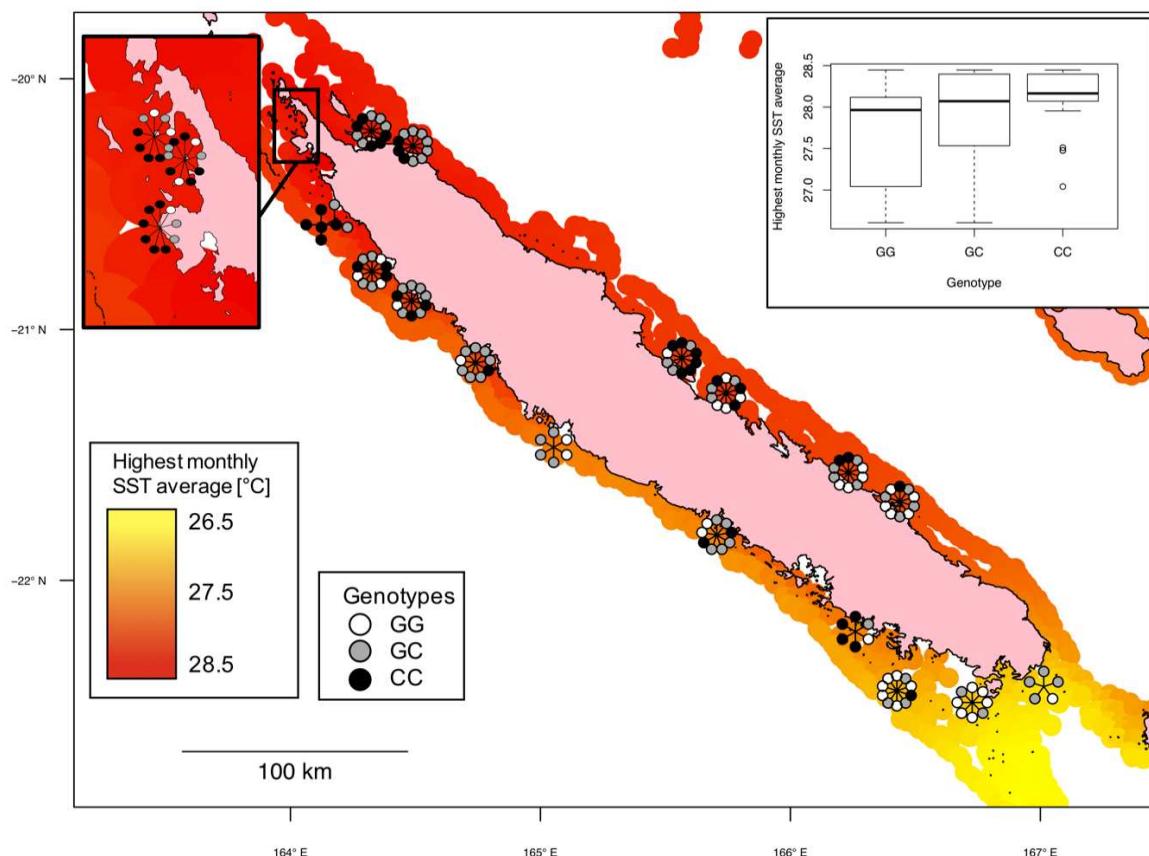
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158 We investigated the presence of SNPs that were associated with 47 environmental gradients
159 describing the seascape conditions in New Caledonia (Fig. 3, Fig. S3, Tab. S2). When a SNP was
160 found associated with multiple environmental descriptors, only the most significant
161 association was kept. In total, 120 significant ($q < 0.01$) genotype-environment associations
162 were found for *A. millepora*, 90 for *P. damicornis* and 100 for *P. acuta* (Tab. 2a; Tab. S3, S4).
163 In all of the three species, we investigate the environmental descriptors that most frequently
164 associated with significant SNPs were those related to sea surface temperature (SST; 63 in
165 *A. millepora*, 47 in *P. damicornis*, 43 in *P. acuta*). Among these, we found that putative
166 adaptive signals related to bleaching alert frequencies (73 genotype-environment
167 associations) were more frequent than those relating to the standard deviation (42) and
168 average temperatures (38; Tab. 2b).

169 When we focused on the environmental descriptors not relating to temperature, we
170 observed that those describing chlorophyll concentration were those associated to more
171 SNPs (64 across the three species), followed by salinity (40; Tab. 2a). In contrast, current
172 velocity variables were the best environmental descriptors associated with fewer genetic
173 variants (10 across the three species, Tab. 2a).

174

Figure 3. Example of significant genotype-environment association. The map displays the superposition between environmental gradient (here highest monthly SST average) and the distribution of an associated ($q < 0.01$) SNP of *Acropora millepora*. Every circle corresponds to the SNP genotype for an individual colony. For illustrative reasons, genotypes are radially distributed around the sampling locations. The boxplot in the top-right corner shows how the environmental variable distributes within each genotype. The SNP represented here is located on the contig xpSc0000535 (position 118526) of *A. millepora* genome, and the closest annotated gene codes for ATP-dependent DNA helicase Q5.



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176

Table 2. Significant genotype-environment associations in *Acropora millepora*, *Pocillopora damicornis* and *Pocillopora acuta* Table a displays the number of SNPs significantly associated ($q < 0.01$) with environmental descriptors; Table b, the complete list of environmental descriptors related to sea surface temperature (averages and standard deviations at two different spatial resolutions and indices of bleaching alert frequencies - BAF). Note that when a SNP was significantly associated to multiple environmental descriptors, only the best association was kept. The detailed list of the SNP-environment associations is available in the supplementary Table 4.

a) All environmental descriptors

Environmental descriptor	<i>A. millepora</i>	<i>P. damicornis</i>	<i>P. acuta</i>
Sea surface temperature	63	47	43
Alkalinity	11	7	14
Chlorophyll concentration	15	22	27
Sea current velocity	4	4	2
Suspended particulate matter	8	1	2
Salinity	19	9	12
Total	120	90	100

b) Sea surface temperature only

Environmental descriptor		<i>A. millepora</i>	<i>P. damicornis</i>	<i>P. acuta</i>	
Sea surface temperature	5 km	Overall average	1	0	
		Average warmest month	1	1	
		Average coldest month	1	1	
		Overall standard deviation	1	2	
		Standard deviation hottest month	0	0	
	1 km	Standard deviation coldest month	11	1	
		BAF _{0°C}	2	1	
		BAF _{4°C}	3	9	
		BAF _{8°C}	1	1	
		BAF _{CRW}	2	3	
Overall average		0	1	1	
Average warmest month		5	2	4	
Average coldest month		12	4	3	
Overall standard deviation		4	3	2	
Standard deviation hottest month		2	1	0	
Standard deviation coldest month		2	4	1	
BAF _{0°C}		10	8	8	
BAF _{4°C}		3	3	3	
BAF _{8°C}		2	1	1	
BAF _{CRW}		0	1	4	
Total		63	47	43	

177

178 *Functional annotations of heat stress associated SNPs*

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180 Genes around SNPs that were associated with heat stress were annotated with gene ontology
 181 (GO) terms to investigate the molecular functions potentially altered by a genetic variant. GO
 182 terms were ranked for over-representation according to the Fisher exact test *P*-value (Tab. 3,
 183 S5). Among the top 50 ranks we found GO terms describing molecular functions such as
 184 “mismatched DNA binding”, “heat shock protein binding”, “chaperone binding”, “unfolded
 185 protein binding”, “cytoskeletal protein binding” and “actin binding” in *A. millepora* (Tab. 3a,

186 S5a); “actin filament binding receptor activity”, “exonuclease activity”, “endonuclease
187 activity” and “death effector domain binding” in *P. damicornis* (Tab. 3b, S5b); “NAD binding”,
188 “nucleotide binding” and “mitogen-activated protein kinase binding” in *P. acuta* (Tab. 3c,
189 S5c). Four terms (“signaling receptor binding”, “receptor regulator activity”, “organic cation
190 transmembrane transporter activity”, “enzyme activator activity”) recurred among top
191 ranked GO terms in at least two different species (Tab. 3). Each of the three species displayed
192 top ranked GO terms referring to “oxidoreductase activity” acting on different molecules
193 (Tab. 3).

194

Table 3. Functional annotations of heat stress associated SNPs. For each of the studied species, *Acropora millepora*, *Pocillopora damicornis* and *Pocillopora acuta*, the tables display the list of GO terms describing molecular functions that are overrepresented in the genomic neighborhoods (± 50 kb) of heat stress associated SNPs. For each GO term, the tables display the id (GO.ID), the term description (Term), the occurrence in the genome split in 100 kb windows (#Ann.), the observed (#Obs.) and expected (#Exp.) occurrence in the neighborhoods of heat stress associated SNPs and the *P*-value associated to the Fisher exact test comparing the expected and observed occurrences. For every species, the tables show a subset of the top 50 GO terms. The complete list of top ranked GO terms is available in the supplementary material (Tab. S5). GO terms in bold are those appearing in the top GO list of two different species.

a) <i>A. millepora</i>						
Rank	GO.ID	Term	#Ann.	#Obs.	#Exp.	<i>P</i> -value
3	GO:0003779	actin binding	278	11	4.25	<0.01
10	GO:0016670	oxidoreductase activity, acting on a sulfur group of donors, oxygen as acceptor	10	2	0.15	0.01
12	GO:0098772	molecular function regulator	724	19	11.07	0.01
15	GO:0008047	enzyme activator activity	244	9	3.73	0.01
16	GO:0030547	receptor inhibitor activity	11	2	0.17	0.01
25	GO:0016701	oxidoreductase activity, acting on single donors with incorporation of molecular oxygen	38	3	0.58	0.02
26	GO:0019209	kinase activator activity	39	3	0.6	0.02
32	GO:0030983	mismatched DNA binding	16	2	0.24	0.02
33	GO:0043028	cysteine-type endopeptidase regulator activity involved in apoptotic process	16	2	0.24	0.02
34	GO:0008092	cytoskeletal protein binding	644	16	9.85	0.03
39	GO:0015101	organic cation transmembrane transporter activity	19	2	0.29	0.03
41	GO:0051082	unfolded protein binding	83	4	1.27	0.04
43	GO:0060090	molecular adaptor activity	167	6	2.55	0.04
44	GO:0016620	oxidoreductase activity, acting on the aldehyde or oxo group of donors, NAD or NADP as acceptor	22	2	0.34	0.04
47	GO:0031072	heat shock protein binding	88	4	1.35	0.04
50	GO:0030545	receptor regulator activity	131	5	2	0.05

b) <i>P. damicornis</i>						
Rank	GO.ID	Term	#Ann.	#Obs.	#Exp.	<i>P</i> -value
2	GO:0005506	iron ion binding	126	6	1.78	0.01
7	GO:0016705	oxidoreductase activity, acting on paired donors, with incorporation or reduction of molecular oxygen	167	6	2.36	0.03
8	GO:0051015	actin filament binding	127	5	1.79	0.03
11	GO:0048487	beta-tubulin binding	24	2	0.34	0.04
15	GO:0008047	enzyme activator activity	251	7	3.55	0.06
17	GO:0004888	transmembrane signaling receptor activity	1045	20	14.76	0.06
19	GO:0008528	G protein-coupled peptide receptor activity	324	8	4.58	0.08
20	GO:0005102	signaling receptor binding	749	15	10.58	0.08
30	GO:0005096	GTPase activator activity	139	4	1.96	0.13
40	GO:0038023	signaling receptor activity	1142	20	16.13	0.14
42	GO:0016796	exonuclease activity, active with either ribo- or deoxyribonucleic acids and producing 5'-phosphomonoesters	47	2	0.66	0.14
46	GO:0016894	endonuclease activity, active with either ribo- or deoxyribonucleic acids and producing 3'-phosphomonoesters	11	1	0.16	0.15
47	GO:0030169	low-density lipoprotein particle binding	11	1	0.16	0.15
49	GO:0035877	death effector domain binding	11	1	0.16	0.15

c) <i>P. acuta</i>						
Rank	GO.ID	Term	#Ann.	#Obs.	#Exp.	<i>P</i> -value
5	GO:0022857	transmembrane transporter activity	702	18	10.17	0.01
9	GO:0015368	calcium:cation antiporter activity	11	2	0.16	0.01
10	GO:0005102	signaling receptor binding	749	18	10.85	0.01
21	GO:0015101	organic cation transmembrane transporter activity	16	2	0.23	0.02
35	GO:0000166	nucleotide binding	1297	25	18.79	0.03
41	GO:0031435	mitogen-activated protein kinase binding	23	2	0.33	0.04
42	GO:0030545	receptor regulator activity	137	5	1.98	0.05
45	GO:0051287	NAD binding	57	3	0.83	0.05
47	GO:0016651	oxidoreductase activity, acting on NAD(P)H	59	3	0.85	0.05
49	GO:0090722	receptor-receptor interaction	26	2	0.38	0.05

197 **Discussion**

198

199 *Different types of heat stress adaptation*

200

201 In each of the studied species, we detected genotype-environment associations that might
202 underpin local adaptation (Tab. 2a). Approximately half of these associations concerned
203 descriptors of sea surface temperature (SST; Tab. 2a), partially because these are the most
204 numerous types of descriptors employed in the analysis (20 out of 47; Tab. S2a).

205 As we focused on SST-related associations, we found that those involving bleaching alert
206 frequencies (BAFs) were more frequent (73) than those related to temperature variations (42)
207 and temperature averages (38; Tab. 2). Coral bleaching is a major threat for coral survival,
208 and bleaching conditions emerge when SST variation exceeds seasonal averages (Liu et al.
209 2003; Hughes et al. 2017). BAFs descriptors account precisely for this selective constraint (SST
210 variation over average), and this might explain why genotype-environment associations with
211 BAFs were more frequent. Previous work on coral seascape genomics also reported a
212 predominance of adaptive signals related to BAF (Selmoni et al. 2020b). Coral adaptation
213 appeared to be also driven by SST averages (regardless of variations) or by SST variations
214 (regardless of the averages; Tab. 2b). This kind of adaptation might relate to bleaching (e.g.
215 being adapted to high thermal variability promotes bleaching resistance; Safaie et al. 2018),
216 or to other types of heat stress responses (e.g. impaired injury recovery at elevated average
217 SST; Bonesso et al. 2017).

218

219 *Candidate molecular targets for heat stress adaptation*

220

221 Previous research reported that reefs exposed to high frequency of daily thermal variability
222 showed reduced bleaching prevalence (Safaie et al. 2018). One of the reasons might be that
223 corals at these sites manage to rapidly readjust their cellular homeostasis (Ruiz-Jones and
224 Palumbi 2017). This view is supported by the numerous GO terms describing activity
225 regulators (for instance “signaling receptor binding”, “receptor regulator activity”, “enzyme
226 activator activity”, “molecular function regulator”; Tab. 3, S5) found surrounding heat stress
227 associated SNPs. The fact that these terms are not heat stress specific, however, invites to a
228 cautious interpretation.

229 In contrast, we also detected several genes coding for well-established molecular actors of
230 corals thermal stress responses in the neighborhood of heat stress associated SNPs (Tab. 4,
231 S4). In some rare cases, we found that SNPs fell directly in the coding sequence of genes, but
232 more frequently the SNPs were located several kb distance from genes (Tab. 4). However, this
233 does not exclude causative effects, as (1) the SNP detected could be physically linked to an
234 adaptive SNP in the coding sequence; (2) the adaptive SNP could be located several kb from
235 the target gene, as it is often the case (Brodie et al. 2016). Hereunder, we highlight the
236 different molecular functions and the related proteins that were found as potential targets
237 for thermal adaptation in corals.

238

239 - *DNA repair*

240 Heat stress impacts the integrity of nucleic acids and elicits mechanisms that promote DNA
241 damage-repair and RNA stability (Henry et al. 1992; Sottile and Nadin 2018). A previous
242 seascape genomics study on *Acropora digitifera* found five SNPs associated with heat stress
243 to be proximal to genes coding for Helicase Q (Selmoni et al. 2020b). Here we found Helicase
244 Q5 in the genomic neighbourhood of a SNP associated with heat stress in *A. millepora*
245 (Tab. 4). Helicases Q are required for efficient DNA repair during the initiation of the
246 replication fork (Sharma et al. 2006). Another family of helicases participating to this process
247 are Helicases MCMs (Daniel et al. 2013) and one was found next to an heat stress associated
248 SNP in *A. millepora* (Fig.3; Tab. 4).

249 In addition, we found proteins involved in DNA damage-repair that are known to be
250 differentially expressed in corals under heat stress. For instance, claspin (Palumbi et al. 2014;
251 Smits et al. 2019) and RAD51/54 homologs (Maor-Landaw and Levy 2016) were found here
252 surrounding heat stress associated SNPs in *A. millepora*, while DNA damage-binding protein 1
253 (Li et al. 2006) and DNA polymerase delta catalytic subunit (Prindle and Loeb 2012) in
254 *P. damicornis* (Tab. 4). Of note, GO terms describing molecular functions related to DNA
255 repair (“mismatched DNA binding”, “exonuclease activity”, “endonuclease activity”,
256 “nucleotide binding”) were found as over-represented in genes surrounding heat stress
257 associated SNPs in the three species (Tab. 3).

258

259 - *Protein folding*

260 One of the main groups of gene annotations surrounding heat stress associated SNPs
261 concerned molecular chaperones (Tab. 4). These are proteins that intervene in cellular
262 responses to heat stress, where they assist the folding or unfolding of proteins in the
263 endoplasmic reticulum notably (ER; Oakley et al. 2017). In corals, the role of these proteins in
264 heat response, as well as their up-regulation under thermal stress, have been reported in
265 several studies (Desalvo et al. 2008; Ishikawa et al. 2009; van Oppen and Lough 2009; Desalvo
266 et al. 2010; Rosic et al. 2011; Maor-Landaw and Levy 2016; Oakley et al. 2017; Ruiz-Jones and
267 Palumbi 2017). The annotation analysis for the related GO terms (e.g. “heat shock protein
268 binding”, “unfolded protein binding”), revealed that this function was over-represented in
269 genes close to the heat stress associated SNPs of *A. millepora* (Tab. 3a). In the three species,
270 the genomic neighborhood of heat stress associated SNP contained several classes of
271 chaperones: DnaJ homologs (four in *A. millepora*, one in *P. damicornis*), Tubulin-specific
272 chaperone A and NudC domain-containing protein (*A. millepora*; Zheng et al. 2011), prolyl 3-
273 hydroxylase 1 (*P. damicornis*; Ishikawa et al. 2009) and selenoprotein-F (*P. acuta*; Ren et al.
274 2018). Another important class of chaperones are “Protein disulfide-isomerase”, which
275 catalyze the formation or breakage of disulfide bonds in proteins and produce reactive oxygen
276 species (ROS) as byproducts (van Oppen and Lough 2009; Oakley et al. 2017). For example,
277 the disulfide-isomerase gene expression has been shown to be upregulated in *Stylophora*
278 *pistillata* under experimental heat stress (Maor-Landaw et al. 2014). Disulfide-isomerase 2
279 genes were found next to heat stress associated SNPs in each of the studied species (Tab. 4).
280

281 - *Oxidative stress response*

282 In parallel to the protein folding and recycling response, coral cells under heat stress
283 accumulate ROS (Oakley et al. 2017; Nielsen et al. 2018). This accumulation can derive from
284 the leakage of ROS from the damaged photosynthetic machinery of the endosymbiont, as
285 well as from the endogenous production of the host mitochondria elicited under heat stress
286 (Oakley et al. 2017; Nielsen et al. 2018). ROS accumulation causes oxidative stress, and the
287 GO terms describing the inherent responses (“oxidoreductase activity”) were found as over-
288 represented in genes next to heat stress associated SNPs in the three studied species (Tab. 3).
289 For instance, we found genes coding for Peroxidasin homolog and Isocitrate dehydrogenase
290 subunit beta next to heat stress associated SNPs in *A. millepora* (Tab. 4). Peroxidasin homolog
291 is in the first line of defence against ROS accumulation, displays high rates of evolution in

292 *A. millepora* and was found highly up-regulated under heat stress in *Monastrea faveolata*
293 (= *Orbicella faveolata*; Voolstra et al. 2009; Voolstra et al. 2011; Louis et al. 2017). Isocitrate
294 dehydrogenase is one of the few sources of NADPH in the animal cell and was found
295 upregulated in *E. pallida* under heat stress (Kültz 2005; Oakley et al. 2017). NADPH is an
296 essential substrate to contrast ROS accumulation (Oakley et al. 2017; Patel et al. 2018) and
297 another gene implicated in its metabolism, Quinone oxidoreductase PIG3, was found close to
298 heat stress associated SNPs in *P. acuta* (Tab. 4; Zangar et al. 2004). In *P. damicornis*, we found
299 the Glutathione peroxidase 5 gene, belonging to a family of well-characterized antioxidants
300 contrasting ROS accumulation in corals (Nielsen et al. 2018).

301 In the host mitochondria, ROS production occurs in a series of redox reactions across the inner
302 membrane (the electron transport chain; Lutz et al. 2015). One of the main components of
303 this chain is the respiratory complex I (NADH-ubiquinone oxidoreductase), and we found
304 genes coding for two of its subunits in the genomic neighborhood of heat stress associated
305 SNPs in both *Pocillopora* species (Tab. 4). The mechanism leading to ROS leakage from host
306 mitochondria into coral cell cytoplasm is poorly known (Dunn et al. 2012; Oakley et al. 2017;
307 Nielsen et al. 2018). However, it is noteworthy to mention that in *A. millepora* the SNP most
308 strongly associated with heat stress was close to the MIC60 gene (Tab. 4, S4a). MIC60 is a
309 subunit of the MICOS complex, a key protein in the maintenance of the mitochondrial inner
310 membrane architecture, through which ROS are produced, and the outer membrane, through
311 which ROS diffuse into cytoplasm (Muñoz-Gómez et al. 2015; Zhao et al. 2019).

312

313 - Inflammatory response and apoptosis

314 The effects of ROS depends on the level of accumulation: medium levels elicit an
315 inflammatory response, while excessive levels lead to cell apoptosis (Patel et al. 2018).
316 Mitogen-activated protein kinases (MAPK) are key actors in the inflammatory response (Son
317 et al. 2013; Courtial et al. 2017; Patel et al. 2018). In corals, MAPKs were shown to repress
318 ROS accumulation in *S. pistillata* (Courtial et al. 2017) and were found in proximity of a SNP
319 associated to heat stress in another seascape genomics study on Japanese *A. digitifera*
320 (Selmoni et al. 2020b). Here we found a MAPK coding gene around heat stress associated
321 SNPs in *A. millepora* (MAPK1), and genes coding for a MAPK activating protein *P. damicornis*
322 (Putative MAPK-activating protein FM08) and *P. acuta* (TNF receptor-associated factor 6;
323 Tab. 4; Mason et al. 2004).

324 ROS excess eventually results in cell apoptosis (Patel et al. 2018). In each of the three species,
325 we found apoptosis-related genes near heat stress associated SNPs: Programmed cell death
326 protein 6 and Death-associated protein kinase 2 (*A. millepora*); Death effector domain-
327 containing protein (*P. damicornis*); and apoptosis regulator Bcl-2 and Death-associated
328 protein 4 (*P. acuta*; Tab 4). These proteins participate in the caspase-mediated apoptotic
329 cascade involved in the coral bleaching process (Ahmad et al. 1997; Dunn et al. 2007; Valmiki
330 and Ramos 2009; Tchernov et al. 2011; Yuasa et al. 2015; Oakley et al. 2017).

331

332 - *Cell structure*

333 Heat stress has been shown to lead to cytoskeleton reorganization (Wilson et al. 2016). In
334 Cnidaria, cytoskeletal proteins displayed changes in abundance under experimental heat
335 stress in *Exaiptasia pallida* and *A. palmata* (Ricaurte et al. 2016; Oakley et al. 2017). Here we
336 found several genes implicated in the cytoskeletal architecture in proximity of SNPs
337 associated to heat stress: myosin III (twice), unconventional myosin VIIb and actin
338 (*A. millepora*), unconventional myosin-Id (*P. damicornis*), Myosin heavy chain and actin-1
339 (*P. acuta*; Tab. 4). Moreover, the GO terms “actin binding” and “cytoskeletal protein binding”
340 were over-represented in the set of genes neighbouring heat stress associated SNPs in
341 *A. millepora*, and “actin filament binding” in *P. damicornis* (Tab. 3a-b). Of note, in three of the
342 myosin genes the heat stress associate SNPs were found inside the coding sequence (Tab. 4,
343 S4).

344

Table 4. Candidate molecular targets for coral adaptation to heat stress. Annotations of genes surrounding (± 50 kb) heat stress associated SNPs were sorted by molecular function. For six specific types of molecular functions, the table displays the genes potentially involved in heat stress response (Putative molecular target) and the position of the corresponding SNP associated with heat stress (SNP position, in format contig/chromosome: base position). The CDS tag indicates SNP falling inside the coding sequence of the putative molecular target. The background colours correspond to the species where the candidate molecular target was found (pink: *Acropora millepora*, blue: *Pocillopora damicornis*, green: *Pocillopora acuta*). The role that each molecular function has in coral heat response is described in the last column.

Molecular function	SNP position	Putative molecular target	Role in coral heat response
DNA damage repair	xpSc0000535:118526	ATP-dependent DNA helicase Q5	Heat stress and UV radiation can provoke DNA damage. Proteins in this list participate to the DNA damage-repair mechanism during replication.
	chr7:20210292	DNA helicase MCM9	
	chr13:7351849	DNA repair protein RAD51 homolog 2	
	chr13:20841309	Claspin	
	NW_020844825.1:56001	DNA polymerase delta catalytic subunit	
	NW_020847490.1:185714	DNA damage-binding protein 1	
Protein folding / chaperone / heat stress responses	chr7:2529319	DnaJ homolog subfamily B member 6	Molecular chaperones are activated in early response to thermal stress to assist protein folding in the endoplasmic reticulum.
	Sc0000122:685084	Sacsin (DnaJ homolog subfamily C member 29)	
	chr3:18355053	DnaJ homolog subfamily A member 3	
	chr5:20803650	NudC domain-containing protein 2	
	chr5:10318559	Tubulin-specific chaperone A	
	chr13: 18019180	Protein disulfide-isomerase 2	
	chr12:16446416	prolyl 3-hydroxylase 1	
	NW_020844635.1:120545	DnaJ homolog subfamily C member 9	
	NW_020844967.1:239041	Protein disulfide-isomerase 2	
	NW_020845264.1:254233	Protein disulfide-isomerase 2	
	NW_020846699.1:47695	Protein disulfide-isomerase 2	
	NW_020847700.1: 75539	selenoprotein-F	
Oxidative stress	chr2:15023513	MIC60	Heat stress leads to the accumulation of ROS. All the proteins in this list participates to the metabolism of ROS. MIC60 is a key structural protein of the inner membrane of mitochondria were endogenous ROS is produced.
	xfSc0000142:60614	Peroxidasin homolog	
	chr7:20210292	Isocitrate dehydrogenase [NAD] subunit beta	
	NW_020843829.1:322689	NADH-ubiquinone oxidoreductase 23 kDa subunit	
	NW_020844825.1:56001	Glutathione peroxidase 5	
	NW_020846699.1:47695	NADH-ubiquinone oxidoreductase B16.6 subunit	
	NW_020845243.1:143874	Quinone oxidoreductase PIG3	
Inflammatory and apoptotic response	chr13:19271503	Mitogen-activated protein kinase-binding protein 1 (MAPK1)	MAPKs are activated in the inflammatory response caused by ROS accumulation. Excessive ROS accumulation elicits the caspases-mediated apoptotic response observed in coral bleaching.
	chr11:11687264	Death-associated protein kinase 2	
	xfs0000077:45906	Programmed cell death protein 6	
	NW_020844212.1: 38069 (CDS)	Partial similarity with DED domain-containing protein	
	NW_020845264.1:254233	Putative MAPK-activating protein FM08	
	NW_020846901.1:188731	apoptosis regulator Bcl-2	
	NW_020846154.1:13424	Death-associated protein 4	
Cytoskeleton	NW_020846522.1:185468	TNF receptor-associated factor 6	ROS accumulation degrades structural proteins and lead to a pronounced reorganization of the cytoskeleton.
	chr1:19795515	Myosin IIIA	
	chr12:22835441 (CDS)	Unconventional myosin-VIIb	
	chr2:15023513 (CDS)	Partial similarity with Myosin III	
	chr3:17298197	Actin, cytoplasmic	
	NW_020844967.1:239041	Unconventional myosin-Id	
	NW_020847027.1:161971	Actin-1	
	NW_020846942.1:1165830 (CDS)	Myosin heavy chain	

345

346 *Limitations and future directions*

347

348 Seascape genomics studies are exploratory analyses that come with the drawback of being
 349 subjected to high false discovery rates (Rellstab et al. 2015; Riginos et al. 2016). This bias is
 350 stressed when the confounding role of neutral genetic variation is not accounted for (Selmoni
 351 et al. 2020a). The preliminary analysis of population structure, however, did not reveal any
 352 cryptic speciation nor isolated reefs among the studied populations (Fig. 2, S1-S2).

353 Furthermore, we used a statistical method (LFMM) and a sampling design allowing to mitigate
354 such confounding effects (Frichot et al. 2013; Selmoni et al. 2020a).
355 There are, nevertheless, some points that could have increased the statistical power of the
356 analysis, as for instance the use of a larger sample size (over 200 individuals per species;
357 (Selmoni et al. 2020a). In addition, the assignment of heat stress associated SNPs to candidate
358 molecular targets for adaptation would have been further facilitated with a higher genome
359 resolution in the sequencing strategy. Higher genome resolution would also allow to infer the
360 structural modification that SNPs falling inside the coding sequence might cause. In the years
361 to come, whole-genome-sequencing on corals is likely to become more affordable and can
362 then be applied to the large sample sizes required for seascape genomics studies.
363 The next step in the characterization of corals' adaptive potential is experimental validation.
364 Our work found several genetic variants that might confer selective advantages against
365 thermal stress (Tab. 4). For each of the studied species, we can now define multiple-loci
366 genotypes of heat stress resistant colonies and test their fitness under experimental heat
367 stress conducted in aquaria (Krueger et al. 2017). As a result, this analysis will allow to 1)
368 further investigate the role of different heat stress associated genotypes and molecular
369 pathways and 2) provide a concrete measure of the thermal ranges that these coral
370 populations might sustain in the years to come. This information is of paramount importance,
371 as it will allow to predict the reefs that are expected to already carry heat tolerant colonies
372 and to define conservation strategies accordingly (Selmoni et al. 2020b). For instance, marine
373 protected areas could be established to preserve reefs with higher adaptive potential against
374 heat stress, where such reefs could provide the breeding stock to restore the damaged ones
375 (Baums 2008; Van Oppen et al. 2015; van Oppen et al. 2017).
376

377 *Conclusions*

378
379 In this study, seascape genomics allowed to uncover genetic variants potentially implicated
380 in adaptive processes against different types of heat stress in three coral species of New
381 Caledonia. These variants were located next to genes coding for molecular actors that
382 participate in well-understood cellular reactions against thermal stress. In addition, the
383 approach pointed out new candidate genes (e.g. Helicase Q) or processes (e.g. signalling
384 receptor binding) that might be implied in such responses. Of note, some of these potential

385 targets for adaptation recurred in the analyses of different species, supporting the robustness
386 and the power of the seascape genomics. Future studies will focus on performing
387 experimental assays to validate the implication of potentially adaptive genotypes and newly
388 identified genes in the heat stress response and to measure the thermal ranges tolerated by
389 the diverse adaptive genotypes.

390

391

392 **Material and methods**

393

394 *Environmental data*

395

396 The seascape genomics approach requires an exhaustive description of the environmental
397 conditions in order to prevent the misleading effect of collinear gradients (Riginos et al. 2016;
398 Leempoel et al. 2017). For this reason, the seascape characterization we used encompassed
399 seven environmental variables: sea water temperature (SST), chlorophyll concentration, sea
400 surface salinity, sea current velocity, suspended particulate matter, alkalinity and bleaching
401 alert frequencies (BAF; Tab. S2). The environmental characterization was performed in the R
402 environment (R Core Team 2016) using the *raster* package (Hijmans 2016) and following the
403 method described in previous work on coral seascape genomics (Selmoni et al. 2020b) with
404 some modifications outlined hereafter.

405 For the description of SST we used two different georeferenced datasets covering the extent
406 of New Caledonia: (1) daily records of SST since 1981 at a spatial resolution of 5 km (SST_{5km};
407 EU Copernicus Marine Service 2017); (2) daily records of SST since 2002 at resolution of 1 km
408 (SST_{1km}; Group for High Resolution Sea Surface Temperature; Chao et al. 2009; Chin et al.
409 2017). The first dataset covers a wider temporal range, therefore providing a more reliable
410 characterization of historical trends. The second dataset covers a smaller temporal window,
411 but the higher geographic resolution allows to portray fine scale thermal patterns with a
412 higher degree of confidence. Both datasets were used to compute, for each pixel of the study
413 area, averages and standard deviations of the warmest month, the coldest month, and the
414 entire observational period. Furthermore, both datasets were used to compute three indices
415 of bleaching alert frequencies (BAF), representing the frequency of days (over the whole
416 period of remote sensing) during which the bi-weekly accumulated heat stress (*i.e.* SST above

417 the average maximum) exceeded 0°C ($BAF_{0^\circ C}$), 4°C ($BAF_{4^\circ C}$) and 8°C ($BAF_{8^\circ C}$) (Selmoni et al.
418 2020b). Similarly, we computed the frequency of the bleaching warning conditions as defined
419 by the Coral Reef Watch (BAF_{CRW}), corresponding to the accumulation of heat over the three
420 previous months (Liu et al. 2003).

421 For the other datasets (chlorophyll concentration, sea surface salinity, sea current velocity
422 and suspended particulate matter; EU Copernicus Marine Service 2017), the spatial resolution
423 ranged between 4 and 9 km (Tab. S2). All the datasets covered a temporal extent of at least
424 20 years before 2018 (the year of sampling) and were processed to compute: (1) highest
425 monthly average, (2) lowest monthly average and (3) overall average. For all the datasets
426 captured at daily resolution (*i.e.* all except suspended particulate matter), we also computed
427 the standard deviation associated with the three means. Seawater alkalinity was estimated
428 by combining SST_{5km} and salinity in a polynomial equation as described by Lee and colleagues
429 (Lee et al. 2006).

430 In total, 47 environmental descriptors (Tab. S2) were computed and assigned to the shapes
431 of the reefs of New Caledonia (UNEP-WCMC et al. 2010), reported for a regular grid (~3,000
432 cells of size: 2x2 km) using QGIS (QGIS development team 2009).

433

434 *Sampling*

435

436 Twenty sampling sites were selected out of the ~3,000 reef cells surrounding Grande Terre,
437 the main island of New Caledonia (Fig. 1). Sampling sites were chosen following an approach
438 that simultaneously maximized environmental contrasts and replicated them at distant sites
439 (the “hybrid approach” described in Selmoni et al. 2020a). The method consists of applying
440 Principal Component Analysis (PCA) and hierarchical clustering to the 47 environmental
441 descriptors in order to separate the ~3,000 reef cells into distinct environmental regions.
442 Next, the algorithm selects the same number of sampling sites within each region in order to
443 maximize physical distance between sites. Increasing environmental variation is expected to
444 raise the sensitivity of seascape genomics analysis, while the replication of environmental
445 gradients is expected to reduce false discovery rates (Selmoni et al. 2020a). Here the number
446 of environmental clusters was five (Fig. S3) and we established at four sampling locations per
447 cluster. When this was not possible (*e.g.* because of logistic constraints during the sampling
448 campaign), additional sampling sites were added to the neighbouring clusters.

449 The sampling campaign was performed from February to May 2018 (under the permits
450 N°609011-/2018/DEPART/JJC and N°783-2018/ARR/DENV) and targeted three flagship
451 species of the Indo-Pacific: *Acropora millepora*, *Pocillopora damicornis* *sensu* Schmidt-Roach
452 et al. (2013) [corresponding to PSH04 *sensu* P. Gélin et al. (2017)] and *Pocillopora acuta* *sensu*
453 Schmidt-Roach et al. (2013) [corresponding to PSH05 *sensu* P. Gélin et al. (2017)]. Of note,
454 *P. acuta* and *P. damicornis* belong to the complex of species formerly named *P. damicornis*
455 (Schmidt-Roach et al. 2014; Johnston et al. 2017; Gélin, Fauvelot, et al. 2018). At every
456 sampling site, we collected up to 20 samples of *A. millepora* and 20 of *Pocillopora* aff.
457 *damicornis* (we did not discriminate between species while sampling as it can be difficult to
458 distinguish them in the field). All the samples were collected in a 1 km area and at a depth
459 ranging between 2-4 m. The centre of this area was used for georeferencing the sampling site.
460 Before sampling, each colony was imaged underwater, then a portion of a branch was
461 sampled with hammer and chisel. Each sample consisted of a 1-2 cm branch that was
462 immediately transferred to 80% ethanol and stored at -20°C. DNA from the 730 samples (370
463 *A. millepora* and 360 *Pocillopora*; Tab. S1) were extracted using the DNeasy 96 Tissue kit
464 (Qiagen) following manufacturer instructions.

465

466 *Pocillopora species identification*

467

468 The 360 *Pocillopora* samples were identified molecularly *a posteriori* of sampling to be
469 assigned to one species or the other (*P. damicornis* or *P. acuta*). Samples were thus genotyped
470 using 13 microsatellite loci, as in Gélin et al. (2017; Online Resource 1). Then, colonies
471 belonging to *P. damicornis* and to *P. acuta* were identified using assignment tests performed
472 with STRUCTURE (v. 2.3.4 ;Pritchard et al. 2000), as in Gélin et al. (2018). Colonies assigned to
473 *P. damicornis* or *acuta* with a probability of at least 0.70 were retained in the final dataset for
474 this study. The *Pocillopora* sampling was composed of 148 *P. damicornis* (more precisely to
475 SSH04b *sensu* Gélin et al. 2017), 159 *P. acuta* colonies (more precisely, a mix of SSH05a and
476 SSH05b *sensu* Gélin, Pirog, et al. 2018) and 53 unassigned colonies (excluded from further
477 analysis; Tab. S1).

478

479 *Acropora species identification*

480 *Acropora* species are genetically and morphologically notoriously challenging in terms of
481 identification and species boundaries detection. However, *A. millepora* can be recognised in
482 the field based on its typical axial and radial corallite shape (Wallace 1999). Back from the
483 field, *in situ* images of each specimen were examined to look for the species diagnostic
484 morphological characters and initial identifications validated.

485

486 *Screening and SNP genotyping*

487

488 All DNA samples from *A. millepora*, *P. damicornis* and *P. acuta* were sent to the Diversity
489 Arrays Technology (Canberra, Australia) for quality check screening and genotype-by-
490 sequencing using the DArT-sequencing method (DArT-seq; Kilian et al. 2012). The restriction
491 enzymes used for library preparation for *A. millepora* and *Pocillopora* samples were PstI and
492 HpaII. Prior to sequencing, all the DNA samples underwent a one-hour incubation with the
493 digestion buffer, followed by a step of quality check for integrity, purity and concentration
494 running 1 µL per sample on a 0.8% agarose gel. Samples from each site were then ranked
495 based on their quality (degree of smearing on the agarose gel). We then selected the samples
496 with the best scores from each site and defined a list of 188 *A. millepora*, 128 *P. damicornis*
497 and 150 *P. acuta* samples that proceeded to the sequencing step in four and five lanes of an
498 Illumina Hiseq2500, respectively. During each step of the workflow (DNA purification, library
499 preparation and sequencing), *A. millepora* and *Pocillopora* samples were kept separated and
500 randomly distributed across the respective batches (e.g. 96-well plates, sequencing lanes) to
501 minimize the risks of technical bias. SNPs were called using the DArT-seq analytical pipeline
502 (DArTsoft14).

503

504 *SNP filtering*

505

506 The DArT-loci (*i.e.* the DNA sequences surrounding each SNP) initially underwent a sequence
507 similarity search (BLASTn; v. 2.7.1; Madden and Coulouris 2008) against a reference genome
508 to retain only those associated with the coral host. For *A. millepora*, the reference genome
509 was the *A. millepora* chromosome-level assembly from Fuller and colleagues (v. 2; Fuller et
510 al. 2019, unpublished data, available on arXiv), while for *P. damicornis* and *P. acuta* we used

511 the only *Pocillopora* reference assembled to date (*P. damicornis sensu lato*; v. 1; Cunning et
512 al. 2018). Only DArT-loci scoring an E-value below 10^{-6} were retained.
513 The processing of the SNPs data followed a pipeline from previous work on coral seascapes
514 genomics (Selmoni et al. 2020b). For each species' dataset, we removed SNPs and individuals
515 with high missing rates ($> 50\%$) by using custom functions in the R environment. Next, we
516 proceeded with imputation of missing genotypes using the *linkimpute* algorithm (based on k-
517 nearest-neighbours imputation; Money et al. 2015) implemented in Tassel 5 (Bradbury et al.
518 2007) using the default settings. Afterwards, we repeated the filtering of SNPs and individuals
519 for missing rates, but this time using a more stringent threshold (5%). We also applied a filter
520 to exclude rare alleles (minor allele frequency $< 5\%$) and highly frequent genotypes (major
521 genotype frequency $> 95\%$). SNPs were then filtered for linkage disequilibrium using the R
522 package *SNPrelate* (function *snpsgdsLDpruning*, LDthreshold=0.3, v.1.16; Zheng et al. 2012).
523 Finally, we applied a filter for clonality: when groups of colonies shared highly correlated
524 genotypes (Pearson correlation > 0.9) only one colony per group was kept.
525

526 *Neutral genetic structure analysis*

527
528 Prior to the seascapes genomics analysis, the neutral genetic structure of the studied
529 populations was investigated by running a PCA on the genotype matrix of each species using
530 the R *stats* package (*prcomp* function). Firstly, we visually inspected the percentage of
531 variance of the genotype matrix explained by each principal component (eigenanalysis); in
532 highly structured populations the first principal components (PCs) are expected to explain a
533 larger proportion of the variance, when compared to the subsequent PCs (Johnstone 2001;
534 Novembre et al. 2008). We ran a Tracy-Widom test as implemented in R package *AssocTests*
535 (Wang et al. 2017) to determine the number of PCs underlying a non-random genetic
536 structure ($P < 0.05$; Patterson et al. 2006).

537 We also visualized the spatial distribution of the main axis of variation (PC1), in order to
538 investigate the presence of geographical structures (Novembre et al. 2008). Finally, we
539 focused on the SNP-specific loadings on PC1 and their distributions across the genome. In
540 fact, groups of genetically isolated individuals (e.g. hybrids, cryptic species) are expected to
541 display genomic islands of low-recombination (i.e. groups of physically close SNPs
542 contributing to high loading on the main axis of variation; Nosil et al. 2009; Li and Ralph 2019).

543 We therefore visualized the distribution of average PC1-loadings by genomic windows of 50
544 and 100 kb. In these calculations, only genomic windows containing at least 5 SNPs were
545 retained.

546

547 *Seascape genomics*

548

549 The seascape genomics analyses were performed separately on the three species using the
550 LFMM method implemented in the *LEA* R package (v. 2.4.0; Fritchot et al. 2013; Fritchot and
551 François 2015). This method associates single environmental gradients and individual SNPs
552 variations in mixed models, where the confounding effect of neutral genetic variation is
553 accounted for through latent factors (Fritchot et al. 2013).

554 Briefly, the first step of the LFMM pipeline is to estimate the number of latent factors (K;
555 Fritchot and François 2015). This parameter corresponds to the number of ancestral
556 populations and can be estimated by using the *snmf* function of the *LEA* package. The method
557 processes a genotype matrix to estimate individual admixture coefficients under different K's,
558 and then evaluates the quality of fit for each K via cross validation (Fritchot and François 2015).
559 We ran ten replicates of this analysis for all the studied species, and found that the optimal
560 number of K (according to the lowest entropy criterion) ranged from 2-4 for *A. millepora*, 6-8
561 for *P. damicornis*, and 10-12 for *P. acuta* (Fig. S4).

562 We then proceeded to the genotype-environment association analysis with LFMM. Since this
563 method does not accept missing genotypes, we first ran the *impute* function of the *LEA*
564 package. For each species, this function inferred the missing genotypes out of the ancestral
565 genotype frequencies previously calculated with the *snmf* function. Finally, we ran the
566 association analysis between the SNPs of each species and the environmental condition
567 descriptors. This was done by using the *lfmm* function, setting K to the ranges previously
568 estimated for each species and running five replicates of each analysis. Since *lfmm*
569 calculations can be computationally intensive, when two or more environmental descriptors
570 were highly collinear (absolute value of Pearson correlation > 0.9), only one was used in the
571 analysis.

572 LFMM returns *P*-values describing the statistical significance of every genotype-environment
573 association under different values of K. For each association model related to the same
574 environmental variable, *P*-values were corrected for multiple testing using the *q-value*

575 method (R package *q-value*, v. 2.14, Storey 2003) and deemed significant if $q < 0.01$ under at
576 least one level of K. When a SNP was significantly associated to multiple environmental
577 variables, or with an environmental variable highly correlated with others (*i.e.* those excluded
578 from the LFMM calculations), we defined a main environmental descriptor as the variable
579 most strongly correlated (Pearson) with the SNP.

580

581 *Annotation analysis of heat stress associated SNPs*

582

583 For each of the three studied species, we annotated the genomic neighbourhood of every
584 SNP in order to characterize the possible functional implications of a genetic variant. Firstly,
585 we uniformed the annotations of genes from different species in order to facilitate
586 comparisons. We did this by retrieving the positions of genes in the two reference genomes
587 (Cunning et al. 2018; Fuller et al. 2019) and the corresponding predicted protein sequences.
588 These sequences were used to perform a similarity search (blastP; Madden and Coulouris
589 2008) against the Uniprot/swissprot database (*metazoa* entries, release 2020_01;
590 Boeckmann et al. 2003). Each gene was annotated with the best significant hit (E-value < 0.01)
591 and inherited protein name and gene ontology (GO) terms describing molecular functions
592 when existing (Ashburner et al. 2000).

593 Afterwards, we focused on the annotation of genes surrounding SNPs associated with heat
594 stress descriptors. Significant SNPs were deemed “heat stress associated” if they were best
595 correlated with an environmental descriptor relating to average temperature, standard
596 deviation of temperature, or bleaching alert frequency. We mapped every SNP associated
597 with heat stress as the genes located within a ± 50 kb window. We selected this window size
598 because genes associated with a SNP can be located hundreds of kilobases away (Visel et al.
599 2009; Brodie et al. 2016), with 50 kb being roughly the median contig size in the *P. damicornis*
600 reference genome (Cunning et al. 2018).

601 We then computed the observed occurrence of each GO term among the genomic
602 neighbourhoods of significant SNPs. As a comparison, we split the reference genome into
603 100 kb windows and computed the expected occurrence for each term. This procedure was
604 performed separately for each of the three studied species, using the respective reference
605 genome. The statistical analysis of differences between the expected and observed GO term
606 occurrences (enrichment analysis) was performed using the Fisher exact test method as

607 implemented in the R *topGO* package (v. 2.34; Alexa et al. 2006). As suggested in the *topGO*
608 guidelines, we ranked GO terms according to the *P-value* of the Fisher test and discarded
609 those occurring fewer than 10 times throughout the genome.

610

611

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613

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621

622

623 **References**

624

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